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## 2 **Projecting Global Mercury Emissions and Deposition Under the Shared** 3 **Socioeconomic Pathways**

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### 11 **Key Points:**

- 12 • Future (2010-2300) anthropogenic Hg releases to air, land and water (0.7-1.7 Tg) are  
13 similar to historical (1510-2010) releases (1.1-2.8 Tg)
- 14 • Cumulative anthropogenic Hg emissions to air (2010-2300) vary by a factor of two  
15 across scenarios (110-230 Gg)
- 16 • Deposition declines from near-term reductions in anthropogenic Hg emissions are  
17 amplified by reductions in re-emissions from land and ocean

### 18 **Abstract**

19 Mercury (Hg) is a naturally occurring element that has been greatly enriched in the environment  
20 by activities like mining and fossil fuel combustion. Despite commonalities in some CO<sub>2</sub> and Hg  
21 emission sources, the implications of long-range climate scenarios for anthropogenic Hg  
22 emissions have yet to be explored. Here, we present comprehensive projections of anthropogenic  
23 Hg emissions (2020-2300) and evaluate impacts on global atmospheric Hg deposition.  
24 Projections are based on four shared socioeconomic pathway (SSP) narratives ranging from  
25 sustainable reductions in resource and energy intensity to rapid economic growth driven by  
26 abundant fossil fuel exploitation. There is a greater than two-fold difference in cumulative  
27 anthropogenic Hg emissions between the lower-bound (110 Gg) and upper-bound (230 Gg)  
28 scenarios. Hg releases to land and water are approximately six times those of direct emissions to  
29 air (600-1470 Gg). At their peak, anthropogenic Hg emissions reach 2200-2600 Mg a<sup>-1</sup> sometime  
30 between 2010 (baseline) and 2030, depending on the SSP scenario. Coal combustion is the  
31 largest determinant of differences in Hg emissions among scenarios. Decoupling of Hg and CO<sub>2</sub>  
32 emissions sources occurs under low- to mid-range scenarios, though contributions from artisanal  
33 and small-scale gold mining remain uncertain. A projected future shift in speciation of Hg  
34 emissions toward lower gaseous elemental Hg (Hg<sup>0</sup>) and higher divalent Hg (Hg<sup>II</sup>) will result in  
35 a higher fraction of locally-sourced Hg deposition. Projected re-emissions of previously  
36 deposited anthropogenic Hg follow a similar temporal trajectory to primary emissions,  
37 amplifying benefits of primary Hg emissions reductions under the most stringent mitigation  
38 scenarios.

## 39 Plain Language Summary

40 Mercury is a global pollutant that is emitted alongside greenhouse gases like carbon dioxide  
41 (CO<sub>2</sub>) when fossil fuels such as coal are burned. Researchers have projected how emissions of  
42 greenhouse gases and climate are likely to change in the future, but relatively little is known  
43 about future Hg releases. Here, we project future Hg emissions between 2020 and 2300 using  
44 growth scenarios developed by climate change researchers. Under low emission scenarios, Hg  
45 emissions are projected to peak between 2010 and 2030. Under the high emission scenario, Hg  
46 releases continue near present-day levels until after 2060, and decline more slowly than other  
47 scenarios thereafter. Large variability in projected releases (cumulatively a two-fold difference)  
48 is apparent across the low and high scenarios. Globally, the intensity of coal combustion and  
49 how quickly it is phased out is the largest driver of future Hg releases. We then use global  
50 models to simulate future atmospheric Hg deposition, identifying multiple factors responsible for  
51 changing deposition patterns and amounts. We find there is a penalty for delaying reductions in  
52 Hg emissions because of increased reemissions from the land and ocean in the future. This work  
53 emphasizes the benefits of near-term stringent global reductions in anthropogenic Hg releases.

## 54 1 Introduction

55 Mercury (Hg) is a naturally occurring element, but human activities such as mining and fossil-  
56 fuel combustion have released approximately 1.5 Tg of Hg from stable geologic reservoirs to the  
57 atmosphere, land, and water over the past 500 years (1510–2010) (Streets et al., 2019a). This has  
58 significantly altered the natural biogeochemical Hg cycle (Geyman et al., 2023; Streets et al.,  
59 2019a) and adversely affected the health of exposed humans and wildlife (Basu et al., 2023).  
60 Reemission of anthropogenic Hg from terrestrial and aquatic ecosystems extends its lifetime in  
61 the biosphere (referred to as “legacy Hg”). Past work has quantified the impacts of historical  
62 emissions and legacy Hg cycling on the global Hg cycle (Amos et al., 2013; Angot et al., 2018;  
63 Guerrero & Schneider, 2023; Nriagu, 1994). However, estimates of future anthropogenic Hg  
64 emissions based on the most recent future growth scenarios adopted by the Intergovernmental  
65 Panel on Climate Change (IPCC) and their drivers are not presently available (O’Neill et al.,  
66 2016), limiting our ability to project future scenarios of Hg pollution.

67 Shared Socioeconomic Pathways (SSPs) and associated emission scenarios offer unified  
68 narratives ranging from sustainable development to fossil fuel-driven growth (O’Neill et al.,  
69 2016, 2017). They provide a framework for bounding development and greenhouse gas  
70 trajectories extending to the year 2300 (Meinshausen et al., 2020). However, these narratives do  
71 not explicitly address Hg emissions or the technological transformations underpinning future  
72 changes in Hg emission intensity. We fill this gap by developing a methodology for projecting  
73 time-dependent change in activity-specific Hg emission factors. This approach is built upon  
74 detailed parameterizations developed for historical emission inventories (Streets et al., 2011,  
75 2019a), which facilitates intercomparison of past and future emissions.

76 Anthropogenic Hg is released to the atmosphere in both the elemental (Hg<sup>0</sup>) and oxidized (Hg<sup>II</sup>)  
77 forms. Hg<sup>II</sup> has an atmospheric lifetime against deposition of a few days (Corbitt et al., 2011),  
78 while the lifetime of Hg<sup>0</sup> is greater than one year (Horowitz et al., 2017; Shah et al., 2021).  
79 Previous studies projected Hg deposition in 2035 and 2050 based on anthropogenic Hg  
80 emissions estimated from older IPCC Special Report on Emission Scenarios (SRES) and

81 independent estimates (Corbitt et al., 2011; Pacyna et al., 2016; Streets et al., 2009). Results  
82 suggested future increases in  $\text{Hg}^{\text{II}}$  emissions relative to  $\text{Hg}^0$  are likely to increase the proportion  
83 of regional Hg deposition from emitting countries. The SRES emission trajectories ranged from a  
84 best-case scenario that showed relatively flat anthropogenic Hg emissions between 2006 and  
85 2050, to a scenario characterized by higher energy and economic growth (A1B) that  
86 approximately doubled 2006 emissions levels (Streets et al., 2009). Results from Pacyna et al.  
87 (2016) ranged from a slight increase under a current policy scenario to -85% under a maximum  
88 feasible reduction case. However, these past studies did not consider future changes in marine  
89 and terrestrial Hg reservoirs that affect reemission of  $\text{Hg}^0$  (Amos et al., 2013, 2014). This is  
90 important because terrestrial and marine  $\text{Hg}^0$  evasion are thought to account for large fractions of  
91 the atmospheric Hg undergoing contemporary and future deposition (e.g., ~60% of 2050  
92 deposition to the contiguous United States) (Corbitt et al., 2011).

93 The main objective of this work was to better understand how future anthropogenic Hg  
94 emissions and deposition vary among the most recent socioeconomic development pathways  
95 used by the IPCC. To do this, we developed new decadal projections of primary anthropogenic  
96 Hg emissions for the years 2020-2300 based on four distinct SSPs spanning a wide range of  
97 radiative forcings. Using a suite of global modeling tools, we quantified how changes in primary  
98 Hg emissions and re-emissions from terrestrial and marine ecosystems affect regional  
99 magnitudes and global patterns in atmospheric deposition. This work provides insights into how  
100 different fossil fuel use scenarios are likely to affect global Hg burdens and the potential effects  
101 of different pollution control efforts.

## 102 **2 Methods**

### 103 2.1 Description of Development Narratives (SSP Scenarios)

104 Forecasts of future anthropogenic Hg emissions were developed in accordance with the scenarios  
105 used in Phase 6 of the Coupled Model Intercomparison Project (CMIP6), organized under the  
106 auspices of the IPCC (O'Neill et al., 2016). The scenarios were based on a framework combining  
107 narratives of global development with emissions and climate projections from integrated  
108 assessment and climate models (O'Neill et al., 2016). The first component, the SSPs, are  
109 comprised of five distinct narratives describing alternative courses of societal development, as  
110 well as quantitative descriptions of population, economic growth, and urbanization (Dellink et  
111 al., 2017; Jiang & O'Neill, 2017; Kc & Lutz, 2017). The SSPs were elaborated using integrated  
112 assessment models (IAMs) to provide quantitative descriptions of energy use, greenhouse gas  
113 emissions, and land-use change (Riahi et al., 2017).

114 For each SSP narrative, multiple IAM trajectories were defined to describe both a baseline  
115 scenario, which assumes no additional climate policies or climate change impacts, and mitigation  
116 scenarios, in which further policies are adopted. SSP narratives can be combined with different  
117 climate forcing pathways to describe the physical response of the climate system. The climate  
118 forcing is defined according to the long-term global average radiative forcing ( $\text{W m}^{-2}$ ; Myhre et  
119 al., 2013). Following convention, the scenarios used in this work are written as:  $\text{SSP}_{x-y}$ , where  $x$   
120 is the SSP and  $y$  is the radiative forcing pathway ( $\text{W m}^{-2}$ ) (O'Neill et al., 2016). Throughout the  
121 rest of this work, scenarios will be referred by their specific name (e.g., SSP1-2.6) or collectively  
122 as “SSPs.”

123 For this work, we adopted the three major scenarios that were initially chosen for long-term  
 124 extensions (LTE) to 2300 under CMIP6 (O'Neill et al., 2016), namely, SSPs 1-2.6, 5-3.4, and 5-  
 125 8.5. Scenario SSP1-2.6 is a lower-bound on future emissions, reflecting strong emphasis on  
 126 sustainability and intensive control of climate-forcing agents; SSP5-3.4 is known as the  
 127 “overshoot” scenario and reflects short-term growth in fossil-fuel use and minimal consideration  
 128 of climate control measures until 2040 followed by aggressive mitigation thereafter; SSP5-8.5 is  
 129 an upper bound on emissions, in which fossil-fuel use continues with little consideration of  
 130 climate mitigation or transition to clean technologies. Subsequently, Meinshausen et al. (2020)  
 131 extended additional scenarios to beyond 2300, so we added a fourth scenario to our projections,  
 132 the so-called “middle-of-the-road” scenario, SSP2-4.5 (Fricko et al., 2017).

133 The detailed raw activity data that are used in this work to characterize Hg emissions under these  
 134 four scenarios out to 2100 follow the work of Rao et al. (2017), with LTE from 2100 to 2300  
 135 following the work of Meinshausen et al. (2020). The Hg emissions forecasts reported in this  
 136 work map the trajectories of CO<sub>2</sub> forecasts contained in the previously mentioned publications  
 137 and the emissions of other species as described in subsequent studies (Gidden et al., 2019; Lund  
 138 et al., 2019; Turnock et al., 2020). However, this work contains the first reported projections of  
 139 Hg emissions under CMIP6 scenarios.

140 The SSP scenarios were used primarily to project fossil-fuel use (coal and oil), the manufacture  
 141 and use of refined industrial products, and economic parameters. However, there are no specific  
 142 variables from which to calculate the extraction and production of raw materials and basic  
 143 products because these are not directly relevant to emissions of greenhouse gases. However, they  
 144 are very important as sources of mercury. Thus, for 21<sup>st</sup> century production of these materials,  
 145 we used forecasts specifically generated by industry models, as follows: copper, zinc, and lead  
 146 (Sverdrup et al., 2019); iron (Morfeldt et al., 2015); mercury (Sverdrup & Olafsdottir, 2020); and  
 147 gold (Sverdrup et al., 2012); as well as the basic industrial products steel (Morfeldt et al., 2015)  
 148 and cement (Zhang et al., 2018). The work of Watari et al. (2020, 2021) was valuable in guiding  
 149 the pathways of metals extraction and use to the end of the century.

## 150 2.2 Mercury Emission Calculation

151 Mercury emissions under a future climate scenario ( $f$ ) were calculated using Equation (1):

$$E_{t,r,s,f} = E_{2010,r,s} \times \left( \frac{A_{t,r,s,f}}{A_{2010,r,s,f}} \right) \times \left( \frac{EF_{t,r,s,f}}{EF_{2010,r,s,f}} \right), \#(1)$$

152 where  $E$  = emissions (Mg a<sup>-1</sup>);  $t$  = decadal future year,  $r$  = world region,  $s$  = source type;  $A$  =  
 153 activity level (in various units); and  $EF$  = emission factor (g per unit of activity).

154 Future Hg emissions were calculated by extrapolating 2010 base-year emissions, as reported in  
 155 (Streets et al., 2017, 2019a), in accordance with the energy and activity forecasts for each of the  
 156 four future scenarios ( $f$ ) described in the previous section. Emissions were calculated for each  
 157 decade between 2020 and 2300, though they reach zero between 2190 and 2250 across scenarios  
 158 in accordance with SSP prescriptions. Emissions are zero beyond 2250 across all scenarios.

159 Emissions were calculated at world region level ( $r$ ). The SSP forecasts were developed for five  
160 world regions: OECD, REF (Russia and Eastern Europe), ASIA, MAF (Middle East and Africa),  
161 and LAM (Latin America). This is a very coarse division of the world, which may be adequate  
162 for CO<sub>2</sub> studies, but is too aggregated for studies of Hg emissions and transport. In particular, the  
163 OECD region is spread across the globe, with contributions from Western Europe, North  
164 America, Australasia, and Japan. The prior Hg emission estimates for 2010 (Streets et al., 2019a,  
165 2019b), upon which this work is based, were calculated for 17 world regions and subsequently  
166 aggregated to seven: North America (NAM), South America (SAM), Western Europe (EUR), the  
167 Former Soviet Union (FSU), Africa/Middle East (AFM), Asia (ASA), and Oceania (OCA). By  
168 examining 2010 Hg emissions in the world regions used in each of these two studies, a simple  
169 equivalence was determined and applied in this work, as follows: NAM = 0.5 OECD + 0.1  
170 LAM; SAM = 0.9 LAM; EUR = 0.3 OECD; FSU = REF; AFM = MAF; ASA = ASIA + 0.05  
171 OECD; and OCA = 0.15 OECD. Emissions at the global scale (GLO) are thus identical. Though  
172 this is less than an ideal solution—because of potential future differences in the rates of  
173 development among the countries comprising the coarse SSP regions—it is certainly an  
174 improvement over the five SSP regions from the perspective of estimating Hg transport and  
175 deposition.

176 The SSP scenarios contain more than 600 activity components ( $A$ ), covering all aspects of future  
177 energy, industrial, and economic development. The prior 2010 Hg emission estimates were  
178 developed for 18 source types ( $s$ ) (Streets et al., 2017). Because none of the SSP scenarios  
179 provide activity components that can be unequivocally associated with several of the Hg source  
180 types, emissions from 11 source types were projected individually in this work: mining (copper,  
181 zinc, lead, iron, mercury, gold, and artisanal gold), steel production, cement production, coal  
182 combustion, and oil combustion. Six other source types (municipal waste incineration, other  
183 waste combustion, electrical and measuring equipment, chemicals manufacturing, caustic soda  
184 production, and dental) were projected in aggregate. Silver mining was not included in this work  
185 because emissions from 2010 onwards are expected to be zero.

186 The change in activity levels in the future,  $A_t/A_{2010}$ , reflects the growth or shrinkage in activity  
187 for a particular source type, in a particular region, under a particular scenario. Some may grow,  
188 some may decline, depending on the influence of the world economy in general and the pressure  
189 of the climate change scenario. For example, coal combustion may increase under a lax climate  
190 scenario or decline under a stringent one. But  $A_t/A_{2010}$  only characterizes the change in the size of  
191 the source type, it says nothing about the transformation of it over the time period in question.  
192 New, high-performing technologies will undoubtedly replace older, low-performing ones; and  
193 the imposition of new emission control regulations may or may not force the use of add-on  
194 emission control technologies or even a complete change in production technology. All of these  
195 will influence Hg emissions. Except for Carbon Capture and Storage (CCS) and inferences from  
196 SO<sub>2</sub> emissions, the SSP scenarios say nothing about technology transformation that will  
197 influence Hg emissions. This, therefore, is the most challenging aspect of forecasting future Hg  
198 emissions. Our approach was to develop the ratio  $EF_{t,r,s,f}/EF_{2010,r,s,f}$  as an indicator of how future  
199 emission rates will decline from their 2010 values. Note that this ratio is never greater than one  
200 (i.e., future emission rates are never higher in the future than in 2010 per unit of activity).

201 In previous work, we developed a methodology for estimating the time-development of emission  
202 factors for historical periods (pre-2010) using transformed normal distribution functions (Streets

203 et al., 2011). These region-specific functions were parameterized based on a detailed review of  
204 experimental measurements around the world, as illustrated in that paper for copper smelting. It  
205 was shown, for example, that the Hg emission rate for copper smelters pre-1900 was an  
206 uncontrolled value of 27.5 gHg/MgCu worldwide. After 1900, emission rates declined, along  
207 different pathways in different parts of the world, reaching values in 2010 ranging from a low of  
208 0.60 gHg/MgCu in western Europe to a high of 11.6 gHg/MgCu in Africa.

209 The challenge in this work was to reasonably represent the continuation of historical *EF* trends  
210 out beyond 2010 in the absence of guidance from the SSP forecasts. In this work, we focused on  
211 the period 2010–2100, where technology transformation will have the biggest effect. For each of  
212 the 11 key source types mentioned above, we assumed that by 2100 the emission rates in every  
213 world region will have declined from their 2010 levels to the present-day emission rate of the  
214 lowest-emitting region. Thus, in the case of copper smelting, all regions emit at 0.60 gHg/MgCu  
215 by 2100, though the trajectories to reach it vary by region. This 2100 rate then continues for all  
216 years beyond 2100. Intermediate decadal years in *EF* were determined as linear trends.

217 For the purposes of atmospheric modeling, speciation of these future atmospheric Hg emissions  
218 into elemental Hg ( $\text{Hg}^0$ ) and divalent Hg compounds ( $\text{Hg}^{\text{II}}$ ) was determined in a similar way, by  
219 extrapolation of the speciation splits in each region in the year 2010. These range rather widely,  
220 from a high value of the  $\text{Hg}^0/\text{Hg}$  ratio of 0.88 in South America to a low of 0.25 in western  
221 Europe. This reflects the different emission characteristics of, for example, mining in South  
222 America and industrial manufacturing/well-controlled coal combustion in western Europe. As  
223 time goes on it can be expected that the  $\text{Hg}^0/\text{Hg}$  ratio will decline everywhere, as artisanal and  
224 small-scale mining techniques (with high  $\text{Hg}^0$  emission rates) are retired, and industrial processes  
225 and combustion become increasingly well controlled, leading to conversion of elemental Hg to  
226 collectable, oxidized Hg compounds. In this work, we assumed that the 2010 ratios in high-  
227 emitting regions approach the western European level or the North American level (0.33) by  
228 2100, eventually leveling out at a technology-limiting value of 0.2 everywhere. Slight variations  
229 were applied among the four scenarios to reflect anticipated patterns of future technology  
230 transformations.

231 Finally, future releases of Hg to land and water were estimated following the method of Streets  
232 et al. (2017, 2019a). In essence, this involves subtracting the air emissions from the total Hg  
233 contained in the raw material that is processed. It was not possible to determine *a priori* the fate  
234 of these releases by source type, technology level, or world region. Uncertainties in the few  
235 quantitative estimates that have been made and the vast quantity of unknown factors for sources  
236 in remote parts of the world essentially rule this out.

### 237 2.3 Global Atmospheric Hg Deposition

238 We used the GEOS-Chem global chemical transport model to simulate the atmospheric fate and  
239 deposition of mercury emissions under each emissions scenario. The model version (12.8) used  
240 here included detailed multi-phase oxidation of elemental Hg and gas-phase photolysis of  
241 oxidized Hg species from Shah et al. (2021). Simulations were run using 2014–2019 MERRA-2  
242 meteorology (Gelaro et al., 2017) on a  $2 \times 2.5$ -degree horizontal grid with a 72-layer vertical  
243 domain extending through the top of the stratosphere. The first two years of each simulation

244 were used for initialization and the final three years were averaged for analysis to reduce the  
245 effects of meteorological variability on simulated deposition patterns.

246 We conducted 5-year simulations for each decade from 2020-2100 for each scenario in addition  
247 to a common baseline scenario for the 2010 emission year. We hold meteorology constant across  
248 decadal snapshots to isolate the effects of variation in future emissions. Emissions from the  
249 seven world regions were distributed onto a  $0.25 \times 0.25$ -degree grid based on the spatial  
250 distributions established in prior work (Steenhuisen & Wilson, 2019). Relative fractions of North  
251 American emissions from Canada, the United States, and Central America were scaled to reflect  
252 2015 values reported in Streets et al. (2019b) for consistency with past work.

253 We constructed source-receptor functions to quantify changes in atmospheric Hg deposition  
254 resulting from shifts in terrestrial and oceanic  $\text{Hg}^0$  evasion and for each anthropogenic emission  
255 region. Following Corbitt et al. (2011), we define each source-receptor function  $F_{ij}$  as:

$$F_{ij} = \frac{D_{ij}}{E_i} \#(2)$$

256 where  $D_{ij}$  is the total mercury deposition flux to receptor grid cell  $j$  from emissions in region  $i$ ,  
257 and  $E_i$  is the magnitude of emissions from region  $i$ .

## 258 2.4 Global Biogeochemical Box Model (GBBM)

259 An updated version of the multi-compartment Global Biogeochemical Box Model (GBBM)  
260 developed by Amos et al. (2013; 2014) was used to simulate future shifts in atmospheric  
261 deposition from legacy mercury emissions (Text S1; Table S1-S2). The GBBM represents Hg  
262 cycling between the atmosphere, terrestrial biosphere (fast, slow, protected), ocean (surface,  
263 intermediate, deep), and removal by burial in marine sediment.

264 We added four new compartments to the GBBM that represent waste reservoirs (fast, slow,  
265 protected, and immobilized) to explicitly track the fate of the estimated 1.13 Tg Hg released by  
266 humans to land and water from 1510-2010 (Streets et al., 2019a). These releases included  
267 tailings and waste from mining and metals production, chlor-alkali plants, and substantial  
268 contributions from Hg use in commercial products (Horowitz et al., 2014; Streets et al., 2011). In  
269 prior work, anthropogenic releases to land and water were added to soil compartments of the  
270 GBBM (Streets et al., 2017). However, it is likely that the majority of such releases are  
271 sequestered at contaminated sites and do not have the same diffuse impacts on deposition as  
272 atmospheric sources (Guerrero & Schneider, 2023; Kocman et al., 2017). Therefore, we  
273 independently tracked land and water Hg releases. We parameterized re-emissions of  $\text{Hg}^0$  and  
274 discharges of  $\text{Hg}^{\text{II}}$  to rivers from waste compartments using the same rate coefficients as for the  
275 fast, slow, and protected terrestrial Hg in the GBBM. Land and water Hg releases were  
276 partitioned into waste pools following the methods described in Streets et al. (2017) (Text S1).

277 We simulated the time-dependent fate of future anthropogenic Hg releases separately for each  
278 scenario. In each simulation, the model was initialized by calculating the steady state distribution  
279 of mercury among reservoirs under a constant geogenic Hg flux of  $230 \text{ Mg a}^{-1}$  from subaerial  
280 volcanism and  $50 \text{ Mg a}^{-1}$  from hydrothermal vents (Geyman et al., 2023). The model was then

281 forced with all-time (2000 BCE to 2010 CE) anthropogenic mercury releases from Streets et al.  
282 (2019) followed by emissions specified in each scenario from 2010-2300. Evaluation of model  
283 results for the year 2010 agree with the observational ranges for the atmospheric Hg reservoir,  
284 seawater concentrations, and atmospheric deposition enrichment factors (Table S3, Amos et al.,  
285 2015).

### 286 **3 Results and Discussion**

#### 287 3.1 Future Emission Trends

##### 288 3.1.1 Long-term perspective on anthropogenic Hg releases

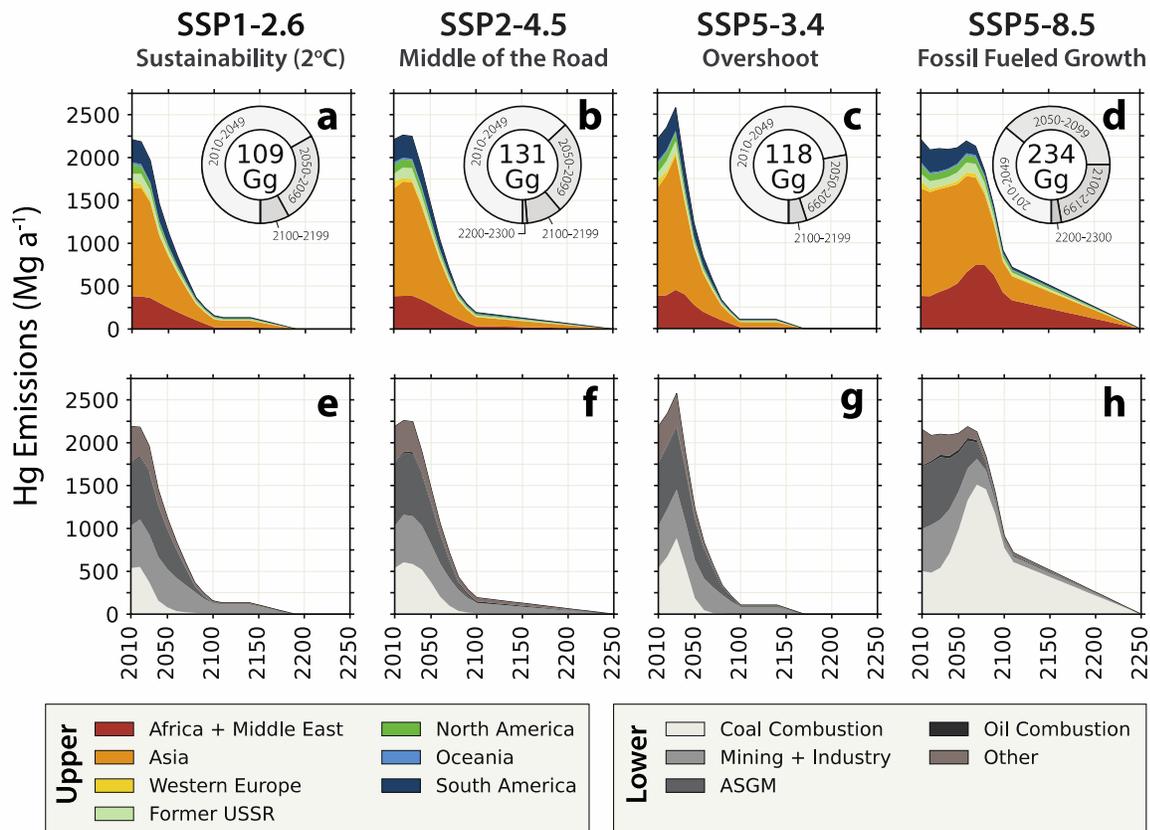
289 Atmospheric Hg emissions are shown by region (Fig. 1; Table S4) and source sector (Table S5)  
290 for each of the four scenarios considered in this work. All primary Hg emissions are projected to  
291 be zero beyond 2250, consistent with SSP assumptions. Projected cumulative Hg releases to air,  
292 land, and water between 2010 and 2300 range from 709 Gg under SSP1-2.6, to 1710 Gg under  
293 SSP5-8.5. Projected anthropogenic emissions between 2010 and 2300 are comparable to  
294 cumulative historical emissions from 1510-2010 (1470 Gg) estimated in past work (Streets et al.,  
295 2019a), suggesting human impacts on the global Hg cycle will be sustained for millennia.

296 Global Hg emissions projected in this work for the year 2050 are substantially lower than similar  
297 forecasts made over a decade prior (2300-4900 Mg a<sup>-1</sup>) using the IPCC SRES scenarios (Streets  
298 et al., 2009). Projected anthropogenic emission declines for the year 2050 under SSP1-2.6 (-1080  
299 Mg a<sup>-1</sup>) are similar to those projected for 2035 by Pacyna et al. (2016) under a scenario involving  
300 full implementation of policy commitments and plans made before 2016. Relative declines  
301 projected in this work are partially attributable to lower Hg emission factors from recently  
302 implemented Hg pollution controls (e.g., Zhang et al., 2023). Thus, lower future Hg emissions  
303 estimated in this work reflect progress made through global policy efforts in recent decades.

304 Across all scenarios, future land and water Hg releases follow qualitatively similar trajectories to  
305 atmospheric emissions. Land and water Hg releases grow as a fraction of total anthropogenic Hg  
306 releases in the future. By 2100, anthropogenic Hg releases to land and water are estimated to be  
307 12 to 25 times greater than anthropogenic Hg emissions to air, compared to only 3.3-fold greater  
308 (7330 Mg a<sup>-1</sup>) in 2010.

309

310



311  
 312 **Figure 1. Global anthropogenic mercury (Hg) emissions to air by world region and source sector.**  
 313 Trajectories under each scenario are shown by world region in Panels (a – d). Inset pie charts show fractional  
 314 emissions occurring during four “snapshot” time periods, arranged clockwise from the bottom: 2010-2049,  
 315 2050-2099, 2100-2200, 2200-2300. Cumulative emissions to air (Gg; 2010-2300) are shown in the center of  
 316 each ring. Panels (e – f) provide a sector-specific breakdown of the same emission. ASGM = artisanal and  
 317 small-scale gold mining.

### 318 3.1.2 Scenario-specific patterns in primary anthropogenic Hg releases

319 Among scenarios considered in this work, SSP1-2.6 is a lower-bound greenhouse gas emissions  
 320 case, and it is known as the “2°C scenario.” SSP1-2.6 is characterized by aggressive reductions  
 321 in greenhouse gas emissions and a nameplate end-century radiative forcing of  $2.6 \text{ W m}^{-2}$   
 322 (Meinshausen et al., 2020). Projected future atmospheric Hg emissions decline continuously after  
 323 2010 (Fig. 1). In 2010, Asia was the largest regional contributor ( $1260 \text{ Mg a}^{-1}$ ; 57% of total) to  
 324 global atmospheric emissions ( $2190 \text{ Mg a}^{-1}$ ). Under SSP1-2.6, a 93% reduction in global Hg  
 325 emissions is projected ( $154 \text{ Mg a}^{-1}$ ) by the year 2100, and Asia remains the largest remaining  
 326 regional emitter ( $91 \text{ Mg a}^{-1}$ , 59%). Beyond 2100, emissions continue to decline at a slower pace  
 327 and reach zero by 2190, in accordance with SSP assumptions.

328 SSP2-4.5 is a middle-of-the-road scenario that reflects moderate socioeconomic, energy, and  
 329 climate mitigation changes and a continuation of growth and development trajectories following  
 330 the status quo (Meinshausen et al., 2020). Relative to 2010, Hg emissions grow slightly until

331 2030 and then decline at a rate similar to SSP1-2.6 (Fig. 1). Hg emissions are  $193 \text{ Mg a}^{-1}$  by  
332 2100, which is only 25% greater than SSP1-2.6 (Fig. 1). SSP2-4.5 is not nearly as close to the  
333 middle of the road for Hg emissions as it is for carbon emissions because it relies heavily on a  
334 switch from coal to natural gas as the primary energy supply under moderate climate mitigation  
335 efforts. This strategy results in continued greenhouse gas emissions and therefore results in  
336 middle of the road climate effects. However, natural gas combustion generates insignificant  
337 quantities of Hg, and so the trajectory is much lower. Hg emissions under SSP2-4.5 continue at a  
338 low level into the distant future, eventually reaching zero in 2250 following SSP assumptions  
339 (Fig. 1). Between 2010 and 2300, cumulative emissions to air are 131 Gg and cumulative  
340 emissions to land and water are 737 Gg under SSP2-4.5.

341 SSP5-3.4 is the overshoot case characterized by a delay in climate-change action until after 2030,  
342 leading to a major rise in Hg emissions between 2010 and 2030. Hg emissions are projected to  
343 rise to  $2580 \text{ Mg a}^{-1}$  by 2030, an increase of 17% over 2010 (Fig. 1). Much of that growth occurs  
344 in Asia. After 2030, emissions decline more precipitously than under the other three scenarios,  
345 reaching levels similar to SSP1-2.6 by 2060. Hg emissions fall to  $107 \text{ Mg a}^{-1}$  by 2100 and reach  
346 zero in the year 2170, which is earlier than SSP1-2.6. Under SSP5-3.4, cumulative emissions to  
347 air are projected to be 118 Gg. Compared to SSP1-2.6, cumulative emissions are heavily  
348 weighted toward the first few decades of the projection period. Cumulative emissions to land and  
349 water are 618 Gg.

350 SSP5-8.5 is the upper-bound case, characterized by continued fossil-fuel use and minimal  
351 consideration of environmental sustainability. Under this scenario,  $\text{CO}_2$  concentrations are  
352 projected to reach levels greater than 2000 ppm by 2200 (Meinshausen et al., 2020). Such levels  
353 have not occurred on Earth since before the onset of the modern Antarctic glaciation over 40  
354 Mya (Rae et al., 2021), and are associated with temperatures  $3.3^\circ\text{C}$  to  $5.7^\circ\text{C}$  higher in 2100 than  
355 during the early-industrial period (1850-1900) (IPCC, 2023). Emissions of Hg to air remain near  
356 present levels until after 2060 and then decline to  $914 \text{ Mg a}^{-1}$  (5-9 times the other scenarios) by  
357 the end of the 21<sup>st</sup> century. The dominant Hg emission regions are Asia, where emissions exceed  
358  $1000 \text{ Mg a}^{-1}$  through 2070, and Africa and the Middle East, where emissions grow from  $383 \text{ Mg}$   
359  $\text{a}^{-1}$  in 2010 to  $744 \text{ Mg a}^{-1}$  in 2080. Emissions decline slowly beyond 2100, remaining at high  
360 levels into the 22<sup>nd</sup> century:  $512 \text{ Mg a}^{-1}$  in 2150 and  $256 \text{ Mg a}^{-1}$  in 2200 (Fig. 1). Emissions do  
361 not reach zero until 2250. Cumulative Hg releases (2010-2300) are 235 Gg to air and 1.47 Tg to  
362 land and water under SSP5-8.5.

363 Several indicators suggest that global anthropogenic Hg emissions are tracking below levels  
364 described in the overshoot scenario (SSP5-3.4) since 2010. While existing emission inventories  
365 report growth in global anthropogenic Hg emissions between 2010 and 2015 (Streets et al.,  
366 2019b), domestic policies have prompted widespread installation of air pollution control devices  
367 over the past decade in China (Zhang et al., 2023) and the United States (Dai et al., 2023).  
368 Trends in  $\text{Hg}^0$  concentration and isotopic composition from long-term monitoring sites in China  
369 also show declines consistent with suggested reductions in regional anthropogenic emissions  
370 (Wu et al., 2023).

371 3.1.3 Decoupling of Hg emissions and radiative forcing under low coal use scenarios

372 Among source sectors in 2010, artisanal and small-scale gold mining (ASGM) was the largest  
373 global source of atmospheric emissions (727 Mg a<sup>-1</sup>), followed by coal combustion (538 Mg a<sup>-1</sup>)  
374 and mining and industry (491 Mg a<sup>-1</sup>). Large-scale (rather than ASGM) gold production was the  
375 largest source of Hg released to land and water (1990 Mg a<sup>-1</sup>), followed by ASGM (1090 Mg a<sup>-1</sup>),  
376 electrical and measurement equipment (1010 Mg a<sup>-1</sup>), chemicals manufacturing (860 Mg a<sup>-1</sup>),  
377 and zinc smelting (670 Mg a<sup>-1</sup>). Coal combustion accounted for a greater fraction of global  
378 atmospheric emissions (25%) in 2010 compared to land and water releases (5%, 370 Mg a<sup>-1</sup>).

379 Differences in anthropogenic Hg emissions to air principally reflect differences in coal  
380 combustion across SSP scenarios. Hg emissions from coal combustion are projected to grow by  
381 2.3 to 12.5% and peak in 2020 under SSPs 1-2.6 and 2-4.5 (Fig. 1). Near-term increases in Hg  
382 emissions from coal combustion reach levels 65% higher than 2010 under SSP5-3.4. Hg  
383 emissions decline thereafter for lower coal-use scenarios (SSPs 1-2.6, 2-4.5 and 5-3.4). By 2050,  
384 they fall to a fraction of 2010 levels (15% under SSP1-2.6 and 69% under SSP2-4.5) and reach  
385 zero before the end of the century. In contrast, coal combustion persists until 2250 under SSP5-  
386 8.5. Coal-related Hg emissions under SSP5-8.5 reach 990 Mg a<sup>-1</sup> (96% higher than 2010) by  
387 2070 and remain 44% higher than 2010 levels by the end of the century (Fig. 1).

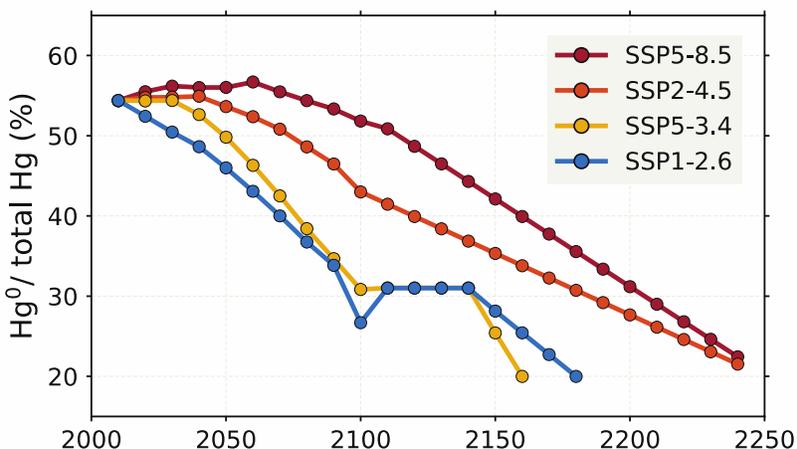
388 Among lower coal-use scenarios, coal combustion is responsible for 16 Gg (SSP1-2.6) to 28 Gg  
389 (SSP2-4.5) of cumulative (2010-2300) Hg emissions to air. Cumulative Hg emissions to air from  
390 coal combustion are ~5 to 9-fold greater under SSP5-8.5 (139 Gg) than the other scenarios, and  
391 coal combustion comprises 59% of all emissions to air between 2010 and 2300 under SSP5-8.5.

392 Despite relative similarities in the phase-out of coal, SSPs 1-2.6, 2-4.5, and 5-3.4 show very  
393 different greenhouse gas emission trajectories. Such differences arise from rates of natural gas  
394 and oil combustion, as well as adoption of carbon capture and sequestration. These factors exert  
395 minimal direct influence on primary anthropogenic Hg emissions, though they produce disparate  
396 climate effects. For example, end-century surface temperatures simulated with the NASA GISS-  
397 E2.1 climate model are 1.8°C – 2.3°C higher than the preindustrial (1850-1880) mean under  
398 SSP1-2.6 compared to 2.7°C – 3.3°C under SSP2-4.5 (Nazarenko et al., 2022). The degree of  
399 warming will modulate future changes in Hg emissions from the natural biosphere (e.g.,  
400 Krabbenhoft & Sunderland, 2013; Schaefer et al., 2020). Therefore, it is important to consider  
401 the consequences of human activity for both direct anthropogenic Hg releases and warming-  
402 driven changes in Hg cycling in the biosphere and ocean (e.g., Schaefer et al., 2020; Schartup et  
403 al., 2019).

#### 404 3.1.4 Changes in Hg emission speciation favor local and regional deposition

405 The fraction of primary anthropogenic Hg emissions released to air as Hg<sup>0</sup> is projected to decline  
406 in the future (Fig. 2), with implications for transboundary pollution. For SSP1-2.6, the fraction of  
407 Hg<sup>0</sup> emitted by primary sources is projected to decline by 28% between 2010 and 2100. It  
408 stabilizes from 2100-2140, and then declines continuously thereafter. By 2180, the last decade  
409 with non-zero emissions for SSP1-2.6, Hg<sup>0</sup> is projected to make up just 20% of total  
410 anthropogenic Hg emissions to air (Fig. 2). Emission speciation under SSP5-3.4 broadly follows  
411 that of SSP1-2.6, with slightly higher Hg<sup>0</sup> fractions through 2100, and an accelerated decline  
412 between 2140 and 2160 from 31% Hg<sup>0</sup> to 20% Hg<sup>0</sup> (Fig. 2). Under SSP2-4.5, Hg<sup>0</sup> comprises a  
413 larger fraction of total Hg emissions to air than under SSPs 1-2.6 and 5-3.4, remaining around

414 55% until 2040 (Fig. 2). The  $\text{Hg}^0$  fraction then exhibits a slow decline with overall emissions and  
 415 reaches 22% by 2240 (Fig. 2). SSP5-8.5 consistently represents an upper bound for the  $\text{Hg}^0$   
 416 fraction, increasing to 57% in 2060, followed by near-continuous declines to reach 22% by 2240.  
 417 While trajectories vary across scenarios, greater fractions of  $\text{Hg}^{\text{II}}$  relative to  $\text{Hg}^0$  in primary  
 418 anthropogenic emissions are expected to produce future deposition patterns from primary  
 419 emissions that increasingly reflect local and regional rather than global sources (Fig. S1).



420  
 421 **Figure 2. Decadal changes in the speciation of primary anthropogenic mercury (Hg) emissions to air.**  
 422 The percentage of total Hg emissions to air released as elemental Hg ( $\text{Hg}^0$ ) is shown for each scenario and the  
 423 remaining fraction consists of divalent mercury ( $\text{Hg}^{\text{II}}$ ).

### 424 3.2. Future Deposition Patterns from Primary Anthropogenic Emissions

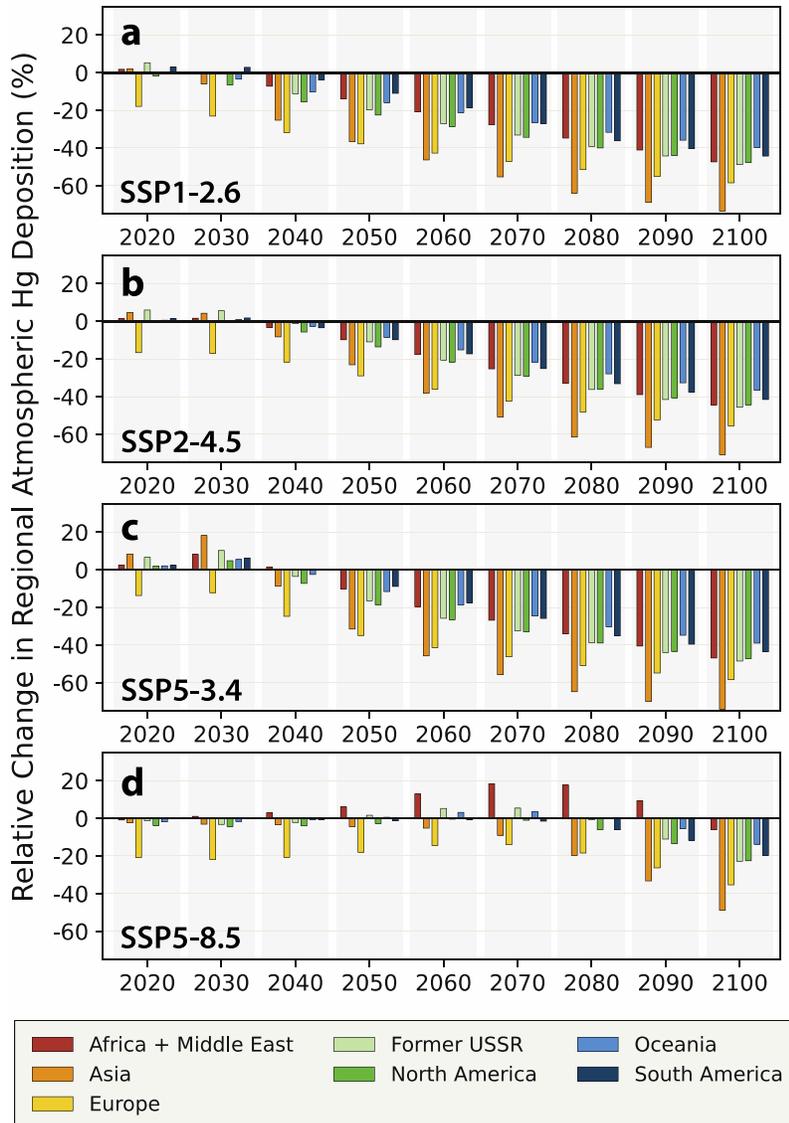
425 Modeled global Hg deposition to land increases by  $22 \text{ Mg a}^{-1}$  (0.9%) between 2010 and 2020  
 426 under SSP1-2.6 emissions. This reflects increases from both primary anthropogenic emissions  
 427 ( $+8 \text{ Mg a}^{-1}$ ) and higher re-emissions from terrestrial and ocean surfaces due to increasing Hg  
 428 reservoirs ( $+14 \text{ Mg a}^{-1}$ ) (Fig. S2). Modeled increases in regional Hg deposition occur during this  
 429 period over all global regions except for Europe and North America (Fig. 3a; 4a). There, declines  
 430 in deposition from primary anthropogenic sources exceed increases in deposition from growing  
 431 global terrestrial and oceanic Hg reservoirs and subsequent re-emissions. From 2040 through the  
 432 end of the century, declines in atmospheric Hg deposition are projected for all regions. By 2100,  
 433 total atmospheric Hg deposition is projected to be less than half of 2010 levels ( $46\%$ ,  $1120 \text{ Mg a}^{-1}$ )  
 434 (Fig. S2). The largest regional deposition declines are projected over Asia ( $-510 \text{ Mg a}^{-1}$ ;  $-$   
 435  $73\%$ ) because it was the largest source region in 2010 (Fig. 3a).

436 Temporal patterns in atmospheric Hg deposition under SSP2-4.5 are qualitatively similar to  
 437 those of SSP1-2.6, characterized by slight but regionally heterogeneous near-term increases that  
 438 give way to continuous decreases. Relative to 2010, global deposition increases by 1.7% in 2020  
 439 and 1.6% in 2030. These increases occur over all regions except Europe, where declines are  
 440 smaller than for SSP1-2.6 ( $-17\%$  in 2020 and 2030) (Fig. 3b). By the end of the century,  
 441 deposition to land is  $1270 \text{ Mg a}^{-1}$  lower than in 2010 (Fig. S2), with regional declines reaching  
 442  $71\%$  ( $-25 \mu\text{g m}^{-2} \text{ a}^{-1}$ ) in Asia (Fig. 3b).

443 Under SSP5-3.4, near-term deposition increases considerably, growing to 2690 Mg a<sup>-1</sup> (+9%) in  
444 2030 over land (Fig. S2). The greatest increases occur in Asia, where total Hg deposition reaches  
445 41 μg m<sup>-2</sup> a<sup>-1</sup> (+18%) in 2030. Deposition grows by greater than 8% over all regions except  
446 Europe. After 2030, accelerated emission reductions relative to SSP1-2.6 result in comparable  
447 end-century deposition, which is 1340 Mg a<sup>-1</sup> lower than in 2010 (Fig. S2).

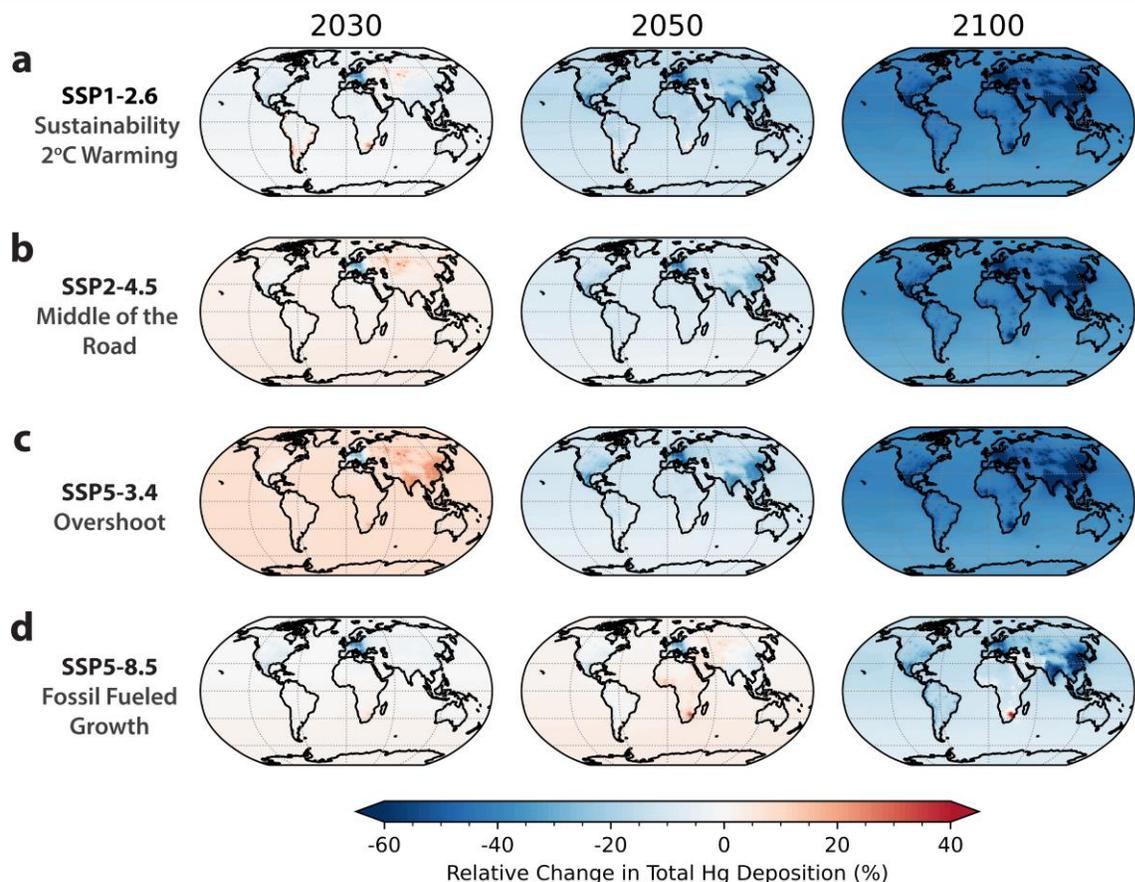
448 Atmospheric deposition under SSP5-8.5 is much higher than the other scenarios through most of  
449 the century. However, 2020 deposition is the lowest of all scenarios, and 2030 deposition is  
450 lower than all scenarios other than SSP1-2.6 (Fig. S2). Growth in emissions from Africa and the  
451 Middle East, combined with sustained emissions elsewhere, produce increasing deposition over  
452 most regions from 2020 until 2070. Most notably, deposition to Africa and the Middle East reach  
453 levels 18% higher (+2.5 μg m<sup>-2</sup> a<sup>-1</sup>) by 2070 and remain above baseline levels through 2100 (Fig.  
454 3d; Fig. S2). End-century total deposition to land is 650 Mg a<sup>-1</sup> lower than in 2010.

455



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457  
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460

**Figure 3. Relative changes in regional atmospheric mercury (Hg) deposition compared to 2010.** Changes in atmospheric Hg deposition are shown by world region (represented by colored bars) for each decade from 2020 to 2100 (x-axis). Each subplot represents temporal trends under a different Shared Socioeconomic Pathway (SSP) scenario (O'Neill et al., 2016).

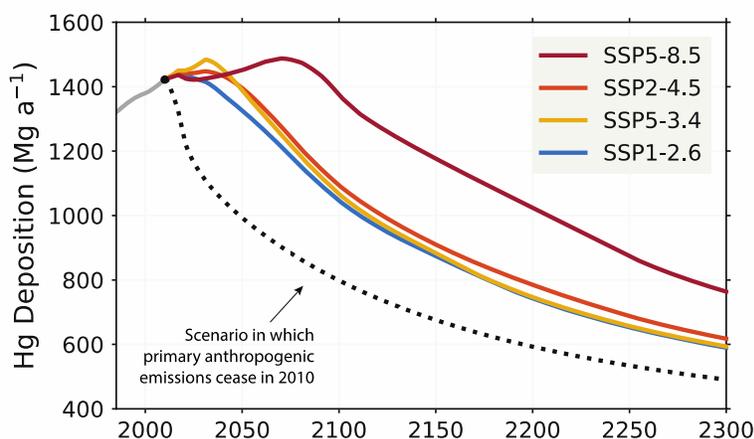


461  
 462 **Figure 4. Trajectories of global atmospheric mercury (Hg) deposition.** Panels represent fractional change  
 463 in total deposition relative to the 2010 baseline by scenario (rows) for three time periods: 2030 (left), 2050  
 464 (center) and 2100 (right). Deposition is calculated as the sum of deposition from primary, legacy, and natural  
 465 Hg emissions using the GEOS-Chem atmospheric mercury model (Shah et al., 2021) and the Global  
 466 Biogeochemical Box Model (GBBM; Amos et al., 2013, 2014).

### 467 3.3 Implications for Legacy Hg Deposition from Terrestrial and Aquatic Emissions

468 Scenarios for deposition of legacy Hg vary among the four SSPs. Legacy deposition peaks  
 469 before 2035 at levels 20 – 62 Mg a<sup>-1</sup> higher than 2010 under SSPs 1-2.6, 2-4.5, and 5-3.4 (Fig.  
 470 5). Legacy deposition under SSP5-8.5 exhibits a larger and later peak, exceeding 2010 deposition  
 471 by 66 Mg a<sup>-1</sup> in 2071 (Fig. 5). By 2100, relative trends in legacy deposition are notably different,  
 472 with declines of 330 – 380 Mg a<sup>-1</sup> relative to 2010 for SSPs 1-2.6, 2-4.5 and 5-3.4, compared  
 473 with a decline of 47 Mg a<sup>-1</sup> for SSP5-8.5 (Fig. 5). These end-century declines in legacy  
 474 emissions are responsible for 26 – 28% of total declines in atmospheric Hg deposition to land  
 475 under SSPs 1-2.6, 2-4.5, and 5-3.4, compared to 7% for SSP5-8.5.

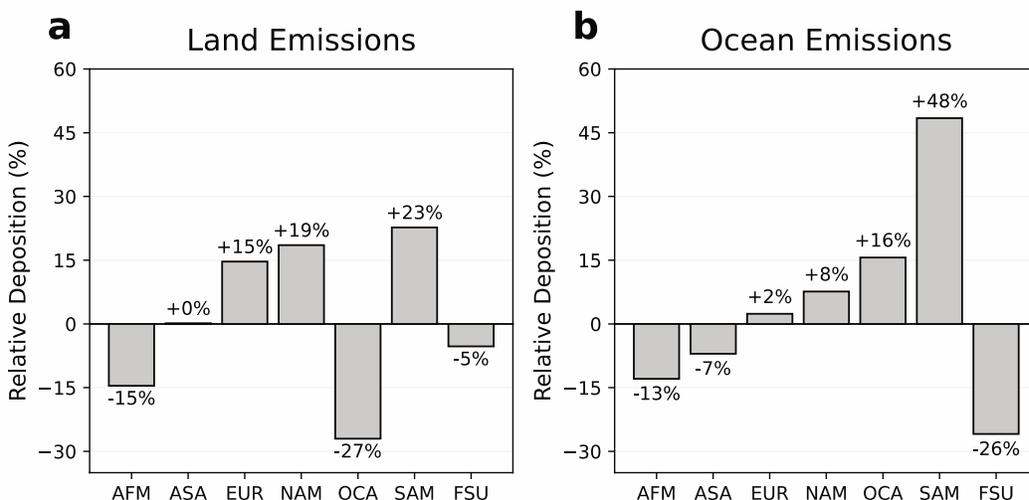
476



477  
 478 **Figure 5. Mercury (Hg) deposition to land from legacy and natural emissions.** Trajectories under future  
 479 emissions (2010-2300) are shown for SSPs 1-2.6 (blue line), 2-4.5 (orange line), 5-3.4 (yellow line), and 5-8.5  
 480 (red line), in addition to a hypothetical scenario where primary anthropogenic emissions are zero after 2010  
 481 (dotted black line). Implications of future trajectories for upper ocean reservoirs are shown in Figure S4.

482 Comparing SSPs 1-2.6 and 5-3.4 provides insights into the consequences of delaying emission  
 483 reductions. Cumulative emissions through 2100 are comparable for SSP1-2.6 (101 Gg) and  
 484 SSP5-3.4 (112 Gg). Additionally, anthropogenic Hg emissions are lower in 2100 under SSP5-3.4  
 485 than SSP1-2.6 (107 Mg a<sup>-1</sup> and 154 Mg a<sup>-1</sup>). Higher deposition to anthropogenic receptor regions  
 486 in 2100 under SSP5-3.4 reflects a legacy deposition penalty for the unrestrained growth of the  
 487 first three decades of the century. Deposition from legacy emissions is 20 Mg higher in 2100 for  
 488 SSP5-3.4 compared to SSP1-2.6, whereas deposition from primary anthropogenic emissions is  
 489 16 Mg lower. This example demonstrates the long-term benefits associated with near-term  
 490 emissions mitigation, as discussed previously in Angot et al. (2018) and Amos et al. (2013).

491 By the end of the century, Hg re-emissions from the ocean and land will become relatively more  
 492 important as sources of Hg deposition (Fig. S3), but deposition will not be evenly distributed by  
 493 world region. Our source-receptor modeling suggests that South America receives the largest  
 494 share of legacy Hg deposition on a per-area basis, receiving 23% higher deposition per unit of  
 495 terrestrial emissions and 48% higher deposition per unit of ocean evasion than the area-weighted  
 496 average of all world regions (Fig. 6). Such high rates of deposition over South America are  
 497 driven by high rates of foliar uptake and wet deposition. In contrast, Oceania receives the lowest  
 498 areal share of terrestrial emissions (73% of average), and the Former USSR receives the smallest  
 499 areal share of Hg sourced from oceanic evasion (74% of average) (Fig. 6). Such low deposition  
 500 is driven by the relative isolation of Oceania from terrestrial emissions and of the Former USSR  
 501 from ocean emissions. These source-receptor relationships are subject to change in the future due  
 502 to shifting patterns of historical anthropogenic Hg loading (e.g., Zolkos et al., 2022) and  
 503 changing biogeochemical dynamics mediating deposition (e.g., Alexander & Mickley, 2015;  
 504 Krabbenhoft & Sunderland, 2013; Yang et al., 2019).



505  
 506 **Figure 6. Regional differences in area-normalized atmospheric Hg deposition from natural and legacy**  
 507 **sources.** Bars represent differences in areal deposition rates between individual receptor regions and the  
 508 average over land. Positive values mean that receptor regions receive greater deposition per unit area than  
 509 average, and negative values mean that receptor regions receive less deposition than average. Individual panels  
 510 show trends in deposition resulting from emissions from the terrestrial biosphere (a) and from the ocean (b).  
 511 Receptor regions are Africa and the Middle East (AFM), Asia (ASA), Europe (EUR), North America (NAM),  
 512 Oceania (OCA), South America (SAM), and the Former Soviet Union (FSU).

513 Globally averaged seawater Hg concentrations in the upper ocean (0-1500 m) are projected to  
 514 increase over the coming decades across all scenarios in this study. The Hg reservoir in the upper  
 515 ocean peaks at 142 to 151 Gg between 2037 (SSP1-2.6) and 2081 (SSP5-8.5) (Fig. S4). Near-  
 516 term increases in upper ocean Hg concentrations are driven by future rather than historical  
 517 emissions. Simulated seawater Hg concentrations begin declining in 2015 under a scenario with  
 518 no future primary anthropogenic Hg releases (Fig. S4). By the end of the 21<sup>st</sup> century, upper  
 519 ocean Hg concentrations are projected to be 15% lower than 2010 under SSP1-2.6 and are 11%  
 520 higher than 2010 under SSP5-8.5.

#### 521 4 Conclusions

522 The SSP scenarios evaluated in this work result in a greater than two-fold difference in  
 523 cumulative anthropogenic Hg emissions between 2010 and 2300. Cumulative anthropogenic  
 524 emissions to air and releases to land and water between 2010 and 2300 range from 710 Gg under  
 525 the low-bound scenario (SSP1-2.6) to 1710 Gg under the upper-bound scenario (SSP5-8.5).  
 526 These future releases are comparable to all-time historical anthropogenic emissions of 1540 Gg  
 527 (80% CI: 1060 – 2800 Gg) (Streets et al., 2019a).

528 Transition of the energy sector away from coal combustion is the largest determinant of  
 529 differences among scenarios, with lower bound and mid-range scenarios (SSPs 1-2.6, 2-4.5, 5-  
 530 3.4) all exhibiting similar cumulative emissions due to declining coal usage. By contrast,  
 531 industrial Hg mining and ASGM releases are projected to grow in relative importance in the  
 532 future. These results imply that under the most stringent climate policies, the largest sources of  
 533 Hg and CO<sub>2</sub> are likely to become more distinct from one another.

534 Numerous factors may affect regional Hg deposition patterns in the future. For most regions,  
535 reducing primary anthropogenic Hg emissions remains the most effective way to reduce  
536 deposition, since 55 – 71% of Hg<sup>II</sup> emissions redeposit to the region of origin, compared to 5 –  
537 13% for Hg<sup>0</sup> (Fig. S5). However, the scenario-based trajectories described in this work indicate  
538 that anthropogenic Hg<sup>II</sup> emissions to air will decline more slowly than Hg<sup>0</sup>, causing Hg<sup>II</sup> to grow  
539 as a fraction of total anthropogenic Hg emissions to air. As the speciation of anthropogenic Hg  
540 emissions shifts towards lower fractions of Hg<sup>0</sup>, a greater proportion of regional emissions will  
541 redeposit to the region of origin. Across all regions, the fraction of self-sourced atmospheric Hg  
542 deposition is projected to increase from 33% in 2010 to 35% – 45% in 2100 (Fig. S1).

543 Regional atmospheric Hg deposition patterns over the next century reflect trends in the  
544 magnitude and spatial distribution of primary anthropogenic emissions as well as re-emissions  
545 from the land and ocean. In regions where anthropogenic Hg emissions reductions are projected  
546 to occur, total Hg deposition is expected to decrease, even during periods where global  
547 deposition is increasing. This means that domestic policies have significant leverage on domestic  
548 Hg deposition (e.g., Dai et al., 2023). However, greater reductions in anthropogenic emissions  
549 are accompanied by larger declines in emissions from the land and ocean. Under SSP1-2.6, 28%  
550 of the reduction in atmospheric deposition from 2010 to 2100 was the result of declines in legacy  
551 emissions. As a result, globally-coordinated efforts to reduce near-term anthropogenic Hg  
552 emissions will produce amplified benefits in terms of long-term declines in atmospheric  
553 deposition (e.g., Angot et al., 2018).

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558 used in this work.

## 559 **Open Research**

560 Mercury emission files, code and model output are available in an OSF repository (peer review  
561 link provided with submission). The version of GEOS-Chem used in this study is available from  
562 (<https://doi.org/10.5281/zenodo.3784796>).

563

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