

Characterising the response of vegetation cover to water limitation in Africa using geostationary satellites

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Key Points:

- We provide observation based metrics from FVC time series over Africa, characterising the dynamics of vegetation during water limitation
- The metrics, derived from daily FVC data with 0.0417°, have strong diagnostic power to understand fine-scale vegetation–water interactions
- Focused on water-limited periods, the metrics can be used to test hypothesis and constrain models on highly uncertain processes

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Abstract

Plant available water is a key driver of ecosystem processes in water-limited systems. The interactions between vegetation, soil moisture, groundwater, and lateral redistribution of moisture in landscapes are complex and very heterogeneous. This complexity, together with the scarcity of relevant observations, creates a major obstacle for large-scale ecohydrological analysis and modelling. Here we exploit recent advancements in remote sensing at high spatial and temporal resolutions to extract relevant information on ecohydrological functioning. Our approach focuses on characterising vegetation dynamics along the seasonal wet to dry season transition, i.e. with progressive water limitation.

We present a set of observation-based metrics to characterise ecohydrological patterns across Africa at 0.0417° spatial resolution. These are derived from the daily time series of Fraction of Vegetation Cover (FVC) over the period 2004–2019 from the geostationary satellite Meteosat Second Generation. The metrics include (i) minimum and maximum FVC, (ii) start day, duration, and FVC integral of the dry season, and (iii) the decay rate of FVC during dry-down. The metrics reflect the potential state, temporal extent, and evolution of the limiting factors of FVC, which, in Africa, are predominantly associated with water availability. They provide information on the relevance of secondary moisture sources such as ground water access or ecohydrological buffering due to deep rooting. Analysis of the metrics reveals large-scale gradients with aridity, as well as regional patterns associated with topographic moisture variations. Our observation-based products have large potential for better understanding and modelling the complex vegetation-water interactions from regional to continental scales.

Plain Language Summary

Despite their importance on global carbon and water cycles, together with the ecosystem services, local-scale processes controlling vegetation dynamics under water-limitation are highly uncertain in large scale studies. This is particularly important in Africa due to the scarcity of ground measurements and stronger dependency of population on ecosystem services. In order to overcome this problem, we developed a set of metrics based on the fractional vegetation cover observed from the European geostationary satellite with daily temporal resolution. The metrics are suitable to diagnose local-scale processes thanks to their high spatial resolution of ~ 5 km. First analyses show consistent continental gradients in the metrics together with strong local variations and corroboration with different datasets from independent sources.

1 Introduction

Africa hosts the largest share of undernourished population, and livelihood of the majority of population relies on ecosystem services, ecosystem productivity and water availability (Müller et al., 2014). African ecosystems contribute strongly to variations in the global carbon cycle (Williams et al., 2007; Valentini et al., 2014; Palmer et al., 2019; Weber et al., 2009), in which large uncertainties remain due to limited observations to model complex ecohydrological interactions happening with a wide spectrum over Africa. Most of the ecosystems in Africa are clearly controlled by water availability not only in arid and semiarid regions, but also in tropical forest in Central Africa (Zhou et al., 2014; Guan et al., 2015). Therefore, understanding the vegetation–water interactions in Africa is crucial.

Vegetation access to water is driven by rainfall but modulated by the interplay among hillslope topography (Fan et al., 2019), soil properties, groundwater (Maxwell & Condon, 2016), and root traits (Maeght et al., 2013). Meanwhile ecosystem properties controlling water use are likely adapted to climate and local hydrological conditions (Gentine et al., 2012). Representation of such complex and fine-scale interactions between veg-

66 etation, soil moisture, and groundwater within diverse landscapes still poses a challenge
 67 for land surface modellers (Clark et al., 2015; Fisher & Koven, 2020), and is hampered
 68 by the scarcity of observational constraints, especially for Africa.

69 Analysis of remotely-sensed vegetation indices provides opportunities to infer con-
 70 trolling environmental factors and land surface characteristics over large spatial domains.
 71 As most of the ecosystems in Africa are subject to moisture limitations, patterns of veg-
 72 etation indices from remote sensing retrievals can be a proxy of underlying ecohydrolog-
 73 ical processes. In addition, annual recurrence of distinct wet and dry seasons over Africa
 74 provides a natural test bed to infer the effects of progressive water limitation and sec-
 75 ondary water resources.

76 Characterising vegetation dynamics during dry season transition is challenging. Clas-
 77 sical phenological metrics (reviewed in Zeng et al., 2020), such as the start and the end
 78 of the growing season, developed from the perspective of energy limited ecosystems. How-
 79 ever, they are not tailored to the particularities of water-limited systems, i.e., vegeta-
 80 tion with the ability to access secondary water resources in hillslope scales. In contrast
 81 to temperature or radiation, soil moisture has a strong memory with a gradual decline
 82 of water over the dry season. Besides, the end of growing season varies with vegetation
 83 properties, e.g., rooting depth, and moisture storage capacity depending on climate, soil
 84 and topographic characteristics. As such, characteristics of vegetation dry-down reflects
 85 the underlying ecohydrological processes and the limiting factors.

86 In this study, we characterise the dynamics of vegetation in the water-limiting pe-
 87 riod, i.e., the dry season, instead of those in the growing season. Here, the dry season
 88 is defined as the time period from the start of the effects of water limitations on vege-
 89 tation cover, i.e., the peak of growing season, to the end of water limitation, i.e., the on-
 90 set of the next growing season (see Fig. 1). The definition takes a vegetation perspec-
 91 tive and complements the more traditional approaches using atmospheric forcing like pre-
 92 cipitation thresholds. Furthermore, distinct temporal features of vegetation dynamics
 93 in the dry season provide indications of ecohydrological properties directly relevant to
 94 vegetation, such as ecosystem water storage capacity, access to secondary water resources
 95 due to groundwater or topographic moisture convergence, and ecosystem water use ef-
 96 ficiency during photosynthesis. Undoubtedly, mapping such ecohydrological character-
 97 istics in space facilitates a better understanding, and subsequent modelling of vegeta-
 98 tion-water interactions. We are aware that many factors influence temporal dynamics of ve-
 99 getation cover and that the potential attribution to ecohydrological phenomena is con-
 100 tingent on the presence of predominately water-limited ecosystems such as over most of
 101 Africa.

102 We provide a set of relevant ecohydrological metrics for the African continent at
 103 $\sim 0.04^\circ$ spatial resolution. As mentioned previously, these metrics were derived using
 104 the vegetation dynamics in the decay phase in the time series. It should here be noted
 105 that the vegetation decay in Africa is mostly associated with water availability. Through-
 106 out this manuscript, we, therefore, use decay interchangeably with dry season, and as-
 107 sociate both of them with vegetation dynamics under water limitation. The ecohydro-
 108 logical metrics were derived from the daily Fraction of Vegetation Cover (FVC) from geo-
 109 stationary satellite observations (see Sec. 2). The set of derived metrics (illustrated in
 110 Fig. 1) encompasses:

- 111 1. the asymptotic values for minimum and maximum of the FVC,
- 112 2. start day and duration of the dry season and the integral of FVC in the dry sea-
 113 son, and
- 114 3. the exponential decay rate of FVC during dry-down.

115 The spatial patterns in minimum FVC, hereafter FVC_{min} , are likely related to the
 116 minimum amount of plant available water that would support vegetation activity via sec-

117 ondry water resources, i.e., deeper soil moisture and/or groundwater. The maximum
 118 FVC may be indicative of the maximum plant accessible water; and together with FVC_{min}
 119 allows for assessing the seasonal changes in vegetation (associated with water limitation).
 120 Since the cumulative water stress is expected to shape plant adaptation to prevailing wa-
 121 ter conditions (Caylor et al., 2009; Good & Caylor, 2011), the identification of the start
 122 and duration of the dry season is fundamental for ecohydrology. The integral of FVC
 123 over the dry season essentially diagnoses the total vegetation activity during the dry sea-
 124 son, and is thus indicative of the vegetations dry season water consumption. The inte-
 125 gral can be used to diagnose ecosystems’ buffering capacity for progressive water lim-
 126 itation during dry season, say, due to deep root distribution.

127 The time scale of FVC decay is estimated for dry-down events, i.e., when ecosys-
 128 tems are predominantly water-limited. To do so, we assume that the plant available wa-
 129 ter is a single-pool linear storage reservoir with an exponential decline over time. Such
 130 assumption has been previously applied to satellite retrievals of surface soil moisture, show-
 131 ing associations of the decay with soil texture and aridity (McColl et al., 2017). Simi-
 132 larly, the e -folding time of evapotranspiration observed from eddy covariance flux tow-
 133 ers revealed patterns of associations with plant height and seasonal aridity (Teuling et
 134 al., 2006; Boese et al., 2019; Martínez-de la Torre et al., 2019). In order to satisfy the
 135 single-pool linear storage model, we considered only the convex part of the vegetation
 136 decay and referred to it as dry-down. In summary, the e -folding time of FVC is an emer-
 137 gent ecohydrological signature of the complex interactions between vegetation, climate,
 138 soil, and possibly groundwater.

139 To derive the ecohydrological metrics for the African continent from high-resolution
 140 remote sensing data (Sec. 2), we developed of a robust methodology (Sec. 3) to deal with
 141 noise, gaps, widely varying dynamics, and data size. The quality diagnostics along with
 142 the derived metrics (Sec. 4), and open code for derivations, enables future advances in
 143 understanding and modelling ecohydrological processes and variability. Initial analysis
 144 and corroboration with independent data illustrates the potential of applications of the
 145 ecohydrological metrics (Sec. 5).

146 2 Data

147 2.1 Fraction of Vegetation Cover

148 The FVC, derived from a spectral mixture analysis of the satellite retrievals, is a
 149 vegetation index summarising the coverage ratio of vegetation per unit total land area
 150 within a grid cell (Trigo et al., 2011). With a range of 0–1, FVC is often used to derive
 151 fundamental vegetation indices such as the Leaf Area Index. The FVC product used in
 152 this study was obtained from the Satellite Application Facility for Land Surface Anal-
 153 ysis (LSA-SAF) of the European Organisation for the Exploitation of Meteorological Satel-
 154 lites (EUMETSAT). The product is based on the retrievals of the Spinning Enhanced
 155 Visible and Infrared Imager (SEVIRI) sensor on board the Meteosat Second Generation
 156 (MSG) satellite (Trigo et al., 2011). As a geostationary satellite, the MSG has a circular
 157 spatial coverage of Earth centred at 0° longitude, and it covers Europe and Africa
 158 entirely (see an example of the original FVC data for a day in Fig. A1). The SEVIRI
 159 is a multispectral optical sensor with 12 spectral bands, and a temporal resolution of 15
 160 minutes. Under the sub-satellite point (nadir), it has 3.1 km spatial resolution in the nor-
 161 mal bands, and a high-resolution band with 1 km spatial resolution. The spatial reso-
 162 lution of the retrieval decreases with distance from the nadir, as for all geostationary satel-
 163 lites. The FVC data product is available at daily temporal resolution spanning the time
 164 period from early 2004 to present. The FVC product, as well as its complete details, are
 165 available at <https://landsaf.ipma.pt/en/products/vegetation/fvc/> .

166 For this study, we selected the spatial domain as the African continent. We resam-
 167 pled the original data to a spatial resolution of 0.0417° (~ 5 km) with the nearest neigh-
 168 bour method (using `gdalwarp` function in GDAL (GDAL/OGR contributors, 2020)). In
 169 terms of temporal domain, we used nearly 16 years of data, from the beginning of the
 170 records in 2004, to the end of 2019.

171 2.2 Ancillary data

172 Climate:

173 To characterise major climate conditions, we used the Köppen–Geiger climate clas-
 174 sification data (Rubel & Kottek, 2010) which is available at 0.0833° spatial resolution.
 175 For the sake of interpretability, we simplified the original climate classes into 6 major
 176 climate groups: arid desert (*BW*), arid steppe (*BS*), tropical humid (*Af* & *Am*), trop-
 177 ical with dry season (*As* & *Aw*), temperate humid (*Cf*), and temperate with dry sea-
 178 son (*Cs* & *Cw*). A small number of grid cells with continental (*D*) or polar (*E*) climates
 179 around Mount Kilimanjaro were discarded. A map of the simplified climate classes can
 180 be found in Fig. D1.

181 Accessible water storage capacity and rooting depth:

182 We compared the integral metric (I_{ds}) against other proxies of plant accessible wa-
 183 ter. For that, we selected the rooting depth and plant available water storage capacity
 184 data from previous studies. For rooting depth, we used two data: Yang et al. (2016), at
 185 0.5° spatial resolution, derived from a carbon cost–benefit model, and potential rooting
 186 depth data from Fan et al. (2017), at 0.0083° spatial resolution, derived with a plant adap-
 187 tion perspective via inverse modelling of root water uptake. For water storage capacity,
 188 we again used two datasets based on hydrological or land surface models: data from Tian
 189 et al. (2019) at 0.25° spatial resolution, and that from Wang-Erlandsson et al. (2016)
 190 at 0.5° spatial resolution. For a consistent comparison across data at different resolu-
 191 tions, we aggregated all data to a common spatial resolution of 0.5° by simple averag-
 192 ing. Note that the spatial aggregation may result in the loss of the spatial variability preva-
 193 lent locally and potentially captured at a high resolution.

194 Topography:

195 To relate the variation of the metrics with local-scale heterogeneity and convergence
 196 of moisture caused by topography, we used the Height Above Nearest Drainage (HAND)
 197 data from Yamazaki et al. (2019). The HAND is a normalised metric derived from to-
 198 pography that is closely related with drainage topology and potential local-scale conver-
 199 gence of soil moisture and groundwater (Nobre et al., 2011). The HAND data used here
 200 is based on the MERIT digital elevation model at a spatial resolution of 3-arc second
 201 (~ 90 m). We used the original high-resolution data after aggregating (simple average)
 202 to the resolution of our ecohydrological metrics (0.0417°).

203 Canopy height:

204 Since canopy height is an important indicator of ecosystem functions and is asso-
 205 ciated mostly with water limitation (Tao et al., 2016), we analysed the covariation of canopy
 206 height with the decay rate of vegetation cover during dry-down. We used the lidar-derived
 207 canopy height data from the retrievals of the ICESat satellite at a spatial resolution of
 208 1 km (Simard et al., 2011). We used the data after aggregating (simple average) to 0.0417° .

209 3 Methodology

210 The derivation of the ecohydrological metrics (see Table 1) is based exclusively on
 211 the daily FVC time series. The method can be divided into four main steps: (i) mask-

212 ing and retrieval of minimum and maximum FVC (FVC_{min} and FVC_{max}), (ii) detec-
 213 tion of start and end of the dry seasons (t_{ds}), (iii) estimation of the dry season FVC in-
 214 tegral and duration (I_{ds} and D), and (iv) estimation of the FVC decay rate during dry-
 215 down (λ). Each methodological step is described in detail in the following subsections.

Table 1: A summary of ecohydrological metrics derived from FVC time series in this study.

Metric	Quality Diagnostic
Minimum asymptotic value of vegetation cover (FVC_{min})	-
Maximum asymptotic value of vegetation cover (FVC_{max})	
Duration of dry season (D)	Variation
Starting day of year of dry season (t_{ds})	Variation
Integral of time series of vegetation cover in dry season (I_{ds})	Variation
e -folding time of vegetation cover during dry-down (λ)	Variation, number of converged estimations

216 3.1 Masking and retrieval of minimum and maximum FVC

217 To remove the effect of outliers within a time series, we selected the 2nd and 98th
 218 percentiles of the entire records of the FVC data as the minimum (FVC_{min}) and the
 219 maximum asymptotic values (FVC_{max}). To maintain a reliable signal-to-noise ratio, we
 220 filtered out any grid cell with (i) $FVC_{max} < 0.1$ (ii) more than one-third of the time
 221 series were missing before further steps. Due to the simplicity of the derivation of FVC_{min}
 222 and FVC_{max} metrics, quality diagnostics were deemed unnecessary, and not derived in
 223 this set of metrics.

224 3.2 Detection of dry seasons

225 Detection of the dry season was based on a procedure using the first derivative of
 226 the smoothed FVC (V') (see Algorithm 1). We smoothed daily time series of the FVC
 227 with a 31-day moving average (V_{sm}). Then each day in the time series was marked as
 228 decay, recovery or stable. To do so, we set two thresholds for dry and wet seasons as th_{dry}
 229 and th_{wet} , respectively. We used the 75th and 70th percentiles of the negative deriva-
 230 tive (V') as thresholds th_{dry} and $-th_{wet}$ for each grid cell. The magnitude th_{dry} is, thus,
 231 bigger than th_{wet} . Only the magnitude of th_{wet} was taken as a positive threshold to de-
 232 tect the increase in FVC.

233 An observation was considered as decay if $V' < th_{dry}$, recovery if $V' > th_{wet}$,
 234 and stable if $th_{dry} \leq V' \leq th_{wet}$. The resulting time series of classes (decay, stable,
 235 or recovery) were then smoothed by retaining the majority of decay and stable against
 236 recovery within a 5-day moving window. Dry season was then identified as the period
 237 from the beginning of a decay to the end of a stable period. In order to ensure the ro-
 238 bustness of the end of the stable period, especially in hyper-arid regions with poor signal-
 239 to-noise ratio, we extended the detected dry seasons until the next significant increase
 240 in V_{sm} ($> 5\%$ of the corresponding seasonal amplitude of FVC). Note that the selec-
 241 tion of the thresholds and the moving window sizes were based on extensive exploration
 242 and visual inspection of the FVC time series. The exploration was a necessary step to
 243 ensure the robustness against noise in the data, as well to address the diversity of FVC
 244 dynamics across African ecosystems. To highlight the complexity, some representative
 245 time series of FVC in selected grid cells across different climates are included in Appendix
 246 Appendix B.

247 After detection of all dry seasons in the time series, we only selected the longest
 248 one per calendar year. This is necessary for regions where vegetation may potentially
 249 have two growing (and drying) seasons within a year. The longest dry season within a
 250 year is likely to be the most indicative of the largest water limitation, and the under-
 251 lying ecohydrological mechanisms. When the detected dry season spanned over two cal-
 252 endar years, it was assigned as the dry season of the starting year. In total, the dry sea-
 253 son detection algorithm (Algorithm 1) yielded 16,423,339 dry seasons in 1,029,847 grid
 254 cells.

Algorithm 1 Detection of dry seasons from the entire time series

- 1: Smooth FVC time series with 31 days moving average; to yield V_{sm}
 - 2: Calculate the first derivative of FVC time series from V_{sm} with daily step size; to yield V'
 - 3: Through the entire time series, set the threshold for decay as $th_{dry} = \text{percentile}(V', 75)$ where $V' < 0$
 - 4: Through the entire time series, set the threshold for growth as $th_{wet} = -1 \times \text{percentile}(V', 70)$ where $V' < 0$
 - 5: Mark each observation for their corresponding period as:
 - if** $V' < th_{dry}$ **then** decay
 - else if** $V' > th_{wet}$ **then** recovery
 - else** stable
 - 6: Smooth the classes with a 5-day moving window by majority voting
 - 7: Label consecutive observations marked with decay and followed by stable ones as *dry season*
 - 8: Extend every *dry season* label until $V_{sm} > \min(V_{sm}) + 0.05 \times (\max(V_{sm}) - \min(V_{sm}))$ is satisfied in the corresponding season
 - 9: For each grid cell, keep only the longest *dry season* per year
-

255 **3.3 Derivation of duration related metrics**

256 We calculated the integral of FVC during dry season (I_{ds}) as the total area under
 257 the FVC time series from the start to end of the dry season, with the area under FVC_{min}
 258 removed. This can be expressed as,

$$I_{ds} = \sum^{dryseason} FVC(t) - FVC_{min} \quad (1)$$

259 Removal of the baseline FVC value (FVC_{min}) enhances the signal of seasonal de-
 260 cay of vegetation with respect to baseline vegetation activity. Note that, upon necessity,
 261 the full integral (total area under the curve) can be calculated as the sum of I_{ds} and $D \times$
 262 FVC_{min} .

263 From the yearly dry season detection, 16 (the number of years) values of D , t_{ds} and
 264 I_{ds} we computed for each grid cell. We selected the median of the 16 values as the rep-
 265 resentative inference to be used for spatial analyses. The median was preferred over the
 266 mean to make the estimation robust against annual variations, for instance, by intermit-
 267 tent rain events in the dry season or issues related to FVC derivation. In addition, we
 268 also calculate and report the normalised robust Standard Error (SE) as an indicator of
 269 variability. The SE is calculated as,

$$SE = \frac{SD_n}{\sqrt{n}} \quad (2)$$

270 where SD_n is the robust standard error, calculated from the Median Absolute De-
 271 viation (MAD) across years (with the assumption of a normal distribution, Rousseeuw
 272 & Croux, 1993), and corrected for the low number of samples ($n = 16$) as:

$$SD_n = MAD \times 1.4826 \times \frac{n}{n-1} \quad (3)$$

273 The robust standard error reflects variability of the metrics among years as well
 274 as methodological uncertainty, and is therefore suitable for customised filtering in the
 275 context of spatial analysis.

276 3.4 Derivation of exponential decay rate

277 Temporal decay of the FVC can be characterised using an exponential function as,

$$FVC = (FVC_{dd} - FVC_{min}) \times e^{-t/\lambda} + FVC_{min} \quad (4)$$

278 where FVC_{dd} is the initial FVC value in the beginning of a dry-down, and λ is the
 279 e -folding time (in days). Note that λ is merely an inverse of the exponential decay rate.
 280 The formulation in Eq. 4 uses λ as it is easier to interpret. In simple terms, λ denotes
 281 the number of days needed to have a decrease in the seasonal amplitude of FVC ($FVC_{dd} -$
 282 FVC_{min}) to $1/e$ of its original value during a dry-down event.

283 Due to the S-shaped character of temporal vegetation dynamics, functions allow-
 284 ing different convexity, e.g., logistic functions, have been used to characterise the veg-
 285 etation decay. As exponential decay functions are strictly convex, the concave part of
 286 the decay is not considered in this study. Note that curvature is concave mostly at the
 287 beginning of dry season, which is of smaller relevance to the metrics presented here. In
 288 addition, the selected exponential decay function takes into account an asymptotic value
 289 of the FVC, as FVC_{min} (see Sec. 3.1) is explicitly included in the formulation (Eq. 4).

290 At the beginning of a dry season, when water demand of the ecosystem is still largely
 291 supported by surface soil moisture, the FVC typically does not decay at an exponential
 292 rate. To identify the dry-down period, for which λ is estimated, we infer insights from
 293 the mathematical properties of the exponential decay function. As the curvature of the
 294 exponential decay function is strictly convex, the first derivative is negative, and the sec-
 295 ond derivative is positive. Therefore, we first discarded the time steps with concave ob-
 296 servations (negative first and negative second derivative). Afterwards, we filtered out the
 297 convex observations before the inflection point of the FVC, that mostly associated with
 298 low signal-to-noise ratio at the beginning of the dry-down. After marking the observa-
 299 tions as either convex or concave, we searched for local minimum of V' in the first third
 300 of the dry season, and identified the inflection point as the start of the dry-down. Note
 301 that, in the above process, second derivative of the FVC (V'') was also smoothed with
 302 a 31-day moving window.

303 This procedure effectively removes observations with concave shape in the dry sea-
 304 son, especially at the beginning of an event. For each event, if more than half of the data
 305 points showed convexity, we estimated λ , together with FVC_{dd} , based on an asymptotic
 306 regression model that minimises least squares error with the Levenberg–Marquardt al-
 307 gorithm (Moré, 1978; Elzhov et al., 2016)). We used both the Nash–Sutcliffe modelling
 308 efficiency (NSE; Nash & Sutcliffe, 1970) and the standard error of the model (SE_m) to
 309 assess the estimates of the model fitting. From the multiple λ estimates, only those with
 310 successful convergence of the Levenberg–Marquardt algorithm with $NSE > 0.5$ and
 311 $SE_m(\lambda) < 0.5 \times \lambda$ were selected, the median of which was taken as the representative
 312 final λ for a grid cell.

Algorithm 2 Identification of dry-down periods and modelling of the exponential decay

- 1: Smooth V' with 31 days moving average; to yield V'_{sm}
- 2: Calculate the second derivative of FVC time series from V'_{sm} with daily step size; to yield V''
- 3: Smooth V'' with 31 days moving average; to yield V''_{sm}
- 4: Mark each observation with $V'_{sm} < 0$ as:
 - if** $V''_{sm} > 0$ **then** convex
 - else** concave
- 5: Ignore convex observations before the inflection point of FVC time series, if there is any
- 6: Ignore concave observations within the dry season and keep the rest as the dry-down period
- 7: Discard any event having more concave observations than convex
- 8: Use Eq. 4 on dry-down period of the dry season to estimate λ
- 9: Filter out the estimations with $NSE < 0.5$ OR $SE_m(\lambda) > 0.5 \times \lambda$

313 After defining the final λ , we estimated the variation as done in Sec. 3.3. Unlike
 314 in Sec. 3.3, the sample size per grid cell (n) may change, as λ estimation may not con-
 315 verge in cases with high noise. We, therefore, also report the number of successful con-
 316 vergences of the Algorithm 2 as an additional quality diagnostic that can be used for fil-
 317 tering λ (mapped in Fig. G1).

318 4 Ecohydrological metrics from vegetation time series dynamics

319 In this section, we present and discuss the ecohydrological metrics derived in this
 320 study (see Table 1). Here we present the metrics independently, but we summarise their
 321 cross-comparison in Fig. C1. For each metric we show the variation in continental scale
 322 by maps along with zoomed inset plots (see Sec. Appendix D for further information and
 323 visual impression by corresponding Google Earth cut-outs) to visualise regional variabil-
 324 ity. Box plots for major bioclimatic regions (see Sec. 2.2 for the definition) provide in-
 325 sights on the co-variation with large scale climate.

326 4.1 Minimum and maximum FVC

327 The spatial distributions of FVC_{min} and FVC_{max} , histograms of the distribution
 328 in the full domain, and six zoomed insets focusing on selected regions are shown in Fig.
 329 2a and 2b (see Fig. E1 for the seasonal dynamics expressed as $FVC_{max} - FVC_{min}$).
 330 At the continental scale, both FVC_{min} and FVC_{max} follow the climate gradient with
 331 the highest and the lowest values in humid and arid regions, respectively. Nevertheless,
 332 compared to FVC_{max} , FVC_{min} has a stronger spatial gradient associated with climate
 333 seasonality within each major climate group (see Fig. 2c). Understandably, the climatic
 334 groups with a distinct dry season have a lower FVC_{min} . This highlights the effect of wa-
 335 ter limitation on vegetation dynamics in regions with distinct seasonality of water avail-
 336 ability (see Fig. D1 for map of simplified climate classes as well as Google Earth views
 337 of the insets).

338 In addition to the climate-associated large scale gradients, the metrics also exhibit
 339 a substantial local-scale heterogeneity. In arid regions, FVC_{min} is higher in areas closer
 340 to the water sources, as can be seen near the Senegal and Gambia rivers (Box-A in Fig.
 341 2a). Positive effect of seasonal flooding on FVC_{min} is also evident near large inland deltas
 342 (e.g., the Okavango Delta and the Sudd swamp, Box-D and Box-F in Fig. 2a, respec-
 343 tively). Such local-scale heterogeneity clearly exhibits the importance of secondary wa-
 344 ter sources in water-limited systems, especially on top of the large climate-driven spa-

345 tial variations, and highlights the usefulness of vegetation-based asymptotic metrics for
 346 ecohydrological studies.

347 4.2 Dry season duration related metrics

348 The dry season duration, D , also follows the climatic gradient at the continental
 349 scale, with the shortest dry season in tropical humid, the longest in arid, and interme-
 350 diate values in the temperate climates (Fig. 3a). Even for tropical and temperate cli-
 351 mate, D consistently increases when the sub-climate includes a dry season (Fig. 3c). The
 352 decrease in D from arid steppe to arid desert climate suggests that the Algorithm 1 may
 353 still be sensitive to very low signal-to-noise ratios in some of the hyper-arid regions with
 354 low FVC and rare, episodic rainfall. Though, such occurrences can be well identified and
 355 filtered using the variation of D , as the values in some hyper-arid regions are relatively
 356 high (Fig. 3b).

357 At local scales, variations in D emerge as a combined effect of climate and other
 358 local ecohydrological factors, such as proximity to the nearest drainage or occurrences
 359 of shallow water table depth. This, once again, is particularly the case in semi-arid cli-
 360 mates. For example, shorter dry seasons appear in seasonally flooded areas like Barotse
 361 Floodplain, the Okavango Delta, and the Sudd swamp, where shallow water tables of the
 362 floodplains support vegetation for longer periods (Box-D and Box-F in Fig. 3a). In these
 363 regions, lateral water transport and moisture convergence in the floodplains provide an
 364 important buffer for vegetation against the climate-driven dryness, which would not be
 365 detectable from precipitation data.

366 I_{ds} shows on average smaller values in humid tropical and arid desert compared
 367 to the other climates of intermediate dryness. However, variation of I_{ds} within climate
 368 groups is much larger when subject to intermediate dryness (Fig. 4a and 4c). The re-
 369 gional inset plots show the impact of shorter dry season duration on I_{ds} in seasonally
 370 flooding wetlands (Box-D and Box-F of Fig. 4a). However, I_{ds} does not only follow the
 371 patterns of D . For example, the variation of I_{ds} in the Lower Zambezi and its tributaries
 372 does not coincide with that of D (Box-E of Fig. 4a and Fig. 3a). The highest values of
 373 I_{ds} in the Lower Zambezi, bear strong similarity with the rooting depth product of (Wang-
 374 Erlandsson et al., 2016), and the previously reported seasonal hydrologic buffer (Kuppel
 375 et al., 2017) in these regions (see Sec. 5.1 for further corroborations).

376 4.3 Exponential decay rate

377 The λ , presented in Fig. 5a, has a mean value of 41 days with a positively-skewed
 378 distribution at the continental scale. We find the lowest λ values throughout the humid
 379 regions and partially in the arid regions, such as edges of the Sahara desert or the Horn
 380 of Africa. The highest λ values are found in the semi-arid and arid regions. Though vari-
 381 ation of λ (Fig. 5b) suggests that the low values of λ in some hyper-arid regions are as-
 382 sociated with higher uncertainty due to low signal-to-noise ratio.

383 Besides the coherent continental-scale spatial patterns, λ also has strong variations
 384 at the local scale. Stronger lateral moisture convergence positively affects the λ in the
 385 arid regions, as seen in the Senegal (Box-A, Fig. 5a) and the Niger (partially in Box-B,
 386 Fig. 5a) rivers in the arid climate. However, lateral moisture convergence does not al-
 387 ways affect λ positively, as seen in the Upper Zambezi and the Okavango rivers and their
 388 tributaries. The λ is high around the Cuando river, the Okavango Delta and the Liny-
 389 ati swamp, but low in the Barotse Floodplain (Box-D in Fig. 5a). Such non-trivial pat-
 390 terns suggest the role of complex interactions between the vegetation traits and local mois-
 391 ture conditions (Fan et al., 2019), which also effect λ (see Sec. 5.2 for further corrobo-
 392 rations and discussions).

5 Corroborating products and potential applications

5.1 Relationship between I_{ds} and plant available soil water holding capacity

Conceptually, plant water storage capacity is related to the vertical distribution of roots, and the water holding capacity of the soil that is determined largely by texture and organic carbon content. The root profile of water-limited ecosystems appears to adapt to the prevailing hydrologic and soil conditions while being constrained by other ecosystem properties and traits (Guswa, 2008; van Wijk, 2011; Fan et al., 2017; Schenk, 2008; Schenk & Jackson, 2002; Laio et al., 2006). Plant water storage capacity controls the propensity and sensitivity of ecosystems to drought stress in dry periods. Various modelling approaches to infer rooting depth or plant water storage capacity have been proposed (explained in detail in Wang-Erlandsson et al., 2016), as it cannot be observed directly but still contains a critical information for global-scale models (Kleidon & Heimann, 1998).

The integral of the FVC during dry season should be positively correlated with plant accessible water storage of the soil, as larger water storage would facilitate vegetation activity for longer period during water-limited conditions. The continental scale pattern of I_{ds} (Fig. 4a) with the largest values in strongly seasonal semi-arid Savannah systems of both hemispheres is qualitatively consistent with the previous observation-based analysis (e.g. Schenk & Jackson, 2002) as well as the optimality-based models (e.g. Kleidon & Heimann, 1998). I_{ds} declines in hyper-arid regions like the Sahel, Horn of Africa, Southern Africa, as well as the Congo rainforest. A similar pattern would be expected for optimal rooting depth, which increases in regions with small differences between rainfall and potential evaporation in annual scales but large differences in seasonal scales (Laio et al., 2006; van Wijk, 2011). The inset plots in Fig. 4a clearly reveal the landscape scale patterns of I_{ds} , presumably, due to topography-driven large variations of moisture. This may reflect enhanced and continued moisture supply due to topographic moisture convergence or shallow water tables along with possible adaptations of rooting depth to these local hydrological conditions (Fan et al., 2017).

We compared I_{ds} with 4 products of plant storage capacity (Wang-Erlandsson et al., 2016; Tian et al., 2019) or rooting depth (Yang et al., 2016; Fan et al., 2017) at 0.5° across Africa. As shown in Fig. H1, there is qualitative agreement of high values of I_{ds} and the storage capacity from Tian et al. (2019) and Wang-Erlandsson et al. (2016) in the Miombo woodlands and, to a lesser extent, also in the northern savannahs. All three also agree on low values in hyper-arid regions like the Sahel, Horn of Africa and in Southern Africa. A pairwise comparison of Spearman's correlation coefficient among the five estimates (Fig. 6) reveals that the strongest agreement is between I_{ds} and storage capacity from Wang-Erlandsson et al. (2016). The overall low-to-moderate correlation values among the previous observation-based products demonstrates the scale of the challenge in estimating plant water storage capacity or rooting depth (Fig. 6).

All four independent products utilised meteorological input data for water balance estimation, and also use remotely-sensed vegetation products in some way. While the products of Wang-Erlandsson et al. (2016) and Tian et al. (2019) are constrained by hydrological earth observations, the rooting depth products of Fan et al. (2017) and Yang et al. (2016) originate largely from different assumptions of optimality and plant adaptation. Our comparison suggests that estimating plant storage capacity based on Earth observation data may be more suitable than the presently-used optimality principles. Using I_{ds} as an indicator of plant water storage capacity has the advantage that it is derived from dense time series of a geostationary satellite alone, requiring no additional meteorological inputs or modelling assumptions that introduce their inherent uncertainties. In a sense, the remote sensing based I_{ds} is directly associated with the actual vegetation growth, and is completely independent of the assumptions and uncertainties of theoretical models or meteorological dataset. Furthermore, I_{ds} features much higher spa-

445 tial resolution than most other storage capacity data, which provides insights in the role
446 of topography-driven subsurface moisture variations.

447 There are many factors other than plant water storage capacity that could influ-
448 ence I_{ds} . Therefore we compared the variation of I_{ds} with the HAND data in different
449 climate groups to find associations of I_{ds} with topographically induced moisture vari-
450 ations (Fig. 7). In dry regions, I_{ds} tends to decline with increasing HAND, apparently
451 above a varying HAND threshold. This implies that shallow water tables may support
452 vegetation with additional moisture availability under dry conditions, as also shown in
453 Koirala et al. (2017). Therefore, I_{ds} is a suitable proxy for analysing the complex pat-
454 terns and processes surrounding groundwater–soil moisture–vegetation interactions.

455 5.2 λ and canopy height

456 The rate of FVC decay during dry-down events, λ , can corroborate to the rate of
457 decrease of plant available water, ecosystem scale water use efficiency, and the propen-
458 sity to senescence. Ecosystems differ widely in their water use strategies, from being wa-
459 ter conservative – typically associated with strong down-regulation of stomatal conduc-
460 tance with water deficiency – to aggressive exploitation of water resources (Laio et al.,
461 2001). Herbaceous plants are typically aggressive water users and cease with the deple-
462 tion of surface soil moisture. Woody plants risk cavitation and death under severe wa-
463 ter stress, and such, trees in places with frequent dry periods benefit from a water sav-
464 ing strategy or senescence for prolonged periods. Konings and Gentine (2017) inferred
465 ecosystem water-use strategies globally based on diurnal variations of vegetation opti-
466 cal depth assuming that those reflect stomatal regulation to maintain leaf-water poten-
467 tial. They found an increase in isohydricity, the degree of stomatal regulation and sub-
468 sequent water savings, with increase in vegetation height, consistent with the need of tall
469 trees to prevent hydraulic failure during drought.

470 If the rate of FVC decay was also related to the ecosystems’ water use strategy in
471 a similar manner, we would expect slower FVC decay (higher λ) with increasing canopy
472 height. For arid regions, we indeed find a tendency of increasing λ with canopy height
473 (Fig. 8), suggesting that λ incorporates ecosystem water use strategy traits as well as
474 direct/indirect effects of soil moisture therein. However, as the climate gets wetter, or
475 over the entire African continent, λ tends to decrease with canopy height. A possible ex-
476 planation would be that water consumption, i.e. transpiration, increases with canopy height
477 resulting in a faster depletion of moisture storage (Koirala et al., 2017), or increasing ecosys-
478 tem water use efficiency with aridity. Even though interpretation of the spatial variabil-
479 ity of λ remains speculative at this point, the initial analysis and considerations given
480 here show the potential of gaining ecohydrological insights, especially for model-data-
481 fusion exercises.

482 6 Conclusions and outlook

483 Using retrievals of the SEVIRI sensor of the geostationary satellite MSG, we de-
484 rived ecohydrological metrics for continental Africa entirely from the temporal dynam-
485 ics of the daily Fraction of Vegetation Cover (FVC) time series from 2004 to 2019 at 0.0417°
486 spatial resolution. Our metrics captures both, continental scale gradients and covaria-
487 tions with climate as well as structured regional variations, e.g. due to topographic fac-
488 tors. This provides an unprecedented opportunity to improve our understanding of eco-
489 hydrological processes across spatial scales over Africa.

490 The minimum asymptotic value of vegetation cover (FVC_{min}) gives indications on
491 where secondary water resources support vegetation in the dry season. Duration and start-
492 ing day of the dry season (D and t_{ds} , respectively) show the effective extent and start
493 of the water-limited period, a critical source of information for any ecohydrological anal-

494 ysis or model study. Because they incorporate the effects of non-climatic factors as well,
495 they are complementary to e.g. precipitation-based dry season delineations and have fur-
496 ther the advantage that they can be estimated at higher spatial resolutions. The inte-
497 gral of FVC time series in dry season (I_{ds}) indicates buffering capacity of vegetation on
498 moisture limitation and shows broad consistency with inferred variations of the plant stor-
499 age capacity or rooting depth. Since this is an important, but at the same time a very
500 uncertain, aspect in ecohydrology, our high-resolution estimate of I_{ds} may help under-
501 stand and model ecohydrological processes more accurately. The spatial patterns of I_{ds}
502 may be used to analyse plant water storage capacity in ecohydrological models and re-
503 place simplistic approaches where this varies only with vegetation type and soil. Finally,
504 the e -folding time of vegetation cover during dry-down (λ) reveals the decay rate of veg-
505 etation during dry season, which emerges from the complex ecohydrological interactions.
506 Using the structured but highly variable spatial patterns of λ , we believe much can be
507 learned about underlying mechanisms by thorough analysis and modelling studies. The
508 suggested algorithms for deriving the metrics and the provision of the code facilitates
509 consistent parallel assessments and helps overcome the technical difficulties of dealing
510 with large volumes of data and the particularities of vegetation cover retrievals from the
511 geostationary satellites. There remain multiple opportunities for further synergistic ex-
512 ploitation with retrievals of surface temperature from geostationary satellites which could
513 provide complementary indicators on variations of moisture states inferred from an en-
514 ergy balance perspective.

515 **7 Data and code availability**

516 All ecohydrological metrics and their quality diagnostics derived and presented in
517 this study are available in standardised netCDF data format in [https://doi.org/10](https://doi.org/10.17871/bgi_ehydro_afr_2020)
518 [.17871/bgi_ehydro_afr_2020](https://doi.org/10.17871/bgi_ehydro_afr_2020) (use [ftp://ftp.bgc-jena.mpg.de/pub/outgoing/ckucuk/](ftp://ftp.bgc-jena.mpg.de/pub/outgoing/ckucuk/ecoHydro_Afr)
519 [ecoHydro_Afr](ftp://ftp.bgc-jena.mpg.de/pub/outgoing/ckucuk/ecoHydro_Afr) to download the data anonymously) (Küçük et al., 2020).

520 The R scripts developed for the implementation of the methodology are available
521 for research uses. They can be accessed through [https://github.com/caglarkucuk/](https://github.com/caglarkucuk/EcohydroMetrics_Africa.git)
522 [EcohydroMetrics_Africa.git](https://github.com/caglarkucuk/EcohydroMetrics_Africa.git) (also at [ftp://ftp.bgc-jena.mpg.de/pub/outgoing/](ftp://ftp.bgc-jena.mpg.de/pub/outgoing/ckucuk/EcohydroMetrics_Africa_Repository.zip)
523 [ckucuk/EcohydroMetrics_Africa_Repository.zip](ftp://ftp.bgc-jena.mpg.de/pub/outgoing/ckucuk/EcohydroMetrics_Africa_Repository.zip) to ensure anonymity) and cited as
524 Küçük et al. (2020).

525 The ancillary data from Tian et al. (2019) was obtained by contacting to the cor-
526 responding author. All other datasets were obtained from the public domain using the
527 information in the cited literature (see Sec. 2).

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530 for Global Biogeochemical Cycles. Diego G. Miralles acknowledges funding from the Eu-
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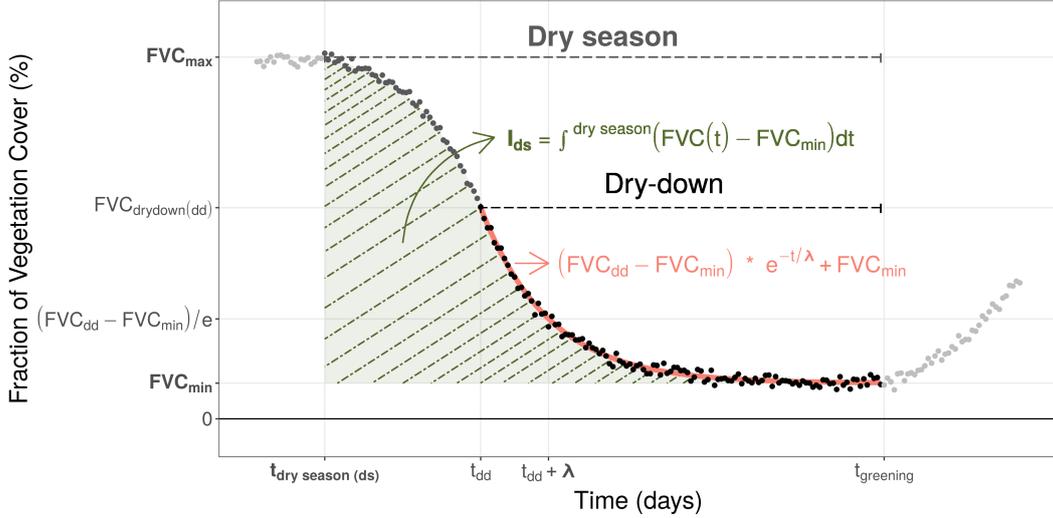


Figure 1: Conceptual plot of the ecohydrological metrics derived from time series using synthetic data. Points represent observations for wet season, early dry season and dry season with dry-down in light grey, grey and black, respectively. Dry and wet seasons are defined by presence of decay, i.e., first derivative of the time series, while dry-down period is defined by the convexity of the decay, i.e., using both first and second derivatives (see Sec. 3.4 for details). The shaded area shows the integral of FVC during dry season. The curve shows the fitted line on the FVC time series during dry-down using the asymptotic exponential decay function. All metrics presented in this study are shown in bold characters.

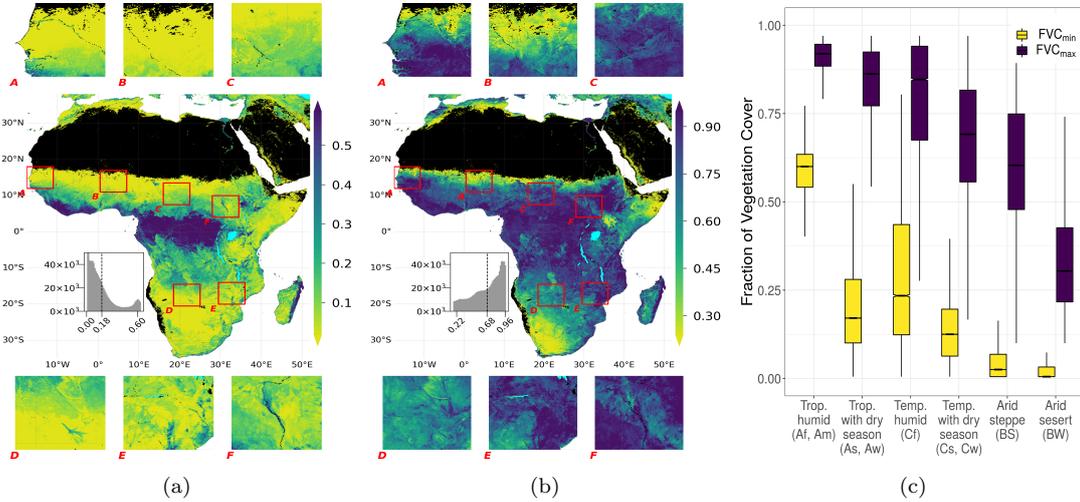


Figure 2: (a) Minimum asymptotic values of FVC, FVC_{min} , (b) maximum asymptotic values of FVC, FVC_{max} , (c) box plot showing the distribution of FVC_{min} and FVC_{max} in different climate groups. A histogram of the metrics mapped can be seen inside the major plot, with a dashed line indicating the mean values of the domain in all maps. See Sec. Appendix D for further explanation of the insets in the map. In all of the following box plots, median values per class are shown in the intermediate line of the boxes, with their 95 % confidence intervals notched. Upper and lower edges of the boxes show the interquartile range (75th and 25th percentiles, respectively) while the error bars show 1.5 times the interquartile range.

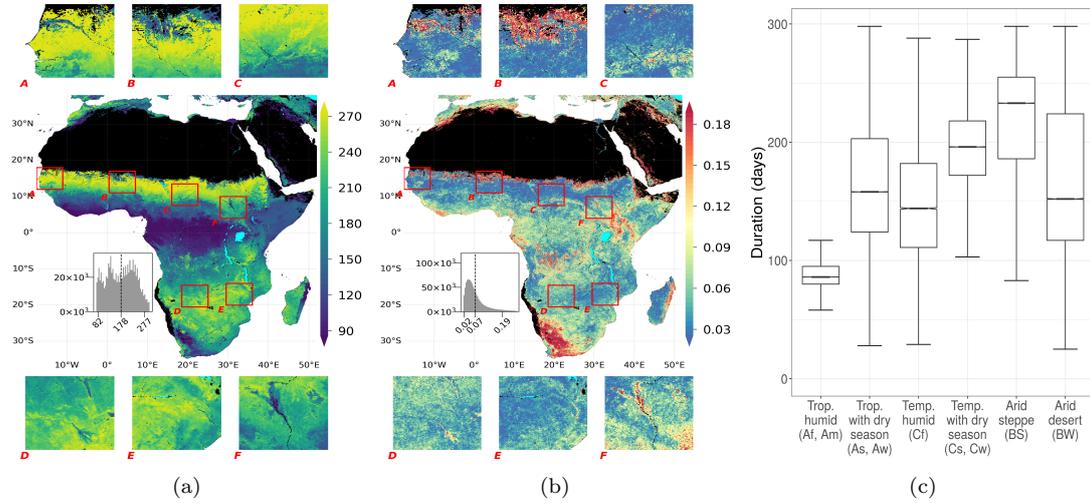


Figure 3: (a) Duration of the dry season (in days), D , (b) variation of D , (c) distribution of D within climate groups (see Fig. 2c for plotting details).

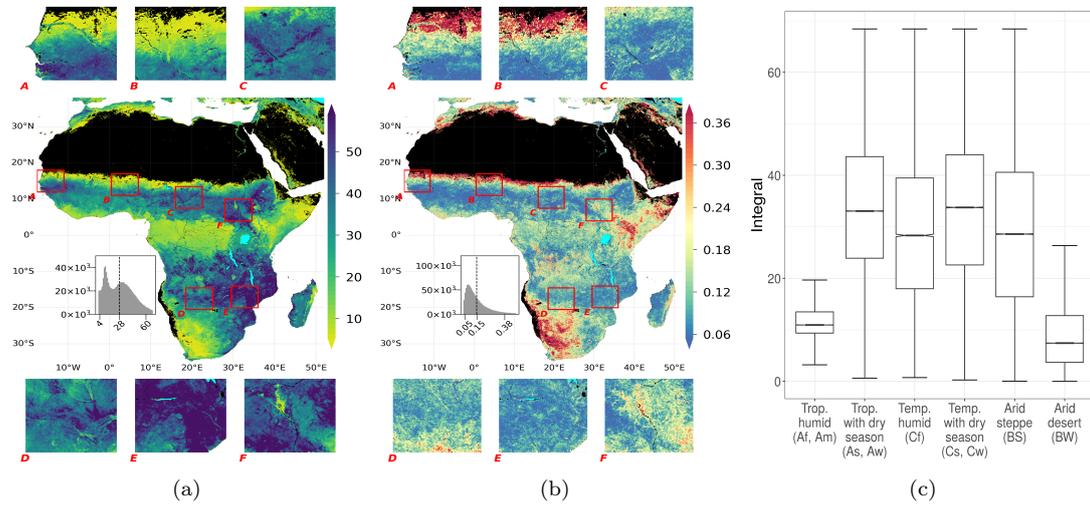


Figure 4: (a) Integral of FVC time series in the dry season, I_{ds} , (b) variation of I_{ds} , (c) distribution of I_{ds} within climate groups (see Fig. 2c for plotting details).

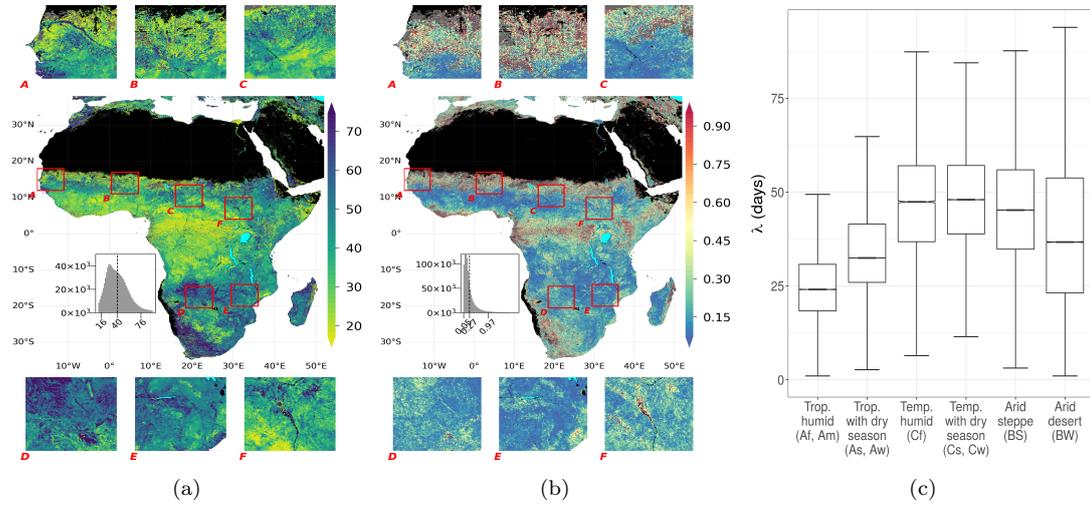


Figure 5: (a) e -folding time of FVC time series during dry-down (in days), λ , (b) variation of λ , (c) distribution of λ within climate groups (see Fig. 2c for plotting details).



Figure 6: Spearman's correlation coefficients between different plant available water storage products.

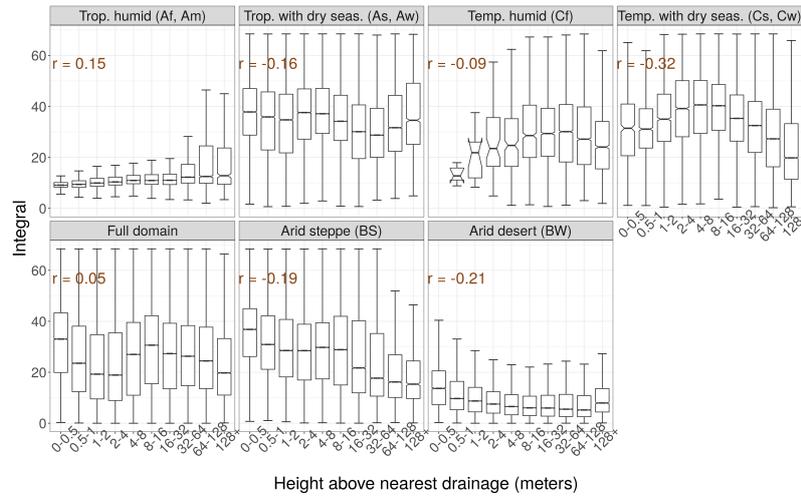


Figure 7: Covariation of I_{ds} and HAND for different climate groups, and for the full study domain. Spearman's correlation coefficients between I_{ds} and HAND are annotated in the panels.

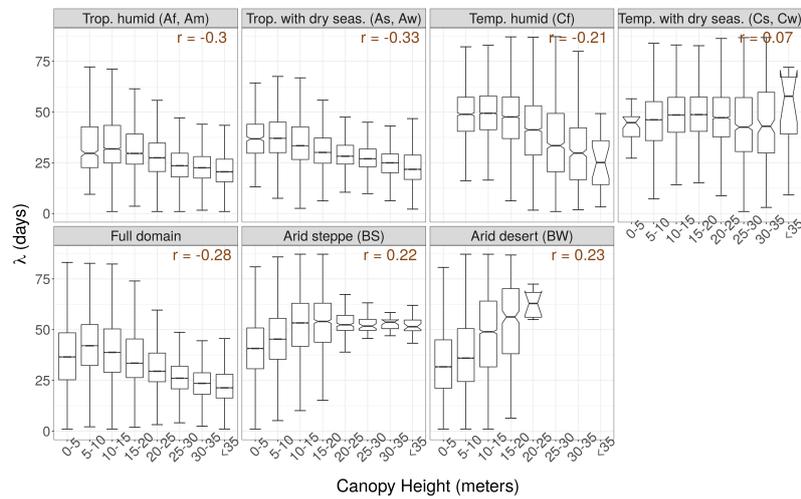


Figure 8: Covariation of λ and canopy height for different climate groups, and for the full study domain. Spearman's correlation coefficients between λ and canopy height are annotated in the panels.

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Appendix A An example map of the original FVC data for a single day

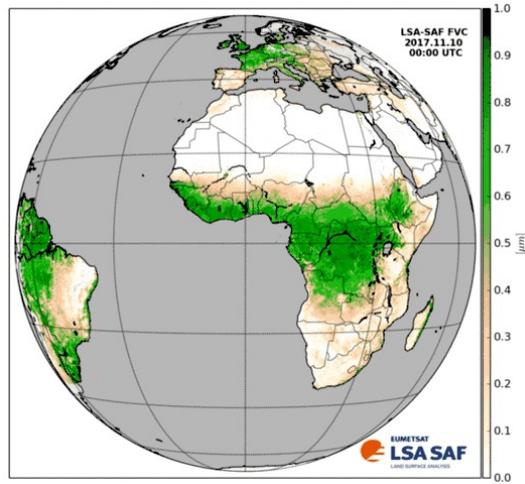


Figure A1: The original FVC data product for a single day, taken from <https://landsaf.ipma.pt/en/products/vegetation/fvc/>

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Appendix B Time series of FVC in example grid cells

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In this subsection; we present 5 years time series of two selected grid cells from each simplified climate class to demonstrate the results of the algorithms in grid cell scale.

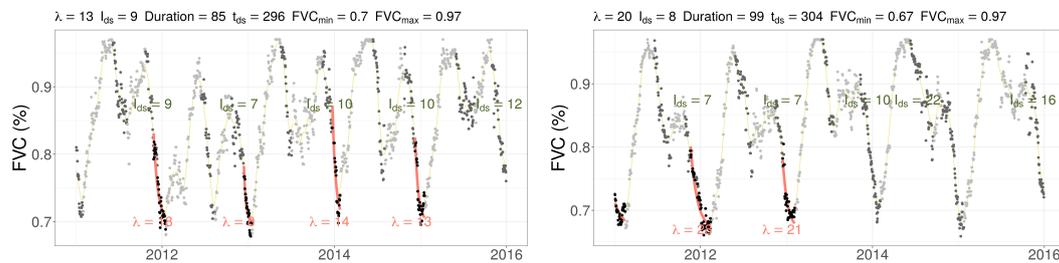


Figure B1: Time series of two grid cells from tropical humid climate (Af, Am) with coordinates (23.645832, 2.562501) and (29.145832, 2.562501), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

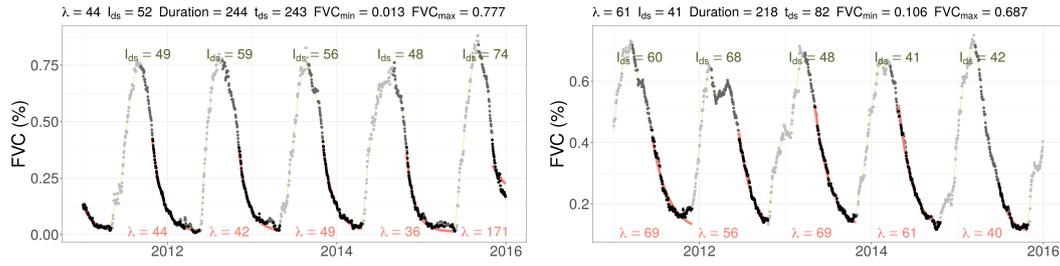


Figure B2: Time series of two grid cells from tropical climate with dry season (As, Aw) with coordinates (9.562499, 10.145834) and (45.812498, -24.479165), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

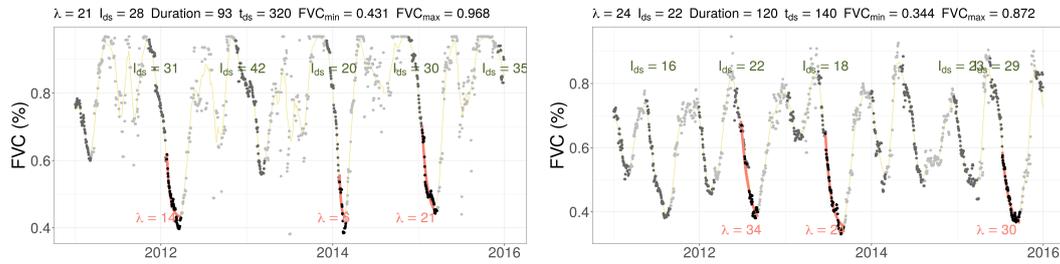


Figure B3: Time series of two grid cells from temperate humid climate (Cf) with coordinates (36.520832, 7.020834) and (30.104165, -1.104165), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

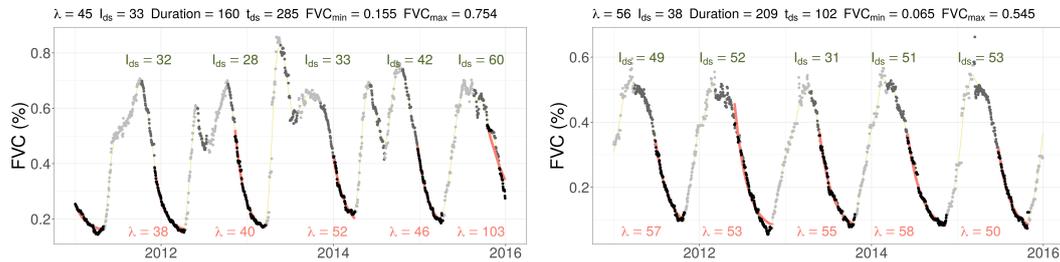


Figure B4: Time series of two grid cells from temperate climate with dry season (Cs, Cw) with coordinates (41.729165, 8.937501) and (47.020831, -20.312498), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

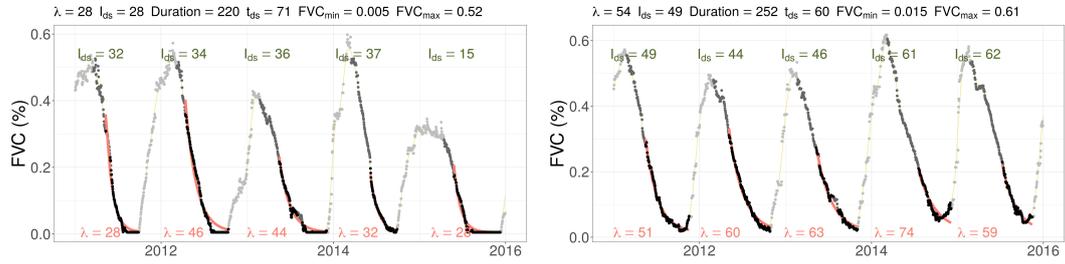


Figure B5: Time series of two grid cells from arid climate with steppe land cover (BS) with coordinates (18.354165, -20.229165) and (30.104165, -18.854165), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

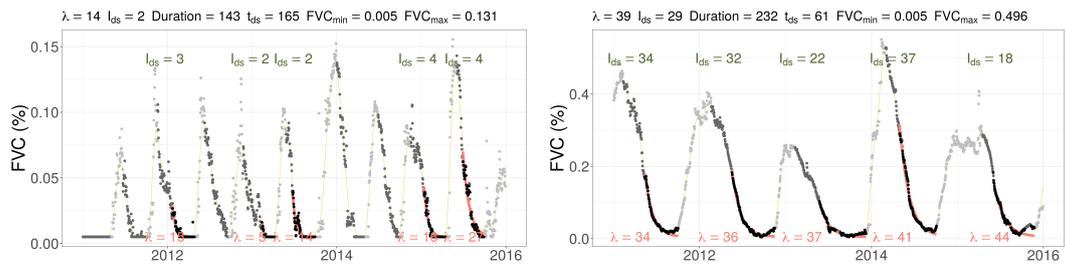


Figure B6: Time series of two grid cells from arid climate with desert land cover (BW) with coordinates (49.395831, 7.479168) and (19.645832, -21.520832), respectively. Seasonal values of λ and I_{ds} are shown inside the plot while final values of the metrics are given in the plot titles.

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Appendix C Density plots of the ecohydrological metrics

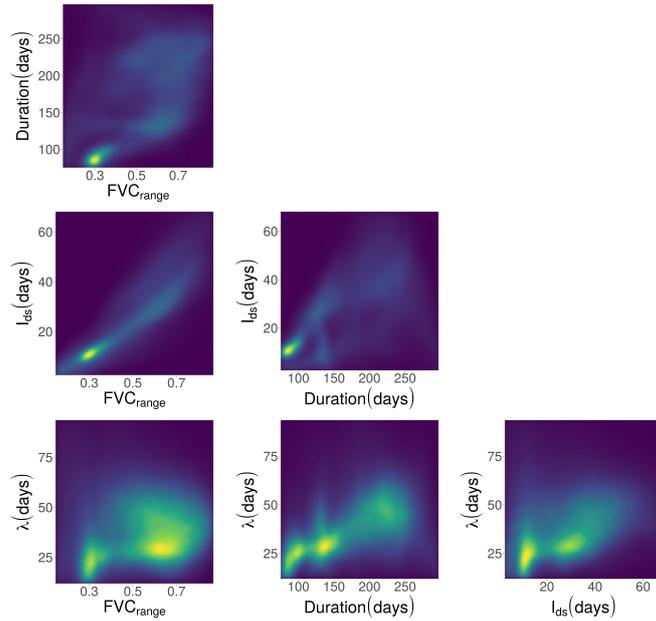


Figure C1: Density plots of the ecohydrological metrics presented in this study. $FVC_{range} = FVC_{max} - FVC_{min}$ is used to summarise the minimum and maximum FVC values.

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Appendix D Map of simplified climate classes and Google Earth view of insets

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Fig. D1 shows the continental map of the simplified climate classes and the Google Earth views of the insets. Box-A: the Gambia and most of the Senegal rivers; Box-B: a small area of the Niger river mostly showing the transition from the Sahara desert to Sahel; Box-C: more on the transition from Sahel to tropical regions; Box-D: located in one of the most complex regions of Africa in terms of topography and lateral flow of water with lower sections of the Okavango and the Cuando rivers and upper section of the Zambezi river, together with multiple seasonally flooding areas like the Okavango delta, the Linyati swamp and the Barotse Floodplain. These seasonal wetlands are vital for the ecosystem and also provides great support against water limitation and heat for not only plants but also animals; Box-E: Lower Zambezi Basin together with the drainage of Lake Malawi to Zambezi. It also covers the Inyanga mountains located between Mozambique and Zimbabwe where a climatic shift happens due to the mountain range. Last but not least, Box-F, which is divided by the White Nile from South to North, covers the Sudd swamp with a climatic gradient from tropical to arid systems.

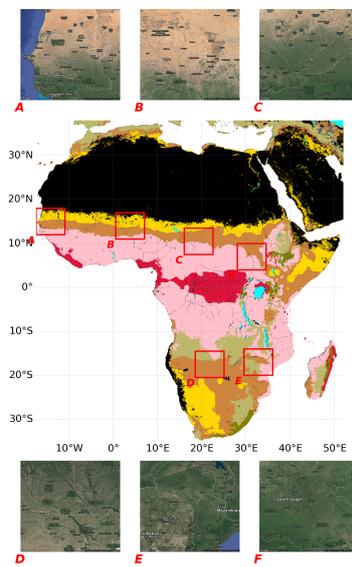


Figure D1: Map of simplified climate classes (from Köppen–Geiger climate classification) in the centre and satellite view of the insets. Map and image data of the insets: Google Earth ©2020 TerraMetrics.

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Appendix E Map of FVC_{range} to show insights of FVC_{min} and FVC_{max}

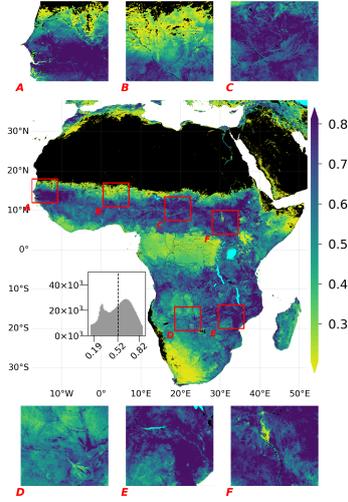


Figure E1: $FVC_{range} = FVC_{max} - FVC_{min}$

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Appendix F Map of starting day of year of dry season

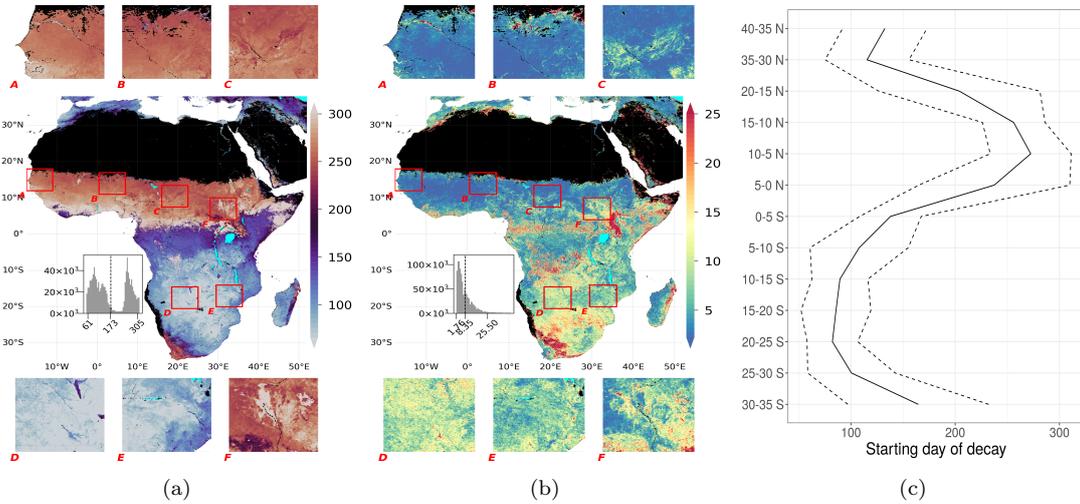


Figure F1: (a) Starting day of the dry season, t_{ds} , (b) SE of t_{ds} across years as a quality diagnostic, (c) latitudinal distribution of t_{ds} , where mean values per bin shown with continuous while standard deviations are shown with dashed lines.

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Appendix G Map of number of convergences of Algorithm 2

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Appendix H Maps of accessible water storage capacity datasets

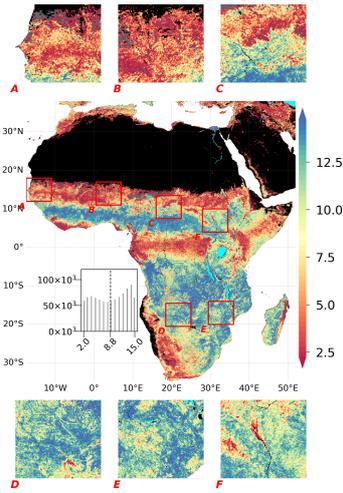


Figure G1: Number of dry seasons in which the Algorithm 2 successfully converged.

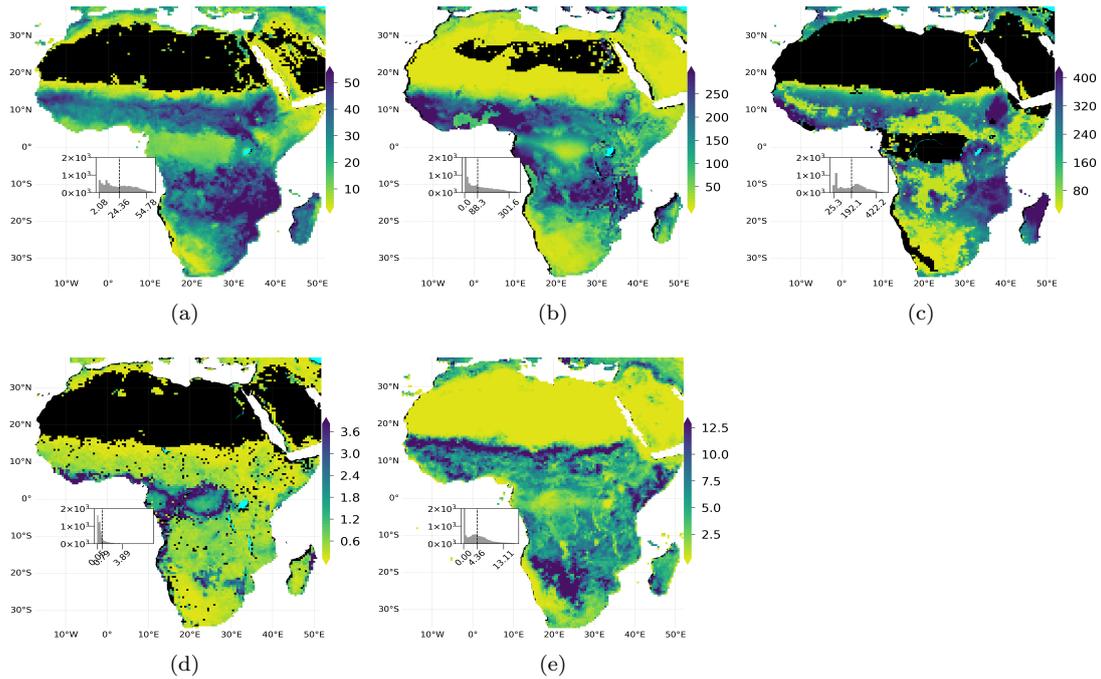


Figure H1: Maps of accessible water storage capacity and rooting depth datasets used in this study. (a) Integrated FVC during dry season, I_{ds} , (b) root zone storage capacity with CRU precipitation data with 2 years of drought return period from Wang-Erlandsson et al., 2016, (c) accessible water storage capacity from Tian et al., 2019 (d) effective rooting depth from Yang et al., 2016, (e) rooting depth from Fan et al., 2017. All products are aggregated to 0.5° and cropped for the study domain.

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