

1 **What are different measures of mobility changes telling**
2 **us about emission reductions during the COVID-19**
3 **pandemic?**

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12 **Key Points:**

- 13 • In light of the COVID-19 pandemic, vehicle emission reductions are in the range
14 of 7-22% in seven investigated urban and rural regions.
- 15 • Recent work used mobility data for assessing the impact of the pandemic on traf-
16 fic emissions. However, we observe errors in excess of 60%.
- 17 • Referencing and representation errors are the main drivers for the discrepancies,
18 which can not be described by functional relationships.

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19 Abstract

20 The COVID-19 pandemic led to widespread reductions in mobility and induced ob-
 21 servable changes in the atmosphere. Recent work has employed novel mobility datasets
 22 as a proxy for trace gas emissions from traffic, yet there has been little work evaluating
 23 these emission numbers. Here, we systematically compare these mobility datasets to traf-
 24 fic data from local governments in seven diverse urban and rural regions to character-
 25 ize the magnitude of errors in emissions that result from using the mobility data. We
 26 observe differences in excess of 60% between these mobility datasets and local traffic data,
 27 which result in large errors in emission estimates. We could not find a general functional
 28 relationship between mobility data and traffic flow over all regions. Future work should
 29 be cautious when using these mobility metrics for emission estimates. Further, we use
 30 the local government data to identify emission reductions from traffic in the range of 7-
 31 22% in 2020 compared to 2019.

32 Plain Language Summary

33 The government-imposed mobility restrictions due to the COVID-19 pandemic led
 34 to observable changes in our atmosphere. We identify atmospheric traffic emission re-
 35 ductions in the range of 7-22% in 2020 compared to 2019 in seven diverse urban and ru-
 36 ral regions using traffic data from local governments. Previous studies investigating these
 37 observed changes used new datasets from tech companies that track user mobility. How-
 38 ever, our work identifies important shortcomings using these new mobility datasets to
 39 directly estimate emissions from traffic, with calculated emission errors larger than 60%.
 40 Further, we could not find a simple functional relationship between these new mobility
 41 datasets and data from local governments on traffic flow, implying caution when using
 42 these mobility metrics for assessing emissions.

43 1 Introduction

44 The COVID-19 pandemic induced widespread changes in society, impacted the global
 45 economy, and has indirectly impacted the environment. Specifically, the emergence of
 46 COVID-19 led to government restrictions on mobility including shelter-in-place orders
 47 and bans on social events (World Health Organisation (WHO), 2020). There has been
 48 much interest in understanding and quantifying how these regulations modulated both
 49 emissions to the atmosphere and the chemical composition of the atmosphere (e.g., Tanzer-
 50 Gruener et al., 2020; Turner et al., 2020; Dietrich et al., 2020). Recent studies have tried
 51 to quantify the impact of the enforced and voluntary restriction of human activities (travel
 52 and work related) on global greenhouse gas (GHG) emissions (Forster et al., 2020; Le Quéré
 53 et al., 2020; Liu et al., 2020) and air pollution (e.g., Venter et al., 2020; Grange et al.,
 54 2020). Many of these studies employed global mobility datasets from Apple Inc. (2020),
 55 Google LLC (2020), and TomTom International BV (2020) and concluded that decrease
 56 in mobility was one of the leading reason of decreased global GHG emissions and air pol-
 57 lution during COVID-19 lockdown periods. These global mobility datasets are highly
 58 attractive as they provide a near-real time estimate of changes in human activity across
 59 nations and over time (Forster et al., 2020). However, in many cases, there is a lack of
 60 transparency about the methodology and, as such, we are left wondering how *exactly* these
 61 datasets relate to emissions (Forster et al., 2020). Further understanding of what these
 62 datasets can tell us about trace gas emissions and atmospheric composition is warranted.

63 Here we investigate these measures of mobility and compare them to data from lo-
 64 cal governments regarding their utility as a proxy for trace gas emissions from traffic.
 65 Through a series of case studies in seven urban and rural regions, we highlight cases where
 66 the mobility data is consistent with local governmental data on traffic flow and, impor-
 67 tantly, cases where the mobility data is inconsistent. We then quantify the potential er-
 68 rors in emission estimates when using these mobility datasets timely and regionally di-

69 vided with a particular focus on CO₂. Finally, we provide an estimate of emission changes
70 due to COVID-19 based on the available local government data for the regions analyzed
71 in our case studies.

72 2 Regions for case studies and investigated datasets

73 We selected seven regions (Oslo, Munich, San Francisco Bay Area, Los Angeles,
74 Cape Town, Norway, California; Supplemental Table S1) as case studies to identify the
75 impact of COVID-19 on traffic emissions. These seven regions encompass both urban
76 and rural regions from four countries on three different continents. They were chosen for
77 their latitudinal coverage and availability of data from local governments on traffic. The
78 distribution of the regions over the latitudes and the coverage of the northern and south-
79 ern hemisphere enable a comprehensive data analysis. Diverse seasonal climate behav-
80 iors are covered, for example the strong and weak temperature seasonality in Oslo and
81 in California (Supplemental Figure S2). While Norway and California are comparable
82 in size, the population of California is around 8 times higher than in Norway. From Sup-
83 plemental Table S1 we see that all of these regions first enacted restrictions on the mo-
84 bility of their populations between March 13 and March 26 in 2020. Los Angeles shows
85 similar effects to San Francisco Bay Area (see Supplemental Section S10).

86 It is important to note that the measures of mobility do not all report the same
87 quantity. Additionally, the metric reported in the mobility datasets differs from the met-
88 rics that are traditionally used to estimate emissions to the atmosphere (e.g., Janssens-
89 Maenhout et al., 2019; Oda et al., 2018).

90 The Apple Inc. (2020) mobility trends report represents the relative request vol-
91 ume of Apple Maps in the categories driving, walking, and public transportation glob-
92 ally. The baseline is the request volume as of Monday, January 13, 2020, reaching from
93 midnight to midnight of the corresponding day in Pacific Time Zone. Apple Inc. (2020)
94 themselves state that increases of their index can occur due to usual seasonality. Also
95 they do not collect user or demographic information and Apple Maps is only available
96 on Apple devices. Therefore, it is unknown whether the use is representative for the en-
97 tire population.

98 The TomTom International BV (2020) traffic index provides congestion levels for
99 416 cities in 57 countries of the world. Due to the COVID-19 pandemic the daily per-
100 centage congestion value for the year 2020 and also the deviation from 2019 are published.
101 The percentage congestion value represents the extra time needed for a trip compared
102 to the uncongested traffic situation. For example, if an uncongested trip takes 30 min-
103 utes and the congestion index currently is 50%, then the trip takes 15 minutes longer.
104 Each weekday is related to the annual average congestion of that same weekday in 2019.
105 The traffic index is calculated with the data of more than 600 million global users who
106 navigate with TomTom technology in navigation devices, smartphones or other techni-
107 cal devices. The uncongested situation is analyzed by looking at free-flow local traffic
108 situations.

109 Here we focus on vehicle traffic and therefore we do not investigate Google LLC
110 (2020) data which provides information about the stay of people at different locations,
111 like e.g. transit stations. As it is included in some recent studies (e.g., Forster et al., 2020;
112 Venter et al., 2020) we have included our investigation in the supplemental material (Sup-
113 plemental Section S11).

114 In contrast to mobile device based data gathering, the local governments measure
115 traffic by point counting stations using microwave radar detectors or induction loops on
116 roads and at traffic lights. For California, we consider the vehicle miles traveled (VMT)
117 metric (California State Senate SB 743, 2015). For all other regions we use the total av-
118 erage daily traffic volume of all point detectors. Data was downloaded directly from the
119 websites or requested from the local governmental departments. For Oslo we reduce the
120 data of Norway by cropping a square with 10 km distance to the city center of Oslo. (Statens

121 vegvesen, 2020; Bayerisches Landesamt fuer Umwelt (LfU), 2020; Caltrans, California
 122 Department of Transportation, 2020; Western Cape Government, Road Network Infor-
 123 mation System, 2020)

124 From Figure 1, we can see that all of the data show an abrupt drop in early March
 125 2020. Interestingly, all of the regions show a nearly synchronous decline even though the
 126 actual government restrictions were implemented over a 3 week period (Supplemental
 127 Table S1). Hence, the San Francisco Bay Area, Munich, and Cape Town show decreases
 128 prior to their actual governmental restriction. We identify deviations, such as the large
 129 increase in summer time in Munich, Oslo, and Norway in the Apple data when compared
 130 to governmental traffic data and TomToms congestion index. All of the regions analyzed
 131 here show substantial differences between mobility and traffic. As such, we are interested
 132 in characterizing what drives these differences and the impacts on bottom-up emissions
 133 inferred using mobility data.

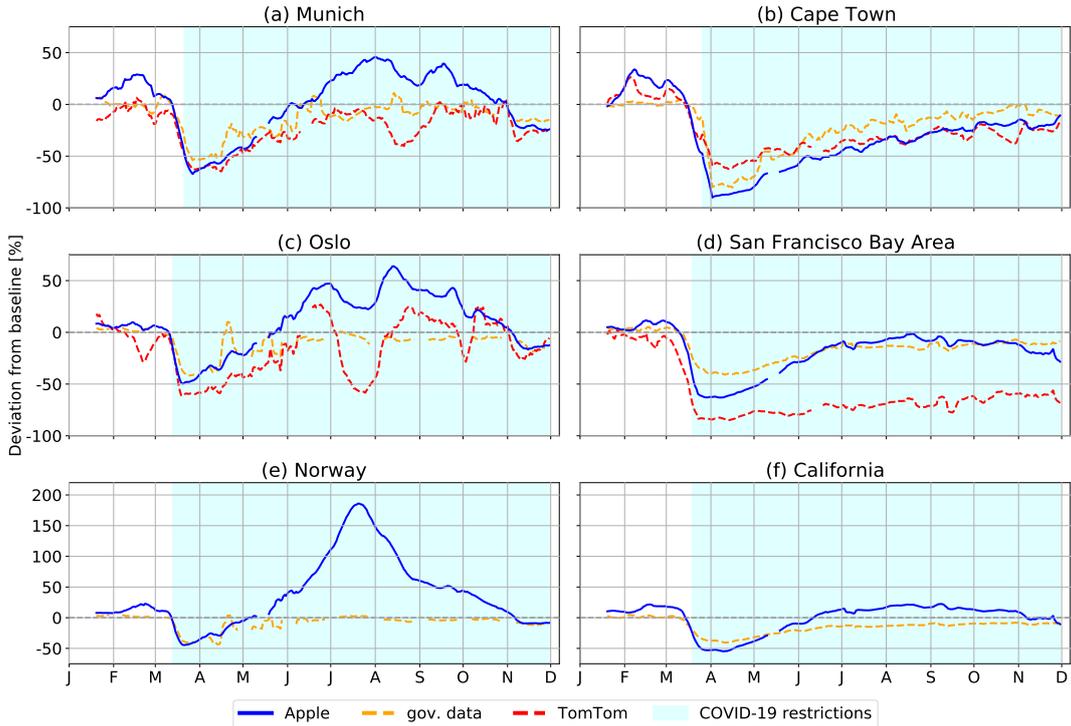


Figure 1. Time series trend comparison of different mobility and traffic datasets. Apple data is relative to its request volume on January 13, 2020. There is no 2019 data for the Apple mobility index as this product was only made public in response to the COVID-19 pandemic. The governmental traffic data each weekday is related to the same weekday of the same calendar week in 2019 and for TomTom data each weekday is related to the annual average congestion of that weekday in 2019. A seven day rolling mean is applied to the data to remove the weekly cycle.

134 **3 Assessing differences between the datasets**

135 As mentioned above, all regions analyzed here show sizable differences between the
 136 temporal evolution of the mobility data and local traffic data (see Figure 1). Addition-
 137 ally, the temporal evolution of these differences varies across regions, and not in an eas-
 138 ily predictable manner. Nevertheless, we are interested in identifying the underly causes

139 of these differences to establish a relationship between mobility and traffic to facilitate
 140 their use in developing bottom-up emission estimates and inferring processes driving changes
 141 in atmospheric composition.

142 Figure 2a shows the monthly deviation from the annual mean traffic flow for six
 143 of the seven study regions using governmental data. We observe little seasonality in Cal-
 144 ifornia (deviations are less than 5%, similar to McDonald et al. (2014)), in contrast to
 145 other regions, which is due, in part, to the temperate climate. The European regions Mu-
 146 nich, Oslo, and Norway show deviations peak of up to 9-12%. Further, we observe the
 147 inverse seasons in the southern to the northern hemisphere in the annual traffic cycle when
 148 we compare Cape Town with the urban study sites Munich and Oslo. Generally the traf-
 149 fic is weaker in the local winter months than in the local summer months in all inves-
 150 tigated regions. The traffic seasonality at higher latitude is larger than at lower latitude
 151 e.g. in California.

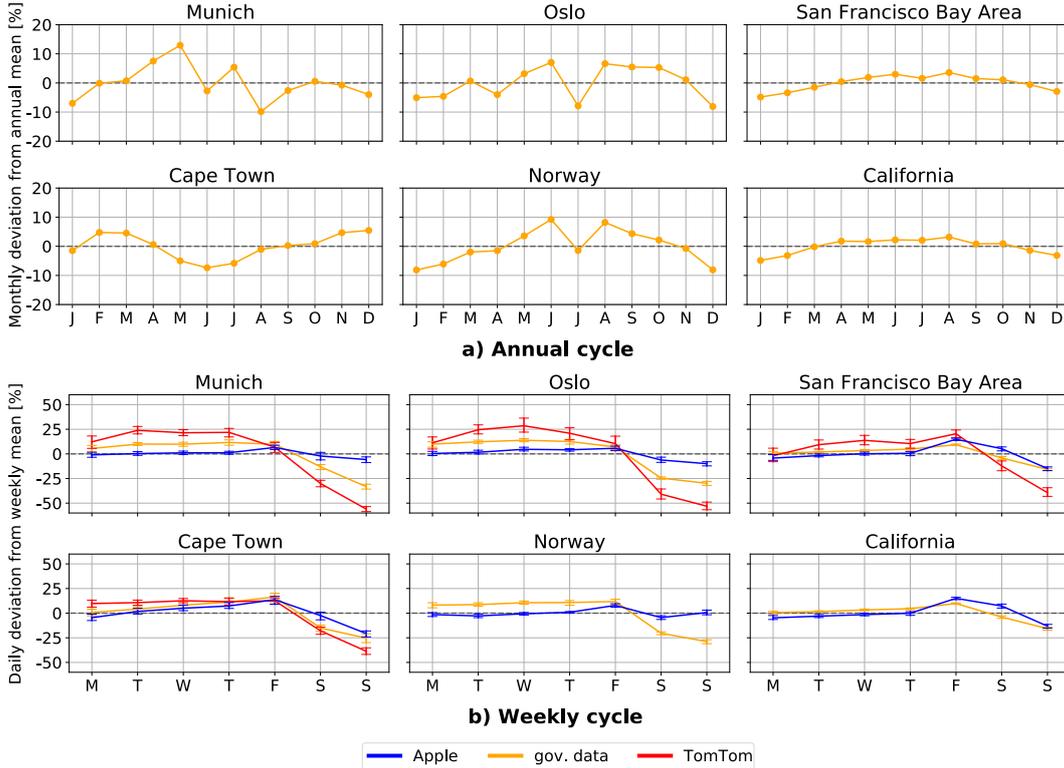


Figure 2. Annual and weekly cycle of traffic and mobility data a) Annual traffic cycle. Deviation of the mean monthly local governmental data of the corresponding month in 2019 to the mean of the year 2019. b) Weekly traffic cycle. Deviation of the daily data of the corresponding weekday to the mean of the corresponding calendar week with 2σ error bars for the time span from 01/14/20 until 11/30/20.

152 Figure 2b shows the daily deviation in traffic flow relative to the weekly mean traf-
 153 fic flow for data from the local government, Apple, and TomTom. All regions show a pro-
 154 nounced decrease in governmental data and TomToms congestion index on the weekend.
 155 A particularly interesting regional difference is the weekly cycle in the TomTom data for
 156 Munich with positive anomalies from Monday through Thursday and a sharp decrease
 157 from Friday through Sunday. This feature is observed in both TomTom and the local
 158 government data, but not Apple mobility data. A similar pattern is seen in Oslo and Cape
 159 Town, but is notably different than San Francisco where all datasets indicate the largest,

160 positive, anomaly on Friday. Apple data indicates the largest positive anomaly on Fri-
 161 days across all regions. The lower traffic values seen on weekends in local governmen-
 162 tal data and TomToms congestion index is also notably smaller in the Apple Maps mo-
 163 bility data.

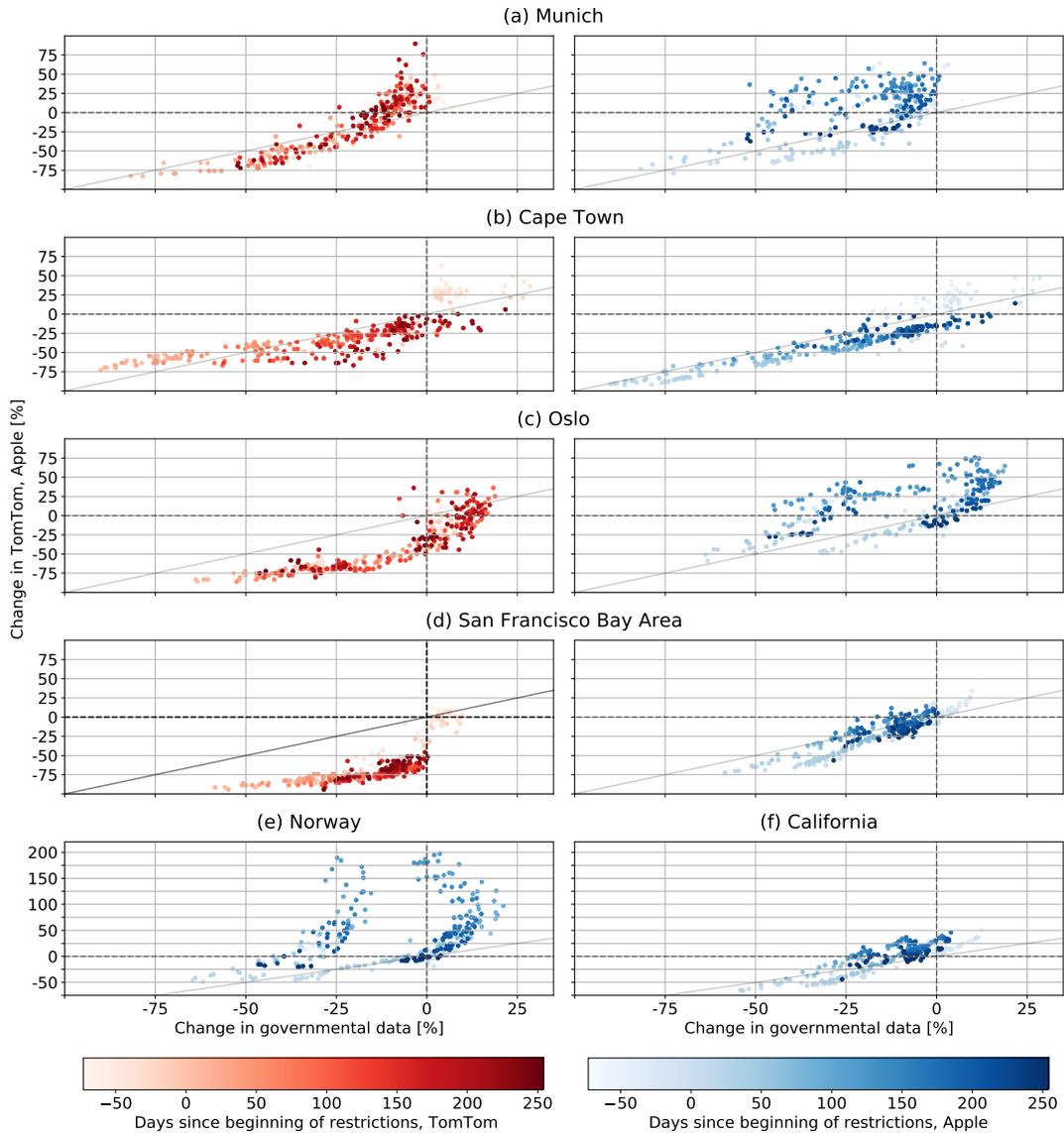


Figure 3. Comparison of different measures of traffic flow. The scatter shows the daily comparison between the governmental data to Apples mobility data and TomToms congestion index. All datasets are referred to their value on January 13, 2020. The coloring of the dots is done by the distance to the first day of local governmental COVID-19 restrictions.

164 The annual traffic cycle (Fig. 2a) and the weekly traffic cycle (Fig. 2b) reveals the
 165 importance of taking annual and weekly seasonality into account, which is however not
 166 the case for Apple data. TomTom data includes weekly cycles but neglects its annual
 167 cycle. Supplemental Figure S1 shows the timeseries of all datasets related to January
 168 13, 2020. We observe large differences between datasets which reveals that the referenc-

169 ing issue only partially explains the differences in Figure 1. These remaining differences
 170 can be attributed to the representation discrepancies that are listed in Section 2.

171 We have highlighted differences between Apple mobility, TomTom congestion and
 172 governmental traffic data (Figure 1). In Figure 3 we assess the relationship between these
 173 metrics using scatterplots. We are interested in comparing the representation of these
 174 metrics and therefore we remove the different baselines by referring all datasets to their
 175 value on January 13, 2020. The coloring of the dots represents the distance to the first
 176 day of governmental COVID-19 restrictions. With increasing brightness the dots are longer
 177 before the first restrictions, while with more darkness they are longer after.

178 From Figure 3 we can see the differences between these metrics cannot be charac-
 179 terized by a relationship that generalizes over all regions. The relationship differs between
 180 cities and is highly scattered for some regions. Removing the impact of weekly cycles by
 181 only comparing weekly means shows a similar trend (Supplemental Figure S5). This in-
 182 dicates that work should be cautious when attempting to estimate trace gas emissions
 183 in response to COVID-19 using (scaled) mobility data, as a number of recent studies have
 184 done (e.g., Forster et al., 2020; Le Quéré et al., 2020; Liu et al., 2020). In supplemen-
 185 tal Figure S15 we have applied the functional relationship of Liu et al. (2020) to the Tom-
 186 Tom congestion index in our study regions and observe big regional differences to real
 187 governmental traffic data.

188 4 Impact of mobility datasets on estimated atmospheric emission change

189 We identify that different measures of traffic and mobility that are currently used
 190 for bottom-up emission estimates deviate strongly from each other. This begs the ques-
 191 tion, “*What do these different measures of traffic and mobility imply about emission changes?*”.
 192 We assess this by assuming that the data from the local governments is the most accu-
 193 rate and look at differences relative to these datasets.

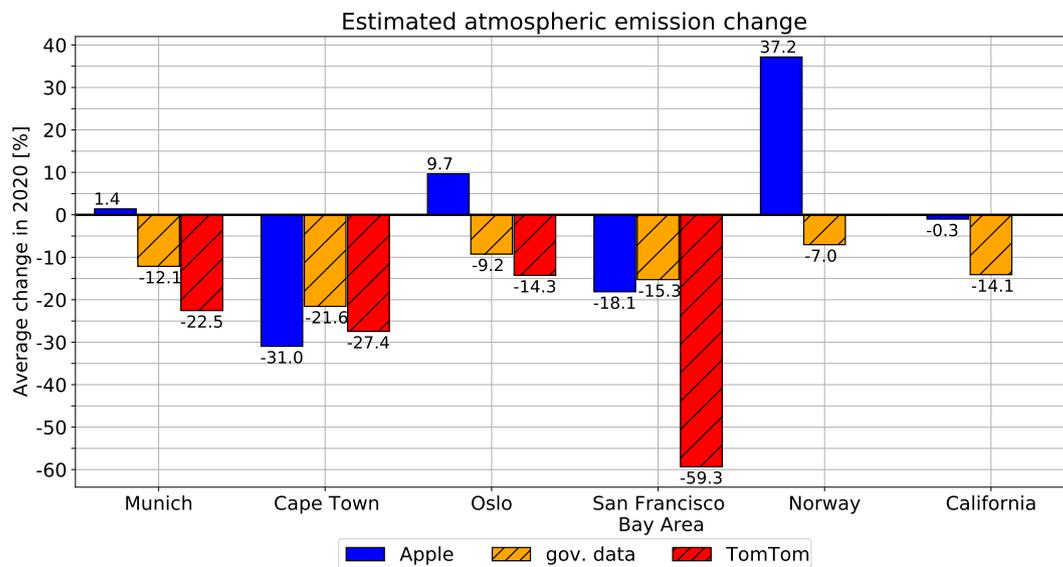


Figure 4. Estimated atmospheric emission change. Traffic emission change in the time span 01/13/2020 until 11/30/20 for six urban and rural regions. Apple data is referenced to January 13, 2020 whereas TomTom and governmental data are to 2019.

194 Figure 4 shows the estimated atmospheric emission change based on those datasets
 195 from January 13, 2020 until November 30, 2020. The bars show the average daily change
 196 of the time series.

197 We quantify the impact of the COVID-19 pandemic on governmental traffic data
 198 which ranges from a decrease of 7.0% to 21.6% depending on the region. The TomTom
 199 congestion index typically indicates a higher decrease than the governmental data. In
 200 the extreme case of the San Francisco Bay Area the TomTom data reduction is about
 201 four times higher than the governmental data reduction. Apple even shows an increase
 202 in Munich, Oslo, and Norway. In Cape Town and the San Francisco Bay Area it shows
 203 a decrease and in California it indicates nearly no change in average over the investigated
 204 period. Supplemental Figure S7 shows the same comparison but also governmental and
 205 TomTom data are related to January 13, 2020 there. Supplemental Figure S8 shows the
 206 time dependent estimated traffic emission change in 2020.

207 Figure 5 shows the difference in trace gas emissions since January 13, 2020 until
 208 the corresponding day on the horizontal axes when TomToms congestion index or Ap-
 209 ples mobility data is used as a proxy for traffic changes instead of governmental traffic
 210 data following Equation 1. If the deviation is negative the usage of the mobility dataset
 211 results in a lower estimated emission number than using the local governmental data:

$$\Delta E(d, g, t) = \frac{\sum_{i=1}^t (d_i - g_i)}{\sum_{i=1}^t g_i} \quad (1)$$

212 where ΔE is the difference in trace gas emissions on the vertical axes in percent; t is the
 213 day on the horizontal date axes; g the local governmental data; and d the datasets of Ap-
 214 pple or TomTom. In Figure 5, the data are denoted as $d^{13,Jan}$, d^{2019} , and g^{2019} , depend-
 215 ing on the baselines that are used for the referencing. In Supplemental Section S8 we use
 216 Eq. 1 with combinations of different baselines for both the local government and mobili-
 217 ty data.

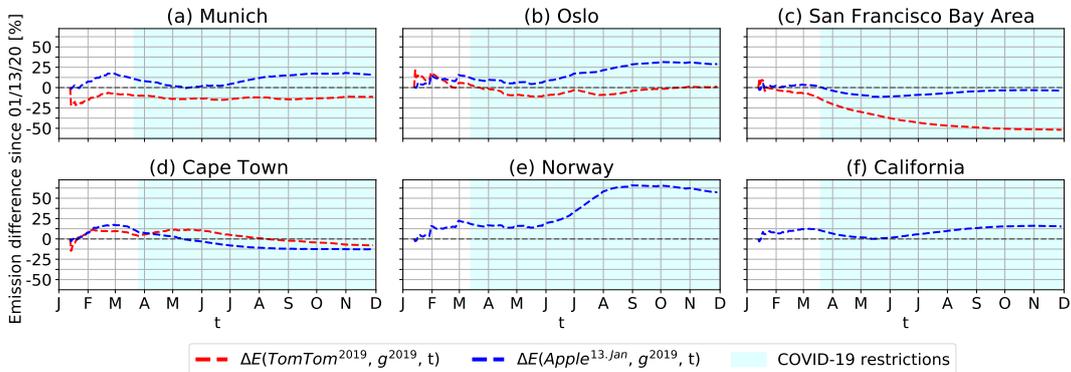


Figure 5. Timeseries of the emission difference (ΔE , Equation 1) of TomToms and Apples data compared to governmental data. The value assigned to one day is the difference in integrated emissions calculation from January 13, 2020 to the corresponding day t using Apples or TomToms data (d) instead of governmental traffic data (g) following Equation 1.

218 We observe in Figure 4 and 5 that the difference between emissions estimates based on
 219 governmental traffic data to estimates based on TomTom congestion index or Apple
 220 mobility data differ for each study region and depend on the timepoint of investigation
 221 (day t after the reference day). The datasets can be a good proxy at one location at a

222 specific time but deviate at another location at the same time (e.g. San Francisco Bay
 223 Area vs. California in end of March). Reasons for this can be caused by the regional an-
 224 nual traffic seasonality that is not taken into account by Apple or TomTom. Relation-
 225 ships between TomTom and Apple data to governmental data can be linear or non-linear
 226 depending on the region (Figure 3, Supplemental Figs S6, and S7). The usual regional
 227 congestion level may also impact the TomTom congestion reduction (Supplemental Fig-
 228 ure S4). The lack of historical data from TomTom and Apple makes it difficult to inves-
 229 tigate the regional differences in the data. The resulting emission differences using mo-
 230 bility datasets are in the range of -13% to 66% and -52% to 21% for Apple and Tom-
 231 Tom, respectively.

232 Taking the San Francisco Bay Area as an example, we calculated the discrepancies
 233 in emission estimates using different datasets. We use Caltrans, California Department
 234 of Transportation (2020) VMT measure (governmental dataset) for the San Francisco
 235 Bay Area as input to the California Air Resources Board’s EMFAC (2014) model to cal-
 236 culate the vehicle trace gas emissions on January 13, 2020. We use the default vehicle
 237 fleet of the model for the ratio of vehicle classes. We then apply the deviations of the
 238 three datasets from January 13, 2020 to the previously calculated vehicle emissions on
 239 that day. For the period of 01/13/20 to 30/11/20 the total differences in the Bay Area
 240 when using Apple instead of VMT are 0.45 Mt CO₂, 452 t NO_x, and 67 t PM which is
 241 a relative vehicle emission difference of -4.55%. Using TomTom instead of VMT results
 242 in an emission difference of 5.7 Mt CO₂, 5653 t NO_x, and 848 t PM (-56.78% in vehi-
 243 cle emission). The percentage error can also be observed in Figure 5 and in Supplement-
 244 al Figure S9 and compared to other regions. These errors in traffic emission estimates
 245 affect the total CO₂ emissions of the San Francisco Bay Area by an underestimate of -
 246 1.6% and -20% using Apple and TomTom, respectively (Supplemental Section S9).

247 5 Discussion and conclusions

248 In this study, we investigated the estimated traffic emission change in 2020 due to
 249 the COVID-19 pandemic in seven urban and rural regions using different measures of
 250 traffic and mobility. Using governmental traffic data, we identify emission reductions in
 251 the range of 7-22% compared to 2019. We compare these results to mobility data pro-
 252 vided by Apple and TomTom and identify two major error sources in emission estimates
 253 when using them as a proxy for vehicle traffic:

- 254 1. **Referencing error.** The impact of the weekly and annual traffic cycle is signif-
 255 icant. Use of a fixed (arbitrary) time-point reference value may yield incorrect con-
 256 clusions (see Figs 1, and 2).
- 257 2. **Representation error.** The datasets investigated here measure different quan-
 258 tities. Local governments typically measure traffic volume and/or vehicle miles trav-
 259 eled, Apples mobility dataset is a measure of their request volume from naviga-
 260 tion systems (Apple Maps), and TomToms congestion index measures urban con-
 261 gestion levels. Even when using the same baseline the deviation of the datasets
 262 is, again, non-trivial (Figure 3, Supplemental Figs S1, S5, S7, and S9).

263 These error sources do not allow us to develop a generalizable relationship between
 264 mobility data and traffic flow over all regions (see Figs 1, 3, 5 and Supplemental Figs S5,
 265 S6, S7, S8), like assumed in Liu et al. (2020) and Forster et al. (2020). Supplemental Fig-
 266 ure S15 shows the error induced by the regression function between TomTom conges-
 267 tion and governmental data used in Liu et al. (2020).

268 We quantify vehicle trace gas emission deviations of -13% to +66% and -52% to
 269 +21% for Apple and TomTom, respectively, compared to data from the local government.
 270 These percentage values depend on the region of interest and time of investigation. In
 271 the case of the San Francisco Bay Area, using the mobility data from Apple and Tom-

Tom results in transportation emission estimates that are, respectively, 0.45 Mt CO₂ and 5.7 Mt CO₂ lower than government traffic data implies, resulting in total emission estimates that differ by -1.6% and -20%.

Despite the widespread use of these mobility metrics, there is a lack of understanding about what exactly they are telling us about changes in trace gas emissions due to COVID-19. Here we quantified the potential errors that might be inferred by using these mobility metrics as a proxy for changes in trace gas emissions. The findings presented here should serve to caution others from directly using these mobility measures as a proxy without additional investigation or adaptation.

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