

1
2 **Preliminary Evidence of Transport-Limited Chemical Weathering and Element**
3 **Immobility in the Ganges Tidal Delta Plain of Bangladesh**

4
5 **J. C. Ayers¹, B. Patton¹, and M. Dietrich¹**

6 ¹ Department of Earth and Environmental Sciences, Vanderbilt University. VU Station B,
7 #351805, 2301 Vanderbilt Place, Nashville, Tennessee 37235-1805, U. S. A.

8
9 Corresponding author: John C. Ayers (john.c.ayers@vanderbilt.edu)

10
11 **Key Points:**

- 12 • In the Ganges tidal delta plain in Bangladesh chemical weathering rates are low and
13 transport-limited.
- 14 • Tidal channel sediments and rice paddy soils are similar in composition to average upper
15 continental crust.
- 16 • Relative to coexisting solids surface waters are depleted in immobile elements but are not
17 in exchange equilibrium.

20 Abstract

21 Previous studies have documented a weathering-limited regime in the upper reaches of the
22 Ganges River Basin. Chemical weathering and element mobility in the lower reaches of the
23 Ganges in the tidal floodplain of Southwest Bangladesh were investigated by comparing
24 compositions of rice paddy soils, precursor tidal channel sediments, surface waters, and extract
25 solutions, which represent the soluble fraction of solids. Little spatial variation in water and solid
26 compositions is observed in each season, indicating similar processes are acting to transport
27 elements across this region. Roughly one to several decades after deposition, rice paddy soils are
28 not significantly different in mineralogy or composition from precursor tidal channel sediments,
29 and both are similar to the composition of average upper continental crust. There is no detectable
30 change in composition of tidal channel water between upstream and downstream sites. Rice
31 paddy and tidal channel waters are saturated in the dominant minerals present in the silt-sized
32 soils and sediments, which are dominantly quartz and clay minerals. Together, these
33 observations indicate the dominance of weathered material and low chemical weathering rates in
34 the tidal floodplain, consistent with a transport-limited regime. Multiple lines of evidence
35 indicate a lack of exchange equilibrium between surface waters and coexisting solids, which may
36 be a common feature in tidal river deltas where transport-limited regimes likely dominate.

37 Plain Language Summary

38 Previous studies of the upper reaches of the Ganges River near the Himalaya have shown that an
39 abundant supply of easily weathered minerals produced by erosion results in rapid weathering
40 and transport of large amounts of dissolved material in the river. We measured the chemical
41 compositions of water from rivers and rice paddies, and of associated river sediments and the
42 soils that form from them, in the lower reaches of the Ganges River tidal delta plain in SW
43 Bangladesh. We found that soils were similar in composition to river sediments, and that river
44 water compositions were constant across our study area. These observations suggest that little
45 material is transferred from solids to water over the course of roughly one decade. Evidence
46 suggests that this is either because the soluble minerals were dissolved in the upper reaches and
47 are not present in sediment in the lower reaches, or the water has dissolved as much of the
48 minerals as possible. Thus, the lower river reaches are dominated by physical movement and not
49 chemical modification of sediments.

50 1 Introduction

51 Roughly 1×10^9 tonnes of sediment per year are eroded from the Himalaya and transported through the Ganges-
52 Brahmaputra-Meghna (GBM) delta (Milliman & Farnsworth, 2013). Chemical weathering of silicates in these
53 sediments and subsequent precipitation of carbonates is hypothesized to have caused drawdown of atmospheric CO₂
54 and global cooling in the Cenozoic (Raymo & Ruddiman, 1992) and to increase seawater ⁸⁷Sr/⁸⁶Sr over the last 40
55 Ma (Edmond, 1992). Recent studies have found that chemical weathering of sediments occurs not just in the
56 Himalayan source and in rivers, but also in the upper delta floodplain (Bickle et al., 2018; Lupker, France-Lanord, et
57 al., 2012). This region is “weathering limited” because the rate of chemical weathering is limited by kinetic and
58 climatic factors and not by the supply of unweathered sediment. Because the rate of sediment supply is high due to
59 rapid exhumation and erosion in the Himalayan source, the rate of chemical weathering is high. Readily weathered
60 minerals such as plagioclase persist in the floodplain, and weathering causes measurable changes in the composition
61 of suspended sediments (Lupker, France-Lanord, et al., 2012) and dissolved loads (Bickle et al., 2018) in the Ganges
62 upstream of Harding Bridge in Bangladesh. Chemical weathering rates should decrease downstream as the rate of
63 supply of unweathered primary minerals slows until a transport-limited regime is reached in the lower Ganges tidal
64 plain, as hypothesized by Bickle et al. (2018), but this has not been tested. In this study the compositions of rice
65 paddy soils with approximately known ages, and tidal channel sediments are compared to qualitatively assess the
66 rate of chemical weathering in the Ganges tidal delta plain in Southwest Bangladesh, ~340 km downstream from

67 Harding Bridge (Fig. 1). A small number of upstream and downstream tidal channel water samples are also
 68 compared to see if floodplain weathering in the tidal plain contributes dissolved solutes. We also evaluated the
 69 mobilities of major and trace elements by measuring apparent distribution coefficients of 22 elements between solids
 70 and coexisting water or deionized water extracts, and investigated spatial and seasonal influences.

71 1.1 Setting

72 Bangladesh has a tropical, seasonal climate, with 80% of rainfall occurring during the monsoon from June to
 73 September (N. T. Chowdhury, 2010). The coastal region of Southwest Bangladesh consists of polders, islands
 74 surrounded by and separated from tidal channels by embankments. The tidal channels experience large seasonal
 75 changes in water salinity from ~0 ppt in the wet season to an average of ~15 ppt in the dry season for our study area
 76 (Ayers et al., 2017). Salinity can also vary on a diurnal and biweekly cycle, with higher salinity at spring high tide
 77 due to the increased proportion of seawater from the Bay of Bengal (Ayers et al., 2017). Tidal channels are the
 78 primary source of irrigation water in this region, because the local groundwaters are largely saline (Benneyworth et
 79 al., 2016; Naus et al., 2019). In the wet season when tidal channels are fresh, farmers grow rice in paddies, while in
 80 the saline dry season some switch to brine shrimp aquaculture, while others with access to fresh groundwater use it
 81 to irrigate rice paddies (Barmon et al., 2010).

82 The region of study is in the western portion of the lower GBM delta (Fig. 1). Besides intensively cultivated polders,
 83 it includes the Sundarbans National Forest, the world's largest littoral mangrove forest (Hale et al., 2019). The study
 84 area is an abandoned lobe of the delta, largely cut off from the main Ganges river stem and dominated by tidal
 85 deposits. In this Ganges tidal delta plain, the deposits are considered estuarine because there is a net landward
 86 transportation of sediment that is mostly sourced from the mouth of the GBM river by way of the inner shelf (M.
 87 Allison & Kepple, 2001; Rogers et al., 2013). About half of the deposits are transported directly from the river
 88 mouth along the shelf and then inland by the tides, while half is from upstream reworked deposits (Rogers et al.,
 89 2013). Tidal deltaic deposits define the surficial geology, with upper silt and clay thicknesses from just a few meters
 90 up to 20 to 30 meters (MA Allison et al., 2003; Ayers et al., 2016). No previous studies have investigated how the
 91 composition of these deposits change after deposition.

92 1.2 Factors influencing compositions of solids

93 The mineralogy and chemical composition of sediments deposited by rivers depends on source characteristics; grain
 94 size sorting resulting from hydraulic effects, such as selective entrainment and settling; and chemical weathering
 95 during transport (Garzanti et al., 2010, 2011). In the Ganges River basin the chemical composition of sediments
 96 primarily depends on mineral sorting and, therefore, grain size and density, and secondarily on depletion of mobile
 97 elements by chemical weathering in the floodplain (Lupker, Blard, et al., 2012). In the Sundarbans, where sediment
 98 deposition is unaffected by human activity, the median grain size is 31 µm (medium silt) near the Sutarkhali tidal
 99 river, decreases inland, and does not change seasonally (R P Hale et al., 2019).

100 Tidal islands in SW Bangladesh were converted to polders in the 1960s by building embankments. This reduced the
 101 frequency of flooding and, therefore, the sediment deposition rate on the polders (Auerbach et al., 2015). Today,
 102 tidal channel sediments are deposited on polders only when embankments are breached deliberately for tidal river
 103 management (Gain, 2017), when sluice gates are temporarily opened, or by storm surges that breach the
 104 embankments. For example, in 2009 Cyclone Aila caused long-term tidal flooding on Polder 32 due to embankment
 105 failures, which is a key area of this study (Fig. 1). This caused an elevated mean sediment deposition rate of 17.8
 106 cm/yr (Auerbach et al., 2015). In contrast, the natural background rate of sedimentation at Polder 32 indicates
 107 average Holocene deposition of ~0.6–0.8 cm/y (Ayers et al., 2016). This compares well to modern measurements in
 108 the Sundarbans that yield average annual sediment deposition (vertical accretion) rates of ~1 cm/y (Rogers et al.,
 109 2013), but with seasonality, from ~0.9 cm/y in the dry season to 1.2 cm/y in the wet season for island interiors
 110 (Bomer et al., 2020). We use these deposition rates to estimate mean elapsed time since sediment deposition, and
 111 thus the amount of time rice paddy soils were exposed to chemical weathering.

112 The silt-sized sediments deposited on polders are transported in tidal channels as suspended sediment. Studies of
 113 Ganges River sediments show that suspended sediment mineralogy and composition in the Ganges depends
 114 primarily on grain size, which varies from fine to medium silt at the surface to coarse silt and fine sand above the
 115 channel bed (Garzanti et al., 2011). Common minerals include quartz, feldspar, mica, calcite, dolomite, and epidote-
 116 amphibole-garnet suites, with slow-settling phyllosilicates and the elements they host concentrated near the surface
 117 (Garzanti et al., 2011). However, about half of the sediment deposited in the tidal delta plain is currently sourced
 118 from the Bay of Bengal through tidal exchange, and includes sediments derived from the Brahmaputra, Meghna, and
 119 Ganges Rivers (Rogers et al., 2013; R. P. Hale et al., 2019). When first deposited on polders, sediments are likely to

120 inherit the texture, mineralogy, and composition of suspended sediments as the tidal river overtops or breaches the
121 embankment.

122 The main pedogenic processes that convert sediment to soil in low-lying regions such as SW Bangladesh are (i)
123 homogenization through bioturbation, (ii) mechanical weathering by plant roots, (iii) oxidation-reduction processes
124 of Fe and Mn, (iv) accumulation of organic matter, (v) conversion of primary to secondary minerals such as clays
125 through acid hydrolysis, and (vi) base leaching from calcareous top soils (Hugh Brammer, 2016; Chen et al., 2011).
126 Oxidation of iron under aerobic conditions produces rust colored soils, whereas the reduction of iron when soil is
127 saturated and anaerobic produces grey colored soils (Horneman et al., 2004). Rice paddy soils in the tidal delta
128 plain are often exposed to aerobic conditions in the dry season and anaerobic conditions in the wet season due to
129 waterlogging, resulting in the formation of both rust and grey colored patches in soils (Brammer, 2016). In 2012 on
130 Polder 32 we observed laminations in pits excavated for ponds. These laminations formed by daily tidal inundation
131 following the breach of embankments during Cyclone Aila in 2009 (Auerbach et al., 2015). Soil horizon
132 development was not observed, indicating the soils were entisols (Fry, 2015) and that sediments remained relatively
133 unaltered at least three years after deposition.

134 **1.3 Chemical Weathering in the Ganges Basin**

135 Previous studies of river chemistry in this region have shown that carbonate weathering dominates during the
136 monsoon due to faster dissolution kinetics, while silicate weathering dominates in the dry season due to longer
137 water-rock interaction times and a greater proportion of groundwater baseflow with high solute content (Tipper et
138 al., 2006). In weathering-limited areas in the upper reaches of large watersheds, rates of exhumation, erosion, and
139 chemical weathering are high and fluid residence times are short (Maher, 2010). In the upper reaches of the Ganges
140 Basin, transported sediments are incompletely weathered, and floodplain weathering contributes a substantial
141 fraction of weathering fluxes (Bickle et al., 2018; Lupker, France-Lanord, et al., 2012). However, no previous
142 studies have explored chemical weathering rates in the lower reaches of the Ganges Basin. We hypothesize that in
143 the Ganges tidal floodplain, chemical weathering rates should be transport-limited and lower than upstream in the
144 alluvial fan and delta floodplain due to an inadequate supply of fresh sediment. In the tidal floodplain the rate of
145 chemical weathering may also be low due to the presence of carbonate minerals in the commonly deposited
146 calcareous alluvium – which buffer pH to alkaline values and limit the weathering of minerals by acid hydrolysis
147 (Brammer, 2016) – and by the presence of swelling clays that make surface deposits impermeable (Benson et al.,
148 1994). On the other hand, tilling and fertilization of rice paddy soil may accelerate the process of chemical
149 weathering (Gandois et al., 2011).

150 **1.4 Sampling Sites**

151 Six sites chosen for this study spanned a wide range of surface water salinity and irrigation practices (Patton, 2018).
152 Five sites were primarily used for agriculture, while one located in the Sundarbans mangrove forest represents
153 natural background (Table 1, Fig. 1). The dominant soil type in this region is noncalcareous grey floodplain soil that
154 is seasonally flooded and can develop deeper than 25 cm in noncalcareous alluvium (H Brammer, 2012). The soils
155 on Polder 32 were deposited immediately after Cyclone Aila seven years before sample collection. For sites outside
156 Polder 32 that were not inundated by Aila, the average time since deposition of the top 15 cm is less certain. Given
157 that the long-term average regional sedimentation rate on polders is ~1 cm/y (Rogers et al., 2013), the average age
158 of the top 15 cm would be ~7.5 years, not much different from the age of sampled deposits on Polder 32. However,
159 it's unclear how much the relatively recent conversion of tidal islands to polders by addition of artificial levees, and
160 their periodic breaching, has affected the average deposition rate, adding uncertainty to the age estimate.

161 **2 Materials and Methods**

162 **2.1 Sample Collection and Field Measurements**

163 Solid and water samples were collected at six sites (Fig. 1) in the dry season in May 2016 and wet season in
164 November 2016 in order to describe spatial and temporal variations. Solid and water samples were collected from
165 the same locations in rice paddies, tidal channels, and the Sundarbans. Rice paddy soil samples were collected from
166 the uppermost 15 cm to represent the cultivated topsoil, or ‘root zone’ (Barmon et al., 2010). Rice paddy samples
167 were collected from inundated paddies, while sample type “Dry Rice Paddy” was collected from a part of the paddy
168 without standing water, usually in the dry season. We found that dry rice paddy samples were not significantly
169 different in composition from rice paddy samples with standing water, so they are grouped together. Tidal channel
170 sediment and water samples were collected from as close to the middle of the channel and mid-depth of the water

171 column as possible, usually from irrigation canals connected to tidal channels because the latter were too deep and
 172 wide for sampling. Roughly one kg of solid was collected for each sample.

173 A Decagon GS3 probe was used to measure soil moisture, temperature, and electrical conductivity in-situ in
 174 November 2016, and a Hach HydroLab DS5 Sonde was used to measure water temperature, pH, redox potential Eh,
 175 and specific conductivity (SpC) in-situ. Sample locations were recorded using a Trimble GeoXT 6000, with a
 176 horizontal accuracy of 50 cm.

177 Tidal channel water samples were collected from the middle of the channel and water column, while rice paddy
 178 water samples were collected from the middle of the water column. Water samples were collected by rinsing a one L
 179 bottle, filling it, and immersing a Hydrolab DS5 Sonde for field measurements. Rainwater samples were collected in
 180 clean glass dishes set out just before a rain event. Next, a syringe was used to withdraw 30 mL from either a one L
 181 bottle after field measurements, or a clean glass dish. Water was then filtered at 0.45 µm into a polyethylene sample
 182 bottle. One drop of concentrated nitric acid (HNO_3) was added to the bottle before Inductively Coupled Plasma
 183 (ICP) analysis. Another 60 mL was filtered and placed in a sample bottle without acid for Ion Chromatography (IC)
 184 and Total Organic Carbon (TOC) analysis.

185 2.2 Sample Preparation

186 Solid samples were air dried, sieved with a No. 10 sieve, and homogenized by powdering in a mortar and pestle.
 187 Two to three grams were used to measure loss on ignition (LOI) at 105°C for H_2O - (adsorbed water) content and
 188 1000°C for H_2O + (structural water) content in an alumina crucible (Pansu & Gautheyrou, 2007). A second split for
 189 extract preparation was dried at 110°C. Ten grams of solid was weighed and added to 50 mL of deionized water to
 190 form a slurry. The slurry was stirred for 15 minutes, then measured for pH using an Accumet pH meter and SpC
 191 using a HANNA Portable Solution Conductivity Measurement Meter (Tan, 2005). The saturated slurries were then
 192 filtered using a vacuum pump, coarse porosity/fast flow filter paper, and a Büchner flask and funnel. The extracts
 193 were filtered for a second time through a 0.45 µm syringe filter. Fifteen mL was placed in a conical vial, and 300 uL
 194 of concentrated nitric acid was added to make a 1% HNO_3 acid solution for ICP analysis. Another 30 mL was saved
 195 unacidified for IC and TOC analysis. All water and extract samples were refrigerated until analysed.

196 After LOI measurement, lithium metaborate (LiBO_2) fusion was completed using the same solid samples to measure
 197 their bulk composition. One hundred mg of solid sample and 600 mg LiBO_2 powder were weighed using an
 198 analytical balance and then mixed with an agate mortar and pestle under acetone. The mixture was air dried and
 199 transferred using weighing paper to a new graphite crucible. The crucible was then placed in a preheated oven for 10
 200 minutes at 1100°C. The molten bead that had formed was poured into 50 mL of hot 1 molar (M) nitric acid (HNO_3)
 201 and stirred until the solid material was no longer visible (a minimum of 5 minutes). The cooled sample was
 202 transferred to a 50 ml centrifuge tube and brought to a volume of 50 mL by adding 1 M HNO_3 .

203 2.3 Laboratory Analysis

204 Samples of water, extracts, and dissolved solids were analyzed using methods described in Ayers et al. (2016), who
 205 also list method detection limits. Anions were measured in water and extract samples using ASTM Method D-4327-
 206 03 on a Metrohm 881 Compact IC Pro. Dissolved inorganic carbon (DIC) and dissolved organic carbon (DOC) were
 207 measured in water and extract samples using ASTM Method D-7573-09 on a Shimadzu model TOC-V CPH/CPN.
 208 Major cations were analyzed for using EPA Method 6010B on a Varian ICP Model 14720-ES ICP-OES. Trace
 209 metals were measured using a Perkin Elmer model ELAN DRC II ICP in both standard and dynamic reaction
 210 chamber (DRC) modes.

211 Select solid samples were analyzed for particle size using a Malvern Mastersizer 2000. A homogeneous subset of
 212 each solid sample analyzed was placed in a 1 L beaker of deionized water after running deionized water through the
 213 instrument to establish background concentrations. Sample were then deflocculated while suspended in deionized
 214 water via sonication before analysis. Samples were analyzed for the grain size range of 0.563 µm to 1.27 mm.

215 Minerals in select solid samples were identified using powder x-ray diffraction at Vanderbilt University. Dried,
 216 homogenized powder samples were dry cast on 20 x 20 mm square sample glass holders with a 0.2 mm indent and
 217 analyzed with a Rigaku SmartLab powder X-ray diffractometer with a Cu source of X-rays with a $\text{K}\alpha$ wavelength of
 218 0.154 nm at 40 kV and 44 mA. The detector was a D/teX Ultra 250 1D silicon strip detector. PDXL software
 219 identified the mineral phases associated with the intensity peaks.

220 Saturation index calculations were performed using the Spec8 program in Geochemists Workbench v.8 with the
 221 default thermo.dat database (Bethke, 2007). Data analysis was performed in R (R Core Team, 2019). Spatial
 222 analysis was conducted using ArcGIS Pro.

223 2.4 Quality Assurance/Quality Control

224 Deionized water was used for water sample blanks to test for water contamination in the field, and replicates of
 225 select samples were collected in order to quantify precision. Water samples were kept refrigerated and were
 226 analyzed less than one month after collection. Our previous work has shown that, except for nitrate ions, measured
 227 concentrations do not change for at least one month after sample collection (Ayers et al., 2017). Charge imbalance
 228 errors for water samples averaged 4.1%, with the maximum being 21%. Measured concentrations for species
 229 reported here were well above the method detection limit (given in Ayers et al., 2016) for most elements, and
 230 sample concentrations were always significantly higher than blank concentrations. For replicate water samples, the
 231 average percent difference in measured major element concentrations was 1.6%, indicating good precision. For
 232 water samples, total dissolved solids calculated from chemical analyses showed excellent correlation with values
 233 measured using the Hydrolab (adj. $r^2 = 0.994$, $n = 52$). For extract solutions a comparison of major elements in
 234 duplicate sample analyses shows an average percent difference of 8.7%. Charge imbalance on extract analyses
 235 averaged 0.68%, with a maximum of 9.0%.

236 Bulk solid procedures were also performed using $\geq 99.9\%$ SiO₂ to test for contamination. Except for SiO₂, oxide
 237 concentrations in these bulk solid blanks were on average 95% lower than concentrations measured in samples.
 238 USGS reference powders AGV-2 and GSP-2 were run as samples through the solid preparation procedures to test
 239 for accuracy, and duplicates were completed to quantify precision. Normalized AGV-2 major oxide concentrations
 240 were on average 12.6% different than the USGS reference, while GSP-2 oxides resulted in an average difference of
 241 5.8%. Most trace element concentrations were within 20%. Duplicates show an average difference of 0.54% over all
 242 elements, indicating minimal error in sample preparation and analysis.

243 2.5 Estimation of Distribution Coefficients

244 The solid-water bulk distribution coefficient $D^{s/w}_i$ is defined as the concentration of element i in the solid C_i^s divided
 245 by that in the water C_i^w , both in ppm. $D^{s/w}$ values were estimated by analyzing water and solids collected at the same
 246 time and location. Only samples from rice paddies were included, as tidal channel water was rapidly flowing over
 247 tidal channel sediment and therefore unlikely to reach exchange equilibrium. Solid/extract $D^{s/e}_i$ values were
 248 estimated two ways. In the first, the measured bulk concentration of element i in the solid C_i^s was used with the
 249 measured concentration in the extract solution C_i^e , such that $D^{s/e}_i = C_i^s/C_i^e$. This assumes that the solid concentration
 250 was unaffected by equilibration with the extract solution. In the second method, the same information was used with
 251 mass balance to estimate the concentration in the residual solid C_i^{res} , which accounts for the amount of i that
 252 dissolved in the extract solution, as follows: $D^{s/e}_i = C_i^{res}/C_i^e$. Because measured $D^{s/e} \gg 10$ for all elements, the
 253 concentration in the solid did not change significantly even for soluble elements, so calculating $D^{s/e}$ using the two
 254 methods yielded nearly identical results, but the more accurate values estimated using mass balance were used. $D^{s/e}$
 255 values were estimated for all solid types: tidal channel sediment, rice paddy soil, and Sundarbans soil.

256 3 Data

257 Sample site locations are shown in Fig. 1 and listed in Table 1. Element concentrations grouped by sample type are
 258 tabulated for water samples (Table 2), extracts (Table 3), and solids (Table 4). Concentrations of the elements Be,
 259 Cd, Pb, and Tl were frequently below our method detection limits, and so are not reported. Furthermore, our
 260 procedure for analyzing solids prevented us from measuring sample concentrations of elements that mainly form
 261 anions (since the 1% HNO₃ solution could not be introduced into the ion chromatograph) or were present in the flux
 262 (Li and B). Concentrations reported as negative were deleted. Although some positive concentrations for the
 263 remaining elements were below the method detection limit, we chose to use the data uncensored with no
 264 substitution, as removing concentrations below the detection limit would bias the results to higher concentrations.
 265 Sample types included tidal channel (TC), rice paddy (RP), and Sundarbans (SB). Some samples classified as “tidal
 266 channel” were from irrigation canals connected to tidal channels.

267 For water samples collected in the same season (dry season in May and wet season in November), and extracts
 268 prepared from solid samples collected in the same season, element concentrations were most often lognormally
 269 distributed, as was reported for water samples from the same region by Ayers et al. (2017). All statistical tests and
 270 plots therefore use \log_{10} values of concentrations. Although for solids many elements were not fit well by a
 271 lognormal distribution, we used a log transform on solid concentrations for consistency, and nonparametric
 272 statistical tests were used. Uncertainties are reported as one standard deviation (1σ).

273 Across sample sites, in-situ soil salinity slightly increased to the SE and with increasing proximity to the Bay of
 274 Bengal. In November, rice paddy water salinity slightly increased to the SSW. In general, soil and water salinity

decrease to the North and with increasing distance from the coastline. Although the sampling area was ~1300 km² (Fig. 1) and covered at least five different soil regions (Table 1), the variation in compositions of water and solid samples between sites was small within each season (e.g., Fig. S1), indicating similar processes are acting to transport elements across this region. We therefore lump data for all sites together in the following plots and discussion.

Values of specific conductivity (SpC), a proxy for salinity, of extracts were similar to in-situ solid conductivity measurements because the extract measurement was made on the solid slurry, not the filtered extract solution (Fig. 2). Salinity was much higher in solids and extracts than water samples (Fig. 2). Solids and extracts also show much less seasonality than water samples, which have higher salinity in the dry season.

To compare the wide range of element concentrations in solid, extract, and water, average concentrations were normalized to average upper continental crust (UCC) in spiderplots (Fig. 3) with elements arranged on the x-axis from lowest mobility during weathering and transport on the left to highest mobility on the right (Gaillardet et al., 2014). Tidal channel sediment (Fig. 3a) and rice paddy soil (Fig. 3b) are similar in composition to average UCC but depleted in Ca, Sr, Na, As, and S. Extract and water samples show strong depletions in immobile elements and weak depletions in the mobile elements relative to solids. Extracts have higher concentrations than tidal channel and rice paddy water for most elements.

3.1 Water

In tidal channel water samples, variations in total dissolved solids (TDS) and SpC reflect variable contributions of seawater from the Bay of Bengal (Ayers et al., 2017). Calculated TDS in water samples ranged from 20 ppm in rainwater to as high as 20,276 ppm in May tidal channel samples. Tidal channel water has much higher salinity in the dry season than the wet season (Fig. 2). Wet season tidal channel and rice paddy water have similar salinity and element concentrations, consistent with observed irrigation of rice paddies with tidal channel water in the wet season (Ayers et al., 2017). In the dry season, most rice paddies have lower salinity than tidal channel water because they are irrigated with groundwater. Concentrations of conservative elements, which show a constant ratio with Cl concentrations, are consistent with these relationships inferred from salinity. Immobile elements such as Al, Fe, Zn, and Cr show little variation in concentration by season or location (as measured by the size of the error bars in Fig. 3), while redox sensitive elements such as Mn and S show larger variations (Table 2). Saturation index calculations on seasonal average water compositions show that rice paddy and tidal channel waters are always saturated in hydroxyapatite, dolomite, muscovite, and the smectite mineral saponite, and usually saturated in illite, kaolinite, calcite, and K-feldspar. Dry season waters are also saturated in the smectite mineral beidellite.

3.2 Extracts

Like water, extracts show little variation in concentrations of immobile elements such as Al, Fe, Zn, Cr, and Ni by season or location (Table 3, Fig. 3). Soluble elements such as Na and redox sensitive elements such as Mn and S show relatively large seasonal variations. Extracts from tidal channel sediments collected in the dry season generally have higher concentrations and SpC values than wet season extracts, though the same is not true for rice paddy samples. Extract pH was nearly always > 7.5 with little seasonal variation, consistent with the presence of carbonate minerals. Sundarbans and tidal channel sediment samples have higher extract TDS and SpC than rice paddies, suggesting that agricultural activity may flush out soluble salts (Fig. 2).

Comparison of sulfur in extracts measured as elemental S by ICP-OES and sulphate SO₄²⁻ measured by IC shows the two are highly correlated ($r = 0.997$). The median molar concentration was higher for S measured by ICP-OES at 6.4×10^{-4} than SO₄²⁻ measured by IC at 6.1×10^{-4} , suggesting a small portion was present as sulfide, and that most S is present as sulphate.

Saturation index calculations show that rice paddy soil and tidal channel sediment extracts are always saturated in hydroxyapatite and kaolinite, and usually saturated in illite, gibbsite, smectite minerals beidellite and saponite, and dolomite ± quartz and calcite. Extracts are undersaturated in hydrous ferric oxyhydroxides that often control the mobility of arsenic and other elements in this region (Hossain et al., 2012). This may be due to their absence in the solids, to disequilibrium, or to their abundance being less than required to saturate the volume of extract solution.

3.3 Solids

The composition of sediment is strongly influenced by grain size (Garzanti et al., 2011). The mode, median, and volume-weighted mean grain sizes for select solid samples from this study are given in Table 5. The average

326 volume-weighted mean of rice paddy soils was 50 μm , tidal channel sediments 23 μm , and the single Sundarbans
 327 sample 48 μm . The median for the Sundarbans sample was 36 μm , similar to the 31 μm previously reported (R P
 328 Hale et al., 2019). We did not measure grain size distributions in enough samples to characterize spatial or seasonal
 329 variability. However, our limited measurements demonstrate that the sampled solids are dominantly silt-sized and
 330 roughly comparable in size, although the rice paddy soil samples are slightly coarser than the tidal channel sediment
 331 samples. Our tidal channel sediment samples are likely finer because they were collected from irrigation canals
 332 connected to the tidal channels. The lower flow velocity in the canals likely caused their bedload sediments to be
 333 finer than tidal channel bedload sediments and more similar to tidal channel suspended sediments.
 334 Powder x-ray diffraction was used to identify minerals in two samples that were also characterized for grain size
 335 distribution: rice paddy soil sample BEMS_08_RPS_06M_May and tidal channel sediment sample
 336 BEMS_TC_01_May. Although the rice paddy sample is significantly coarser (Table 5), the minerals and their
 337 relative proportions assessed qualitatively by peak heights are very similar (Fig. 4). Quartz, illite, chlorite, biotite,
 338 and dolomite have multiple peaks in the 2 θ range of 10–45°. Sample treatment is required to identify smectite. More
 339 detailed powder x-ray diffraction measurements have shown that in Ganges River sediments illite is the most
 340 abundant clay, followed by smectite, kaolinite, and chlorite (MA Allison et al., 2003; Khan et al., 2019).
 341 Average tidal channel sediment and rice paddy soil concentrations of immobile elements are similar to those of UCC
 342 (Fig. 5). Converting to oxide concentrations and normalizing to an anhydrous basis again shows the similarities to
 343 UCC, although FeO is lower and CaO and MgO higher than UCC, most likely due to sediment sorting (Table 6).
 344 However, major elements with intermediate mobilities including Ca, Sr, and Na are depleted, likely due to depletion
 345 of plagioclase feldspar through sedimentary sorting or chemical weathering (Garzanti et al., 2011; Lupker, France-
 346 Lanord, et al., 2012). Mobile elements show variable behavior, with As and S being low relative to UCC. Compared
 347 to water and extracts, solid samples show less variation in element concentrations across seasons and sampling sites
 348 as measured by relative error bar size in Figs. 2 and 3.
 349 Normalizing concentrations in rice paddy soil to tidal channel sediment highlights post-depositional changes in solid
 350 composition (Fig. 6). The only elements that show a substantial difference in concentration between rice paddy soil
 351 (C^{rp}) and tidal channel sediment (C^{tc}) were Na, with higher concentrations in May tidal channel sediment, and Si,
 352 which is higher in November rice paddy soil (Mann-Whitney test p values < 0.05). In both cases the 1σ error bars
 353 overlap with $\log_{10}C^{\text{rp}}/C^{\text{tc}} = 0$, indicating the differences in concentration are very small (Fig. 6).

354 4 Results

355 4.1 Element Correlations

356 Examination of element correlations can shed light on chemical processes. We found statistically significant
 357 correlations (p value for Pearson correlation coefficient < 0.05) for 160 of 406 element pairs in soil extracts (Figs.
 358 S3 and S4), 166/276 pairs in solids (Fig. S2), and 285/378 pairs in water. Log concentrations of conservative
 359 elements such as Na, Cl, K, Mg, and Sr in extract solutions are normally distributed and are all positively correlated,
 360 consistent with simple dissolution of soluble salts (Fig. S3). The same is true for water samples collected in this
 361 region in this study and previous studies (Ayers et al., 2017). In contrast, conservative elements do not always show
 362 good correlations in bulk solids (Fig. S2), which have much higher concentrations of all elements (except S) than
 363 water or extracts (Fig. 3). This indicates that only a small portion of most cations are present in solids as soluble
 364 salts, and generally cations remain bound to solid minerals such as clays during extract preparation or equilibration
 365 with standing water.
 366 In contrast to conservative elements, log concentrations of nonconservative elements in water and extracts are not
 367 always normally distributed and can show positive, negative, or no correlation with other elements (Fig. S4).
 368 Arsenic shows a weak positive correlation with S in extracts, suggesting it may have been present in sulfides.
 369 However, if sulfide oxidation occurred during extract preparation, it was insufficient to reduce pH to < 7.
 370 In bulk solids the metals Ba, Co, Cr, Cu, Fe, K, Mg, Mn, Na, Ni, Si, and Zn show strong ($r > 0.8$, $p < 0.05$)
 371 correlations with Al, indicating they are stored in clay minerals. Arsenic has a strong positive correlation with Fe (r
 372 = 0.78, $p = 2.6\text{E-}6$), suggesting it may be sorbed onto hydrous ferric oxyhydroxides (HFOs) (Fendorf et al., 2010). If
 373 HFOs are present, it is in abundance too low to detect by powder XRD or to saturate the extract solution. In tidal
 374 channel and rice paddy water samples, most elements show positive correlations with TDS, reflecting variable
 375 amounts of seawater mixing, dilution, and evaporative concentration.

376

4.2 Element Partitioning

In soils and sediments, weathering reactions between porewaters and minerals will cause dissolution of primary minerals, precipitation of secondary minerals, and adsorption onto mineral surfaces. Trace elements can be incorporated into solids through adsorption or coprecipitation/substitution. The partitioning of trace element *i* incorporated into multi-phase solids by substitution can be described using the solid-water bulk distribution coefficient, which has the same form as the linear adsorption isotherm: $D^{s/w}_i = C^s_i/C^w_i$. For trace elements, a plot of concentrations in water versus solid should yield a linear array with a positive slope and a zero intercept if Henry's Law is obeyed and solid mineralogy does not change, and the slope should equal the value of the trace element distribution coefficient $D^{s/w}$. This should be true regardless of whether the element is incorporated in solids by coprecipitation (substitution) or adsorption. While D values should be constant only for trace elements, they are still useful for evaluating whether major elements partition more strongly into solid or solution. $D^{s/w}$ values were calculated only for samples of water and rice paddy soil collected at the same time and location. $D^{s/e}$ values were calculated for extracts prepared from the same rice paddy soil or tidal channel sediment sample. D values are roughly lognormally distributed for each element. For most elements, $D^{s/w}$ values are similar to $D^{s/e}$ values (Fig. 7). D values calculated from average concentrations in Tables 2-4 are similar to averages from paired samples, suggesting our results are robust and not site-specific. Also, inter-element variances are larger than intra-element variances, suggesting that D values represent relative element mobilities (Fig. 7).

We observed poor and often negative correlations between concentrations of an element in solid and water or extract. The poor correlations may result from variable water compositions (especially pH) and solid mineralogies, although these would not likely cause negative correlations between concentrations in solids and aqueous solutions. Another potential explanation is disequilibrium partitioning. This is supported by seasonal changes in composition that are large for water and small for soil. From the dry season in May to the wet season in November, metal concentrations in soil and sediment change little (Fig. 5, larger error bars in Fig. 3 for water than for solids), despite surface water compositions changing dramatically from saline to fresh (Fig. 2). The average percentage seasonal difference in element concentrations were 147% and 13% for water and soil, respectively. Disequilibrium exchange is also supported by a lack of systematic dependence on pH, which should strongly affect mineral solubilities and adsorption of dissolved ions, and by the oversaturation of many minerals in extract solutions, including quartz. Also, the 15 minutes that the solid and extract solution remained in contact (the recommended duration for soil extracts cf. Tan (2005)) was likely insufficient to reach exchange equilibrium. Previous studies have shown that while > 75% of metal adsorption in batch experiments occurred within the first 30 minutes, some metals could take up to 300 minutes to reach a steady state solution concentration (Zhang et al., 2019).

While our measured distribution coefficients cannot be considered equilibrium D values, they are still useful as measures of element mobility and for identifying processes that affect element mobility. For sediments and soils in this region that have similar mineralogy, our D values can be used to estimate concentrations in solids from concentrations in standing water or extracts, which are easier to measure. This would be particularly useful for estimating concentrations of toxic elements such as As in agricultural soils.

As expected, D values are highest for the least mobile elements Al and Fe, and lowest for the most mobile element S (Fig. 7). Based on the mobility ranking (the order of elements on the x-axis), D is higher for Si and lower for Ca, Sr, and Na than expected. The higher D for Si is probably due to the relatively high proportion of quartz in the silt-sized sediments and soil (Fig. 4), while low D values for Ca, Sr, and Na likely result from removal of plagioclase feldspar by sedimentary sorting or weathering.

417

4.3 Sediment and Soil Compositions

Chemical weathering during transport and following sediment deposition converts primary minerals into dissolved solutes and secondary minerals, which changes the chemical compositions of solids and fluids. Weathering of sediments causes the following changes in mineralogy of sediments/soils and major element compositions of fluids: conversion of albite and K-feldspar to kaolinite or smectite releases Na and K respectively; conversion of biotite to vermiculite releases K; and calcite and dolomite dissolution release Ca and Mg (Lupker, France-Lanord, et al., 2012). Previous work has shown that progressive weathering of sediments near the Himalayan Front and the Ganges floodplain depletes sediments in the mobile elements that are released by weathering reactions (Lupker, France-Lanord, et al., 2012), particularly elements released by soluble minerals such as carbonates that dissolve rapidly. As a result, alkalis and alkaline earths in the suspended load of the Ganges River are depleted relative to UCC, even in the upper reaches of the Ganges Basin (Garzanti et al., 2011). We also observe depletion in alkaline earths like Ca and Sr and alkalis like Na, which could be caused by the removal of low-density plagioclase feldspar and its Na and

429 Ca components by chemical weathering (Fig. 5, Table 5, lack of plagioclase feldspar in x-ray diffraction patterns in
 430 Fig. 4). In contrast, tidal channel sediments and rice paddy soils both show enrichment in K relative to UCC (Table
 431 5), probably due to illite being the most abundant clay mineral (Fig. 4).
 432 At least half of the sediment deposited in the Ganges tidal plain is derived from the mouth of the main stem Meghna
 433 River (Rogers et al., 2013), which combines the drainage of the Ganges, Brahmaputra, and Meghna Rivers, which
 434 collect water and sediment from all the southern Himalaya. It is, therefore, not surprising that our sediment and soil
 435 samples from the Ganges tidal plain have trace element compositions very similar to average UCC (Fig. 3).
 436 However, plotting only solids highlights small differences between trace element concentrations in solids and UCC
 437 (Fig. 5). Compared to rice paddy soils and tidal channel sediments, normalized concentrations of many trace
 438 elements in Sundarbans soils are higher in the wet season and lower in the dry season (Fig. 5), although we have
 439 only one sample each of wet and dry season Sundarbans soils (cf. Table 4).
 440 Poldering in coastal Bangladesh has greatly reduced the frequency of tidal flooding and sediment deposition. On
 441 Polder 32, the last significant sediment deposition event started with Cyclone Aila in 2009 and ended after
 442 embankment repairs were completed in 2010 (Auerbach et al., 2015). After roughly seven years, element
 443 concentrations in Polder 32 rice paddy soil are only slightly different from tidal channel sediment (Fig. 8), indicating
 444 little effect of in-situ chemical weathering.
 445 Rice paddy soil and tidal channel sediment have similar mineralogy (Fig. 4), major element compositions (Table 6),
 446 and trace element compositions (Figs. 6 and 8). These similarities can only occur if two conditions are met:
 447 sediments deposited inside the polder are similar in composition to sediments deposited in the tidal
 448 channels/irrigation canals; and the influence of post-depositional weathering on mineralogy and composition is
 449 insignificant. Sediment deposition within polders occurred either before embankments were constructed in the 1960s
 450 or during tidal inundation following embankment breaches (Auerbach et al., 2015). None of the sampled areas have
 451 been affected by tidal river management, which involves deliberate flooding to induce sedimentation and an increase
 452 in land elevation to reduce waterlogging (Gain, 2017). Tidal inundation would cause deposition of silt-sized
 453 suspended sediments and not sand-sized bedload sediments within a polder, since silt is the dominant grain size in
 454 suspended sediments of most rivers (Garzanti et al., 2011), and is the dominant grain size in tidal channel sediments
 455 and polder soils within the study area (Table 5). Since polder soils have grain size distributions, mineralogies, and
 456 compositions similar to sediments deposited in tidal channels (although the measured polder soils were slightly
 457 coarser cf. Table 5), the effects of post-depositional chemical weathering within the polders have been insignificant.
 458 An alternative explanation is that since our tidal channel sediment samples were actually collected mostly from
 459 irrigation canals, the rice paddy soils are actually sediments deposited from irrigation water, and that this process is
 460 continuous. This could explain why our rice paddy soil and tidal channel sediment samples have similar
 461 mineralogies and compositions. However, our RP soil is coarser than the TC sediment, and if it was deposited by
 462 irrigation water, it would be finer. Furthermore, water in the irrigation canals moves very slowly due to minuscule
 463 changes in elevation, so that water entering rice paddies has near zero suspended sediment. If sediment deposition
 464 from irrigation water was significant, then irrigation canals would quickly clog with sediment, and rice paddies
 465 would be at higher elevation than adjacent land, neither of which was observed. Finally, a trench dug into surficial
 466 sediments inside Polder 32 showed that all near-surface sediment showed tidal laminations, indicating sediment was
 467 deposited during tidal inundation following Hurricane Aila.
 468 A larger Bangladesh dataset of soil samples collected throughout Bangladesh showed that element concentrations in
 469 paddy and non-paddy soils are positively correlated and similar (M. T. A. Chowdhury et al., 2017). Linear
 470 regression of element concentrations in non-paddy versus paddy soils yielded a slope of less than one for all
 471 elements but arsenic, indicating that only arsenic was added to paddy soils by groundwater irrigation (M. T. A.
 472 Chowdhury et al., 2017). It is unclear whether the non-paddy soils in that study are comparable to tidal channel
 473 sediments in this study, but both studies demonstrate that the influence of rice paddy agriculture on soil composition
 474 in Bangladesh is small.

475 4.4 Chemical Weathering Rates

476 Our results suggest that element concentrations change more slowly in solids than in surface water as seasons
 477 change. Water-solid exchange rates must be too slow for solids to equilibrate with surface waters, which have short
 478 residence times and experience seasonal changes in composition. Solid exchange was also too slow to significantly
 479 modify the composition of the top 15 cm of soil on Polder 32 during the seven years between sediment deposition
 480 after Cyclone Aila in 2009 and our collection of rice paddy soils in 2016 (Fig. 8), although we see seasonal changes
 481 in concentrations of some elements, especially soluble salts (Table 4).

482 As discussed in Section 1.2, the average age of the top 15 cm at sites other than Polder 32 is estimated at ~7.5 years
483 since deposition after Hurricane Aila in 2009 and sample collection in 2016. For polders other than Polder 32 that
484 we collected samples from, it is possible that embankment breaches occurred during Hurricane Aila, but unlikely
485 that they occurred after Aila and before sample collection. If Aila did not cause embankment breaches in these
486 polders, then collected soil samples would be older than 7.5 years, since sedimentation rates in the region have
487 decreased since poldering in the 1960s, but not by more than a few decades. During that time, chemical weathering
488 had little effect on the composition of rice paddy soils at any of the investigated sites, indicating that the rate of
489 chemical weathering is very low. The lack of a change in solid compositions across our field sites is mirrored by no
490 significant change in November tidal channel water composition between upstream (4 samples from sites B3, B8,
491 and B9) and downstream locations (4 samples from sites P32, B2, and SB), which are separated by ~ 50 km (Fig. 9).
492 While changes in element discharge (solute fluxes) are needed to estimate the contribution from chemical
493 weathering, the water discharge likely changes little over this 50 km stretch, so a lack of change in element
494 concentrations is consistent with a lack of chemical weathering inputs. This is also supported by previous work in
495 this region that found concentrations of conservative elements such as Na, Mg, Sr, and Cl in surface waters are
496 explained by simple mixing of rainwater and seawater, with no contributions from chemical weathering (Ayers et
497 al., 2017). The evidence suggests that most or all dissolved solutes in tidal channel water in the study area are
498 derived from other areas, including the fluvial-dominated delta upstream and the Bay of Bengal downstream, and
499 that chemical weathering rates in the study area are low.

500 Far upstream in the Ganges River Basin, in the Himalaya and the alluvial fan, weathering is rapid because
501 exhumation and erosion provide a continuous supply of fresh sediments. In this weathering-limited regime in the
502 headwaters of the Ganges, the chemical weathering rate is directly proportional to the physical erosion rate (Bickle
503 et al., 2018; West et al., 2005), and river sediments show continuous downstream increases in the Chemical Index of
504 Alteration, increases in clay concentrations at the expense of biotite and feldspar, and decreases in Th activity ratios
505 and U/Th caused by chemical weathering (Granet et al., 2007). River waters in the upper reaches are less likely to
506 be calcite-saturated than in the lower reaches (Sarin et al., 1989), while solute concentrations increase downstream
507 from the alluvial fan before leveling off (Bickle et al., 2018). Sediment is weathered during transport as it works its
508 way through the alluvial fan, which may be intermittent due to deposition and reworking in a braided stream
509 network. In this weathering-limited regime, dissolved and suspended sediment fluxes increase downstream. Far
510 upstream of our study area at Farraka, India, just upstream of Harding Bridge (Fig. 1), changes in Ganges River
511 dissolved solute fluxes indicate that floodplain weathering contributes roughly half of major dissolved cation fluxes
512 (Bickle et al., 2018), consistent with estimates based on sediment composition (Lupker, France-Lanord, et al., 2012).
513 In contrast, in transport-limited areas the rate of chemical weathering is limited by the rate of supply of fresh
514 material, and is lower than in the weathering-limited regime (Bickle et al., 2018; West et al., 2005). Our data suggest
515 that a transport-limited regime prevails in the lower reaches of the Ganges Basin, consistent with predictions of a
516 transition from weathering-limited regimes in highlands to transport-limited regimes in lowlands (Bickle et al.,
517 2018; Gaillardet et al., 1999). In the lower river reaches, such as the tidal delta plain where our sampling sites are
518 located, the solid material is more weathered than the fresh material in the alluvial fan. In the more intensely
519 weathered sediment in the lower reaches, reactive/soluble primary minerals have mostly been replaced by secondary
520 minerals such as clays, consistent with our observed sample mineralogies being dominated by the clay mineral illite
521 and insoluble minerals such as quartz (Fig. 4). In the lowlands, water residence times are longer than in the
522 highlands, especially in the floodplain, and tidal channel water that is used for irrigation has high dissolved solids,
523 which lowers the chemical weathering rates of minerals. Water samples from rice paddies and tidal channels are
524 saturated in the dominant minerals observed in our solid samples (illite, dolomite) and minerals that commonly
525 occur in sediments in the lower reaches of the Ganges River basin (smectites and kaolinite), indicating that these
526 minerals will not dissolve in the waters their host solids come in contact with. These waters are even saturated in
527 primary minerals such as K-feldspar, muscovite, and hydroxyapatite. Extract solutions, which are likely more
528 similar to soil porewater compositions than standing water in rice paddies, are even more concentrated in dissolved
529 solutes, and saturated in the same minerals. The fact that water samples are already saturated in most of the minerals
530 they would come in contact with is strong evidence that the driving force for chemical weathering is low. This
531 explains our observed lack of change in soil composition over time (as compared with fresh tidal channel sediment),
532 and lack of change in tidal river water composition across our study area. Somewhere between our downstream
533 study area and Harding Bridge upstream must be the transition from weathering-dominated to transport-dominated
534 regimes. For the main stem Ganges, the transition may be located near the break in river slope that separates the
535 alluvial fan from the fluvial-tidal delta plain and causes a decrease in water velocity and average grain size
536 downstream (Wilson & Goodbred, 2015).

537 Chemical weathering is transport-limited when there is a linear relationship on a logarithmic plot of physical erosion
 538 rate versus chemical weathering rate. These relationships are determined using mass discharge of dissolved solutes
 539 and sediments in main-stem rivers in the fluvial-dominated part of the delta (Bickle et al., 2018). Our study area is in
 540 the tidal floodplain, where it is less clear what physical erosion rate means, and which we have not measured.

541 However, our data show that the chemical weathering rate in the tidal delta plain is low, which is more consistent
 542 with a transport-limited regime than a weathering-limited regime. The area of the delta in the transport-limited
 543 regime should increase with factors that increase the rate of chemical weathering such as temperature and runoff
 544 (West et al., 2005). Since both of these factors are high in the GBM delta, it seems likely that the entire tidal delta
 545 plain is in the transport-limited regime.

546 The instantaneous weathering rate of solid material decreases with increasing time spent in the weathering
 547 environment. The residence time of organic carbon in river sediments in the upstream fluvial-dominated part of the
 548 Ganges Basin in Bangladesh is ~2 kyr on average (Galy & Eglinton, 2011). This suggests that much of the sediment
 549 was stored in the floodplain for over one thousand years, so that dead organic matter could accumulate in soil before
 550 reworking released it to the river. The amount of time elapsed since physical erosion in the Himalaya to deposition
 551 is even greater than one thousand years for the tidal floodplain, since much of the sediment was transported all the
 552 way to the Bay of Bengal, potentially stored there for a significant amount of time, and then transported inland by
 553 the tides before deposition. Thus, by the time sediments are deposited in our field area, they have been thoroughly
 554 weathered.

555 While the older sediment ages and higher solute loads in the tidal delta plain decrease the Gibbs Free Energy change
 556 and rates of chemical weathering reactions, the lack of systematic partitioning (and the oversaturation in many
 557 minerals) shows that solids and surface waters are not in chemical equilibrium. Solids show little seasonal variation
 558 in composition, while surface waters show large variations. Disequilibrium between standing water and the top 15
 559 cm of sediment and soil may result from the solids being dominated by swelling clays such as smectites that are
 560 abundant in the Ganges clay-sized fraction (Borromeo et al., 2019) and that make surface deposits impermeable
 561 (Benson et al., 1994). Swelling will occur in the wet season due to abundant water for interlayer sites in smectites
 562 such as beidellite, while in the dry season abundant Na⁺ may displace Ca²⁺ and Mg²⁺ and cause swelling (Langmuir,
 563 1997). Swelling clays may therefore inhibit chemical weathering reactions in the soil.

564 5 Conclusions

565 Rice paddy soil and tidal channel sediment in the Ganges tidal plain in SW Bangladesh have similar mineralogies
 566 and compositions. Since soils formed from tidal channel sediments deposited at least 7.5 years ago, the lack of
 567 change in composition during that time indicates low rates of chemical weathering characteristic of a transport-
 568 limited regime, in contrast with the documented weathering-limited regimes > 340 km upstream (Bickle et al., 2018;
 569 Lupker, France-Lanord, et al., 2012). Rice paddy and tidal channel waters are saturated in the dominant minerals
 570 present in soils and sediments, indicating low Gibbs Free Energy changes and rates of weathering reactions.
 571 Tidal channel sediment and rice paddy soil show little or no seasonal variation in composition. In contrast, the
 572 composition of water in rice paddies and tidal channels does vary seasonally, with wet season samples being more
 573 dilute. This suggests that compositions of solids change more slowly than solutions, which explains much of the
 574 intra-element variation in our measured apparent solid/water distribution coefficients. While measured distribution
 575 coefficients are not equilibrium values, the relative values for the elements studied are mostly consistent with
 576 previously published trends of element mobility during weathering and transport (Gaillardet et al., 2014), and values
 577 can be used with measured water compositions to estimate compositions of soils and sediments with similar
 578 mineralogies. Disequilibrium partitioning between fine-grained sediments and surface waters may be a common
 579 feature in river deltas due to the presence of swelling clays, the low supply of unweathered material, and saturation
 580 of most available minerals in water, which together keep chemical weathering rates low.

581 Acknowledgments, Samples, and Data

582 The authors have no financial conflicts of interest related to this paper. Thanks to Steve Goodbred for advice,
 583 Rossane DeLapp for quickly analyzing our samples, Md Nazrul Islam (Bachchu) from Pugmark Tours, Saddam
 584 Hossain from Dhaka University, and Chelsea Peters from Vanderbilt University for help in the field. Any opinions,
 585 findings, conclusions, or recommendations expressed in this material are those of the authors and do not necessarily
 586 reflect the views of the National Science Foundation.

587 Research Data associated with this article can be accessed at <http://dx.doi.org/10.17632/6z6bdxrkkb.2>

588 **References**

- 589 Allison, M., & Kepple, E. (2001). Modern sediment supply to the lower delta plain of the Ganges-Brahmaputra
 590 River in Bangladesh. *Geo-Marine Letters*, 21(2), 66–74. <https://doi.org/10.1007/s003670100069>
- 591 Allison, MA, Khan, S., Goodbred, S., & Kuehl, S. (2003). Stratigraphic evolution of the late Holocene Ganges–
 592 Brahmaputra lower delta plain. *Sedimentary Geology*, 155, 317–342. [https://doi.org/10.1016/S0037-0738\(02\)00185-9](https://doi.org/10.1016/S0037-0738(02)00185-9)
- 593 Auerbach, L. W., Goodbred Jr, S. L., Mondal, D. R., Wilson, C. A., Ahmed, K. R., Roy, K., et al. (2015). Flood risk
 594 of natural and embanked landscapes on the Ganges–Brahmaputra tidal delta plain. *Nature Climate Change*,
 595 5(2), 153–157. <https://doi.org/10.1038/nclimate2472>
- 596 Ayers, J. C., Goodbred, S., George, G., Fry, D., Hornberger, G., Benneyworth, L., et al. (2016). Sources of salinity
 597 and arsenic in groundwater in southwest Bangladesh. *Geochemical Transactions*, 17(4), 1–22.
 598 <https://doi.org/10.1186/s12932-016-0036-6>
- 599 Ayers, J. C., George, G., Fry, D., Benneyworth, L., Wilson, C., Auerbach, L., et al. (2017). Salinization and arsenic
 600 contamination of surface water in southwest Bangladesh. *Geochemical Transactions*, 18(1), 4.
 601 <https://doi.org/10.1186/s12932-017-0042-3>
- 602 Barmon, B. K., Kondo, T., & Osanami, F. (2010). Rice-prawn Farming System: Impacts on Soil Quality and Land
 603 Productivity of Modern Variety Paddy Production in Bangladesh. *Asian Journal of Agriculture and
 604 Development*, 7(2), 49–66. Retrieved from http://beta.searca.org/searca/ajad/files/101111132631_Barmon,_Kondo,_Yamaguchi,_and_Osanami.pdf
- 605 Benneyworth, L., Gilligan, J., Ayers, J., Goodbred, S., George, G., Carrico, A., et al. (2016). Drinking water
 606 insecurity: water quality and access in coastal south-western Bangladesh. *International Journal of
 607 Environmental Health Research*, 26. <https://doi.org/10.1080/09603123.2016.1194383>
- 608 Benson, C. H., Huaming, Z., & Xiaodong, W. (1994). Estimating Hydraulic Conductivity of Compacted Clay
 609 Liners. *Journal of Geotechnical Engineering*, 120(2), 366–387. [https://doi.org/10.1061/\(ASCE\)0733-9410\(1994\)120:2\(366\)](https://doi.org/10.1061/(ASCE)0733-9410(1994)120:2(366))
- 610 Bethke, C. M. (2007). *Geochemical and biogeochemical reaction modeling*. Cambridge University Press.
- 611 Bickle, M. J., Chapman, H. J., Tipper, E., Galy, A., De La Rocha, C. L., & Ahmad, T. (2018). Chemical weathering
 612 outputs from the flood plain of the Ganga. *Geochimica et Cosmochimica Acta*, 225, 146–175.
 613 <https://doi.org/https://doi.org/10.1016/j.gca.2018.01.003>
- 614 Bomer, E. J., Wilson, C. A., Hale, R. P., Hossain, A. N. M., & Rahman, F. M. A. (2020). Surface elevation and
 615 sedimentation dynamics in the Ganges-Brahmaputra tidal delta plain, Bangladesh: Evidence for mangrove
 616 adaptation to human-induced tidal amplification. *CATENA*, 187, 104312.
 617 <https://doi.org/https://doi.org/10.1016/j.catena.2019.104312>
- 618 Borromeo, L., Andò, S., France-İanord, C., Coletti, G., Hahn, A., & Garzanti, E. (2019). Provenance of Bengal Shelf
 619 Sediments: 1. Mineralogy and Geochemistry of Silt. *Minerals*, 9, 640. <https://doi.org/10.3390/min9100640>
- 620 Brammer, H. (2012). *The Physical Geography of Bangladesh*. The University Press Ltd.
- 621 Brammer, Hugh. (2016). *Bangladesh: Landscapes, soil fertility and climate change*. Dhaka, Bangladesh: The
 622 University Press Limited.
- 623 Chen, L.-M., Zhang, G.-L., & Effland, W. R. (2011). Soil characteristic response times and pedogenic thresholds
 624 during the 1000-year evolution of a paddy soil chronosequence. *Soil Science Society of America Journal*,
 625 75(5), 1807–1820. <https://doi.org/10.2136/sssaj2011.0006>
- 626 Chowdhury, M. T. A., Deacon, C. M., Jones, G. D., Imamul Huq, S. M., Williams, P. N., Manzurul Hoque, A. F.
 627 M., et al. (2017). Arsenic in Bangladeshi soils related to physiographic region, paddy management, and micro-
 628 and macro-elemental status. *Science of the Total Environment*. <https://doi.org/10.1016/j.scitotenv.2016.11.191>
- 629 Chowdhury, N. T. (2010). Water management in Bangladesh: an analytical review. *Water Policy*, 12(1), 32.
 630 <https://doi.org/10.2166/wp.2009.112>
- 631 Edmond, J. M. (1992). Himalayan Tectonics, Weathering Processes, and the Strontium Isotope Record in Marine
 632 Limestones. *Science*, 258(5088), 1594 LP – 1597. <https://doi.org/10.1126/science.258.5088.1594>
- 633 Fendorf, S., Michael, H. A., & van Geen, A. (2010). Spatial and temporal variations of groundwater arsenic in South
 634 and Southeast Asia. *Science*, 328(5982), 1123–7. <https://doi.org/10.1126/science.1172974>
- 635 Fry, D. C. (2015). *Characterizing temporal and spatial trends in soil geochemistry on Polder 32, Southwest
 636 Bangladesh*. Vanderbilt University. Retrieved from <http://etd.library.vanderbilt.edu/available/etd-07162015-214524/>
- 637 Gaillardet, J., Dupré, B., Louvat, P., & Allègre, C. J. (1999). Global silicate weathering and CO₂ consumption rates
 638 deduced from the chemistry of large rivers. *Chemical Geology*, 159(1), 3–30.
- 639
- 640
- 641
- 642

- 643 [https://doi.org/https://doi.org/10.1016/S0009-2541\(99\)00031-5](https://doi.org/https://doi.org/10.1016/S0009-2541(99)00031-5)
- 644 Gaillardet, J., Viers, J., & Dupré, B. (2014). Trace Elements in River Waters. In H. D. Holland & K. K. B. T.
645 Turekian (Eds.), *Treatise on Geochemistry* (2nd ed., pp. 195–235). Oxford: Elsevier.
646 <https://doi.org/https://doi.org/10.1016/B978-0-08-095975-7.00507-6>
- 647 Gain, A. (2017). Tidal river management in the south west Ganges-Brahmaputra delta in Bangladesh: Moving
648 towards a transdisciplinary approach? *Environmental Science and Policy*, 75(December 2016), 111–120.
649 <https://doi.org/10.1016/j.envsci.2017.05.020>
- 650 Galy, V., & Eglinton, T. (2011). Protracted storage of biospheric carbon in the Ganges-Brahmaputra basin. *Nature
651 Geoscience*, 4(12), 843–847. Retrieved from <http://dx.doi.org/10.1038/ngeo1293>
- 652 Gandois, L., Perrin, A.-S., & Probst, A. (2011). Impact of nitrogenous fertiliser-induced proton release on cultivated
653 soils with contrasting carbonate contents: A column experiment. *Geochimica et Cosmochimica Acta*, 75(5),
654 1185–1198. <https://doi.org/https://doi.org/10.1016/j.gca.2010.11.025>
- 655 Garzanti, E., Andò, S., France-Lanord, C., Vezzoli, G., Censi, P., Galy, V., & Najman, Y. (2010). Mineralogical and
656 chemical variability of fluvial sediments 1. Bedload sand (Ganga-Brahmaputra, Bangladesh). *Earth and
657 Planetary Science Letters*, 299(3–4), 368–381. <https://doi.org/10.1016/j.epsl.2010.09.017>
- 658 Garzanti, E., Andò, S., France-lanord, C., Censi, P., Vignola, P., Galy, V., & Lupker, M. (2011). Mineralogical and
659 chemical variability of fluvial sediments 2. Suspended-load silt (Ganga-Brahmaputra, Bangladesh). *Earth and
660 Planetary Science Letters*, 302(1–2), 107–120. <https://doi.org/10.1016/j.epsl.2010.11.043>
- 661 Granet, M., Chabaux, F., Stille, P., France-Lanord, C., & Pelt, E. (2007). Time-scales of sedimentary transfer and
662 weathering processes from U-series nuclides: Clues from the Himalayan rivers. *Earth and Planetary Science
663 Letters*, 261(3), 389–406. <https://doi.org/https://doi.org/10.1016/j.epsl.2007.07.012>
- 664 Hale, R. P., Wilson, C. A., & Bomer, E. J. (2019). Seasonal Variability of Forces Controlling Sedimentation in the
665 Sundarbans National Forest, Bangladesh . *Frontiers in Earth Science* . Retrieved from
666 <https://www.frontiersin.org/article/10.3389/feart.2019.00211>
- 667 Horneman, A., van Geen, A., Kent, D. V., Mathe, P. E., Zheng, Y., Dhar, R. K., et al. (2004). Decoupling of As and
668 Fe release to Bangladesh groundwater under reducing conditions. Part I: Evidence from sediment profiles
669 1Associate editor: G. Sposito. *Geochimica et Cosmochimica Acta*, 68(17), 3459–3473.
670 <https://doi.org/https://doi.org/10.1016/j.gca.2004.01.026>
- 671 Hossain, M., Williams, P. N., Mestrot, A., Norton, G. J., Deacon, C. M., & Meharg, A. A. (2012). Spatial
672 heterogeneity and kinetic regulation of arsenic dynamics in mangrove sediments: The Sundarbans,
673 Bangladesh. *Environmental Science and Technology*, 46(16), 8645–8652. <https://doi.org/10.1021/es301328r>
- 674 Khan, M. H. R., Liu, J., Liu, S., Seddique, A. A., Cao, L., & Rahman, A. (2019). Clay mineral compositions in
675 surface sediments of the Ganges-Brahmaputra-Meghna river system of Bengal Basin, Bangladesh. *Marine
676 Geology*, 412(September 2018), 27–36. <https://doi.org/10.1016/j.margeo.2019.03.007>
- 677 Langmuir, D. (1997). *Aqueous Environmental Geochemistry*. Upper Saddle River, NJ: Prentice Hall.
- 678 Lupker, M., Blard, P.-H., Lavé, J., France-Lanord, C., Leanni, L., Puchol, N., et al. (2012). 10Be-derived Himalayan
679 denudation rates and sediment budgets in the Ganga basin. *Earth and Planetary Science Letters*, 333–334,
680 146–156. <https://doi.org/10.1016/j.epsl.2012.04.020>
- 681 Lupker, M., France-Lanord, C., Galy, V., Lavé, J., Gaillardet, J., Gajurel, A. P., et al. (2012). Predominant
682 floodplain over mountain weathering of Himalayan sediments (Ganga basin). *Geochimica et Cosmochimica
683 Acta*, 84, 410–432. <https://doi.org/10.1016/j.gca.2012.02.001>
- 684 Maher, K. (2010). The dependence of chemical weathering rates on fluid residence time. *Earth and Planetary
685 Science Letters*, 294(1), 101–110. <https://doi.org/https://doi.org/10.1016/j.epsl.2010.03.010>
- 686 Milliman, J. D., & Farnsworth, K. L. (2013). *River Discharge to the Coastal Ocean: A Global Synthesis*. Cambridge
687 University Press (Vol. 24). <https://doi.org/10.5670/oceanog.2011.108>
- 688 Naus, F. L., Schot, P., Groen, K., Matin Ahmed, K., & Griffioen, J. (2019). Groundwater salinity variation in
689 Upazila Assasuni (southwestern Bangladesh), as steered by surface clay layer thickness, relative elevation and
690 present-day land use. *Hydrology and Earth System Sciences*, 23(3), 1431–1451. [https://doi.org/10.5194/hess-23-1431-2019](https://doi.org/10.5194/hess-
691 23-1431-2019)
- 692 Pansu, M., & Gautheyrou, J. (2007). *Handbook of soil analysis: mineralogical, organic and inorganic methods*.
693 Springer Science & Business Media.
- 694 Patton, B. (2018). *The Effect of Irrigation Source on Arsenic and Salt Concentrations in Soil in Southwest
695 Bangladesh*. Vanderbilt University.
- 696 R Core Team. (2019). R: A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation
697 for Statistical Computing.
- 698 Raymo, M. E., & Ruddiman, W. F. (1992). Tectonic forcing of late Cenozoic climate. *Nature*, 359(6391), 117–122.

- 699 https://doi.org/10.1038/359117a0
700 Rogers, K. G., Goodbred, S. L., & Mondal, D. R. (2013). Monsoon sedimentation on the “abandoned” tide-
701 influenced Ganges-Brahmaputra Delta plain. *Estuarine, Coastal and Shelf Science*, 131, 297–309.
702 Rudnick, R. L., & Gao, S. (2003). Composition of the continental crust. *Treatise on Geochemistry*, 3, 659.
703 Sarin, M. M., Krishnaswami, S., Dilli, K., Somayajulu, B. L. K., & Moore, W. S. (1989). Major ion chemistry of the
704 Ganga-Brahmaputra river system: Weathering processes and fluxes to the Bay of Bengal. *Geochimica et*
705 *Cosmochimica Acta*, 53(5), 997–1009. [https://doi.org/https://doi.org/10.1016/0016-7037\(89\)90205-6](https://doi.org/https://doi.org/10.1016/0016-7037(89)90205-6)
706 Tan, K. H. (2005). *Soil sampling, preparation, and analysis*. CRC.
707 Tipper, E. T., Bickle, M. J., Galy, A., West, A. J., Pomiès, C., & Chapman, H. J. (2006). The short term climatic
708 sensitivity of carbonate and silicate weathering fluxes: Insight from seasonal variations in river chemistry.
709 *Geochimica et Cosmochimica Acta*, 70(11), 2737–2754.
710 <https://doi.org/https://doi.org/10.1016/j.gca.2006.03.005>
711 West, A. J., Galy, A., & Bickle, M. (2005). Tectonic and climatic controls on silicate weathering. *Earth and*
712 *Planetary Science Letters*, 235(1), 211–228. <https://doi.org/https://doi.org/10.1016/j.epsl.2005.03.020>
713 Wilson, C. A., & Goodbred, S. L. (2015). Construction and maintenance of the Ganges-Brahmaputra-Meghna delta:
714 Linking process, morphology, and stratigraphy. *Annual Review of Marine Science*, 7, 67–88.
715 <https://doi.org/10.1146/annurev-marine-010213-135032>
716 Zhang, J., Wang, X., Zhu, Y., Huang, Z., Yu, Z., Bai, Y., et al. (2019). The influence of heavy metals in road dust on
717 the surface runoff quality: Kinetic, isotherm, and sequential extraction investigations. *Ecotoxicology and*
718 *Environmental Safety*, 176(March), 270–278. <https://doi.org/10.1016/j.ecoenv.2019.03.106>
719

720 Tables

721 **Table 1.** Site locations and characteristics

Site	Latitude	Longitude	District	Thana	Union	Soil region ^a
B2	22.56895	89.58131	Bagerhat	Rampal	Rajnagar	Eac: Mixed Ganges River and Tidal Floodplains, non-saline
B3	22.86543	89.30501	Jessore	Keshabpur	Gaurighona	Dba or Fcb
B8	22.72527	89.19153	Satkhira	Tala	Khalishkhali	Dba: Old Southwestern Ganges River Meander Floodplains
B9	22.98833	89.48716	Jessore	Abhaynagar	Subha Para	Fcb: Old Floodplain Basins, Basin Margins
P32	22.50370	89.46009	Khulna	Dacope	Kamarkhola and Sutarkhali	Ebc: Southwestern Ganges Saline Tidal Floodplain
SB	22.45910	89.46735	Khulna	Dacope	Nalian Range	Ebd: Khulna Sundarbans Saline Tidal Floodplain

722 ^a Soil physiographic region (Brammer, 2016).

723

724

725

726

Table 2. Geometric mean element concentrations (mg/L) in water samples by sample type^a

	RP-11 count	RP-11 geomean	RP-11 geosd	RP-5 count	RP-5 geomean	RP-5 geosd	SB-11 count	SB-11 geomean	SB-11 geosd	TC-11 count	TC-11 geomean	TC-11 geosd	TC-5 count	TC-5 geomean	TC-5 geosd
pH	34	8.25	1.07	14	7.97	1.06	2	7.85	1.01	8	7.7	1.03	5	7.25	1.04
SpC	34	1.02	2.38	14	5.64	2.86	2	1.63	1.51	8	1.73	2.24	5	12.5	6.02
Al	34	0.0297	1.35	14	0.0534	1.46	2	0.0298	1.1	8	0.0315	1.32	5	0.0546	1.64
As	34	0.00081	3.08	14	0.0083	2.58	2	0.00352	1.44	8	0.00327	2.5	5	0.00415	2.53
B	34	0.0834	2.37	14	0.262	2.67	2	0.291	1.06	8	0.169	2.65	5	1.15	3.5
Ba	34	0.521	3.08	14	0.175	2.12	2	1.93	1.59	8	2.21	2.13	5	0.138	3.74
Ca	34	42	1.4	14	146	2.49	2	43	1.24	8	51	1.32	5	145	2.22
Co	34	1.31E-05	2.31	14	0.000438	2.15	2	3.65E-05	1.26	8	3.64E-05	2.33	5	0.000333	4.06
Cr	34	0.000491	3.53	14	0.00146	1.35	2	0.00171	1.46	8	0.00143	2.41	5	0.00202	1.28
Cu	33	0.00143	2.2	14	0.00453	2.05	2	0.00215	1.83	7	0.00237	1.26	4	0.00908	3.04
Fe	32	0.000787	3.11	14	0.0257	2.97	2	0.000346	5.8	7	0.000712	1.89	5	0.00356	2.16
K	34	6.6	2.56	14	11	5.06	2	27.5	1.34	8	15.4	2.97	5	105	7.28
Mg	34	21.8	2.3	14	81.9	3.14	2	56.5	1.55	8	46.2	2.32	5	269	4.95
Mn	34	0.000724	3.47	14	0.11	5.68	2	0.00126	1.91	8	0.00115	2.03	5	0.0421	4.51
Mo	34	0.000139	2.2	14	0.00127	2.73	2	0.000661	1.28	8	0.000482	2.45	5	0.00361	2.51
Na	34	106	3.86	14	954	3.51	2	494	1.59	8	292	4.43	5	2100	9.92
Ni	34	0.000125	3.05	13	0.00425	2.11	2	0.000536	1.21	8	0.000185	9.13	5	0.00311	3.62
P	20	0.00785	2.94	14	0.194	3.75	2	0.00277	5.8	5	0.00663	2.34	5	0.0277	1.44
S	34	6.94	3.61	14	21.3	7.19	2	40.3	1.59	8	20.6	4.47	5	167	7.69
Sb	34	5.26E-05	1.72	14	0.000385	2.35	2	0.000125	1.9	8	0.000136	2.31	5	0.000475	1.36
Se	31	0.000111	8.98	14	0.000656	2.88	2	0.00224	1.75	7	0.00189	3.33	5	0.00116	4.52
Si	34	0.965	3.83	14	4.75	2.71	2	3.15	1.01	8	2.16	1.87	5	1.45	1.39
Sr	34	0.251	1.83	14	1	2.64	2	0.439	1.47	8	0.408	1.85	5	1.93	3.82
V	34	0.000752	2.22	14	0.013	1.78	2	0.00365	1.55	8	0.00265	3.11	5	0.0176	2.84
Zn	34	0.0343	1.21	13	0.0017	1.86	2	0.0351	1.4	8	0.0376	1.24	5	0.00607	4.48
Cl	34	160	4.05	14	1450	3.54	2	756	1.65	8	452	4.63	5	3150	9.91
Br	20	19.2	2.06	14	16.2	2.47	2	29.1	1.89	6	34.1	1.6	5	28.3	5.49
NO ₃	3	2.58	1.49	5	1.07	5.4	0	0	0	1	2.05	0	1	0.138	0
DIC	34	31.6	1.36	14	47.1	1.69	2	34.7	1.58	8	29.9	1.27	5	26.3	1.17
DOC	34	33.1	1.28	10	70.6	1.4	2	34.5	1.56	8	14.4	2.96	3	37.2	1.26

727

^a RP = rice paddy, TC = tidal channel, 11 = November, 5 = May, geomean = geometric mean, geosd = geometric standard deviation

728

Table 3. Geometric mean element concentrations in extracts by sample type, in ppm solid^a

	RP-11 count	RP-11 geomean	RP-11 geosd	RP-5 count	RP-5 geomean	RP-5 geosd	SB-11 (n = 1)	SB-5 (n = 1)	TC-11 count	TC-11 geomean	TC-11 geosd	TC-5 count	TC-5 geomean	TC-5 geosd
pH	19	7.99	1.06	12	7.83	1.04	8.46	7.93	7	8.03	1.03	5	7.76	1.04
SpC	19	435	5.34	12	623	3	280	6010	7	2040	4.06	5	3940	3.6
Al	17	1.83	2.24	9	0.241	6.27	5.25	9.32	7	1.38	4.29	3	2.47	2.1
As	19	0.0284	1.82	13	0.0216	1.57	0.0252	0.0466	7	0.0282	1.88	5	0.0526	1.53
B	4	0.488	2.68	3	0.109	2.95			2	1.99	21.3	0		
Ba	19	0.133	1.75	13	0.133	1.36	0.062	0.231	7	0.158	1.93	5	0.318	2.21
Ca	17	71.6	2.74	12	145	1.57	28		5	224	7.04	2	637	4.64
Co	19	0.0016	4.29	10	0.000764	3.99	0.000497	0.00303	7	0.00185	2.59	5	0.00309	4.81
Cr	17	0.0131	2.89	13	0.00474	4.66	0.00935	0.0152	7	0.0104	2.32	4	0.0176	1.23
Cu	15	0.0537	2.58	12	0.0534	2.65	0.085	0.211	7	0.219	4.88	5	0.283	9.45
Fe	19	0	0	13	0.205	4.62	4.68	4.99	6	1.88	2.91	5	0.812	4.91
K	19	24.9	2.66	13	25.6	1.85	29.1	106	7	72.2	5.22	5	110	1.94
Mg	19	27.6	3.43	13	34.9	2.32	17.7	134	7	48.8	3.47	5	166	4.17
Mn	19	0.108	4.27	13	0.0993	3.31	0.046	2.49	7	0.675	7.6	5	2.73	4.36
Mo	19	0.0127	2.19	13	0.00943	1.75	0.0173	0.0209	7	0.0184	1.87	5	0.0202	1.85
Na	19	313	3.68	13	307	2.8	282	3770	7	1210	3.85	5	1500	5.61
Ni	19	0.017	1.82	13	0.0177	1.62	0.0137	0.018	7	0.0186	1.6	5	0.0303	1.75
P	14	0.265	2.47	11	0.266	2.5			3	0.843	2.08	2	0.472	3.96
S	19	98.8	4.68	13	54.5	2.51	104	391	7	224	2.98	5	581	3.24
Sb	19	0.00686	2.13	13	0.00514	1.85	0.00595	0.00321	7	0.0058	1.99	5	0.00518	1.72
Se	19	0.0175	2.03	12	0.0123	2.61	0.01	0.0493	7	0.0195	2.6	5	0.0533	4.75
Si	15	11.9	1.6	12	14.5	1.48	19.4		4	14.7	1.95	1	24.3	
Sr	18	0.366	2.64	13	0.556	1.58	0.147		5	1.65	8.93	4	1.91	3.45
V	19	0.026	1.86	13	0.0186	1.96	0.0221	0.0649	7	0.0256	2.32	5	0.0466	2.26
Zn	13	0.0331	3.7	11	0.0106	6.62	0.0105	0.76	5	0.41	4.7	4	0.794	1.27
Cl	19	469	3.71	13	462	2.85	425	5810	7	1800	3.77	5	2290	5.55
Br	9	26.4	1.9	3	38.3	5.56		22.6	5	17.9	1.97	4	28.2	2.14
NO ₃	6	8.26	3.39	8	29.5	4.19		2.78	2	2.07	1.02	5	3.03	2.11
DIC	19	43	2.22	13	53.8	1.5	61.7	51.9	7	51	1.61	5	47.5	1.25
DOC	19	244	1.49	13	229	1.42	206	196	7	213	1.58	5	305	1.69

729 ^a RP = rice paddy, TC = tidal channel, 11 = November, 5 = May, geomean = geometric mean, geosd = geometric standard deviation

730

Table 4. Concentrations in solid samples by sample type in ppm^a

	Blank-5	RP-11 count	RP-11 geomean	RP-11 geosd	RP-5 count	RP-5 geomean	RP-5 geosd	SB-11 (n = 1)	SB-5 (n = 1)	TC-11 count	TC-11 geomean	TC-11 geosd	TC-5 count	TC-5 geomean	TC-5 geosd
Al	42.4	18	63800	1.23	12	57000	1.29	77800	32500	7	57100	1.35	6	59000	1.2
As	0.187	18	1.63	1.64	12	1.53	2.16	3.45	0.645	7	1.09	2.39	6	1.66	1.27
Ba	3.4	18	768	1.48	12	917	1.34	492	378	7	638	1.62	6	728	1.65
Ca	297	18	7600	1.7	12	11200	1.91	12300	9580	7	13400	1.38	6	10000	1.43
Co	0.214	18	11	1.55	12	8.19	1.54	12.2	8.45	7	8.56	1.59	6	8.95	1.25
Cr	3.81	18	55.3	1.22	12	47.9	1.48	94.1	18.2	7	39.7	1.83	6	51.5	1.24
Cu	3.65	18	29.4	1.31	12	27.5	1.71	37.5	23.4	7	27.3	1.38	6	28.3	1.46
Fe	75.4	18	30500	1.29	12	27200	1.48	47600	10900	7	25100	1.72	6	28900	1.21
K	203	18	36300	1.16	12	36100	1.25	33400	26700	7	33700	1.27	6	34800	1.28
Mg	58.1	18	10700	1.21	12	10000	1.29	13100	10500	7	11300	1.13	6	10600	1.17
Mn	2.25	18	432	1.4	12	408	1.35	693	417	7	592	1.25	6	455	1.09
Mo	2.37	18	1.96	2.05	12	1.8	1.82	12.5	0.945	7	1.84	2.02	6	1.28	1.88
Na	356	18	7990	1.29	12	8090	1.21	10700	11800	7	9530	1.28	6	9850	1.17
Ni	0.929	18	27.8	1.29	12	26.6	1.63	32.1	20.6	7	23.3	1.54	6	24.3	1.3
P	104	18	367	1.77	12	416	1.6	802	98.4	7	276	2.16	6	422	1.12
S	496	12	101	2.31	7	72.5	3.84	535	311	6	71.8	16.6	5	79.6	3.49
Sb	0.0453	18	0.249	2.63	12	0.356	1.99	0.379	0.0246	7	0.109	4.14	6	0.275	2.2
Se	1.31	5	0.0912	5.42	2	0.209	3.72		1		0.286		3	0.205	4.03
Si	241	18	227000	1.27	12	224000	1.19	330000	73700	7	167000	1.84	6	209000	1.11
Sr	0	18	82.9	1.19	12	84.3	1.19	121	88.7	7	95.8	1.15	6	82.4	1.18
V	5.59	18	73.1	1.48	12	78	1.36	103	21.6	7	40.5	2.55	6	72.7	1.25
Zn	1.3	18	77.8	1.21	12	72.2	1.32	83.4	53.7	7	70	1.21	6	73.5	1.16
H ₂ O ⁺		16	73700	1.18	10	78600	1.25	60000	61000	5	68600	1.23	5	73600	1.15
H ₂ O ⁻		16	23400	1.47	10	25200	2.26	22000	38000	5	19100	1.75	5	29000	2

731 ^a RP = rice paddy, TC = tidal channel, 11 = November, 5 = May, geomean = geometric mean, geosd = geometric standard deviation

732

733

734

Table 5: Measured grain sizes of select samples^a

Sample Name	Median (μm)	Mode (μm)	Volume-Weighted Mean (μm)
BEMS_09_TC_01_Nov	16.37	14.98	28.44
BEMS_08_RP_06M_May	36.29	53.68	46.55
SB_SB_1_May	36.02	57.87	47.97
BEMS_03_RP_07_Nov	21.41	23.15	43.16
BEMS_08_TC_01_May	9.89	6.72	20.55
BEMS_08_RP_06M_Nov	39.67	52.41	60.23
BEMS_08_TC_02_May	22.95	34.62	30.33
BEMS_03_TC_01_May	8.44	6.45	14.52

^a Sample Name given as Site_Sample type_Sample #_Month. TC = tidal channel, RP = rice paddy, SB = Sundarbans

735

736

737

Table 6: Concentrations of oxides in solid samples^a

Oxide	RP mean ^b	St Dev	TC mean	St Dev	Avg. UCC ^c	RP anhydrous mean	StDev	TC anhydrous mean	StDev
SiO ₂	59.4	3.75	57.2	4.46	67.1	67.2	3.84	65.3	3.2
Al ₂ O ₃	14	2.37	14.3	1.61	15.5	15.8	2.64	16.4	2.01
FeO	1.35	0.317	1.58	0.44	5.07	1.54	0.36	1.81	0.514
MgO	5.07	1.14	5.1	0.888	2.50	5.76	1.3	5.85	1.13
CaO	4.62	1.11	5.01	0.732	3.61	5.23	1.24	5.73	0.862
Na ₂ O	2.04	0.36	2.25	0.272	3.29	2.31	0.406	2.58	0.341
K ₂ O	1.46	0.87	1.91	0.869	2.82	1.66	1.02	2.17	0.938
P ₂ O ₅	0.117	0.0368	0.121	0.0314	0.15	0.133	0.0409	0.139	0.036
H ₂ O ⁺	8.78	1.53	8.91	1.52					
H ₂ O ⁻	3.01	1.35	3.59	2.1					
Sum	100		100			100		100	

^a Four samples with anomalously low SiO₂ were not included in these summary statistics. Concentrations of oxides normalized to 100%.^b Mean values are 5% trimmed mean, RP = rice paddy soil, TC = tidal channel sediment^c Rudnick and Gao (2003)

738

739

740

741

742

743

744 Figure Captions

745 **Figure 1.** Map of solid sample collection locations specified as Site_Sample #_Sample type. Water samples were collected from the
746 same locations. The sample types are RP rice paddy, TC tidal channel, and SB Sundarbans. Detailed site maps are given in Figures
747 S5-S8.

748
749 **Figure 2.** Boxplot comparing \log_{10} of specific conductivity SpC in mS/cm of samples classified by group (rice paddy soil RP,
750 Sundarbans SB, and tidal channel sediment TC), month (dry season May = 5, wet season November = 11), and sample type. In-situ
751 solid SpC measurements were not made in May.

752
753 **Figure 3.** Average compositions of all solid, extract, and water samples normalized to average upper continental crust (Rudnick &
754 Gao, 2003). The one standard deviation error bars represent the variability across sites, sample types, and season. Elements are
755 arranged on the x-axis from lowest mobility during weathering and transport on the left to highest mobility on the right (Gaillardet et
756 al., 2014). a) Tidal channel samples, n = 12 extract, 13 solid, and 13 water samples. b) Rice paddy samples, n = 32 extract, 30 solid,
757 and 48 water samples. The red horizontal line represents average upper continental crust.

758
759 **Figure 4.** Powder x-ray diffraction patterns for rice paddy soil sample BEMS_08_RPS_06M_May in red and tidal channel sediment
760 sample BEMS_TC_01_May in blue. The y-axis is observed intensity and is truncated for visualization.

761
762 **Figure 5.** Average compositions of rice paddy soil, Sundarbans soil, and tidal channel sediment in May and November. The horizontal
763 red line represents average upper continental crust (Rudnick and Gao, 2003).

764
765 **Figure 6.** Concentration in rice paddy soil Crp normalized to concentration in tidal channel sediment Ctc by sample collection month,
766 plotted as log values with one sigma error bars.

767
768 **Figure 7.** \log_{10} D values for solid/extract $D^{s/e}$ and solid/water $D^{s/w}$ with one sigma error bars. $D^{s/e}$ is an average of 29 rice paddy soil-
769 extract pairs, 13 tidal channel sediment- extract pairs, and two Sundarbans soil- extract pairs, while $D^{s/w}$ is an average of 16 rice paddy
770 soil-water pairs.

771
772 **Figure 8.** Average compositions of Polder 32 November samples of rice paddy soil and tidal channel sediment. No samples were
773 collected in May. The horizontal red line represents average upper continental crust (Rudnick and Gao, 2003).

774
775 **Figure 9.** Comparison of average \log_{10} concentrations in tidal channel water samples from upstream (4 samples from sites B3, B8, and
776 B9) and downstream locations (4 samples from sites P32, B2, and SB).

780

Figure 1.

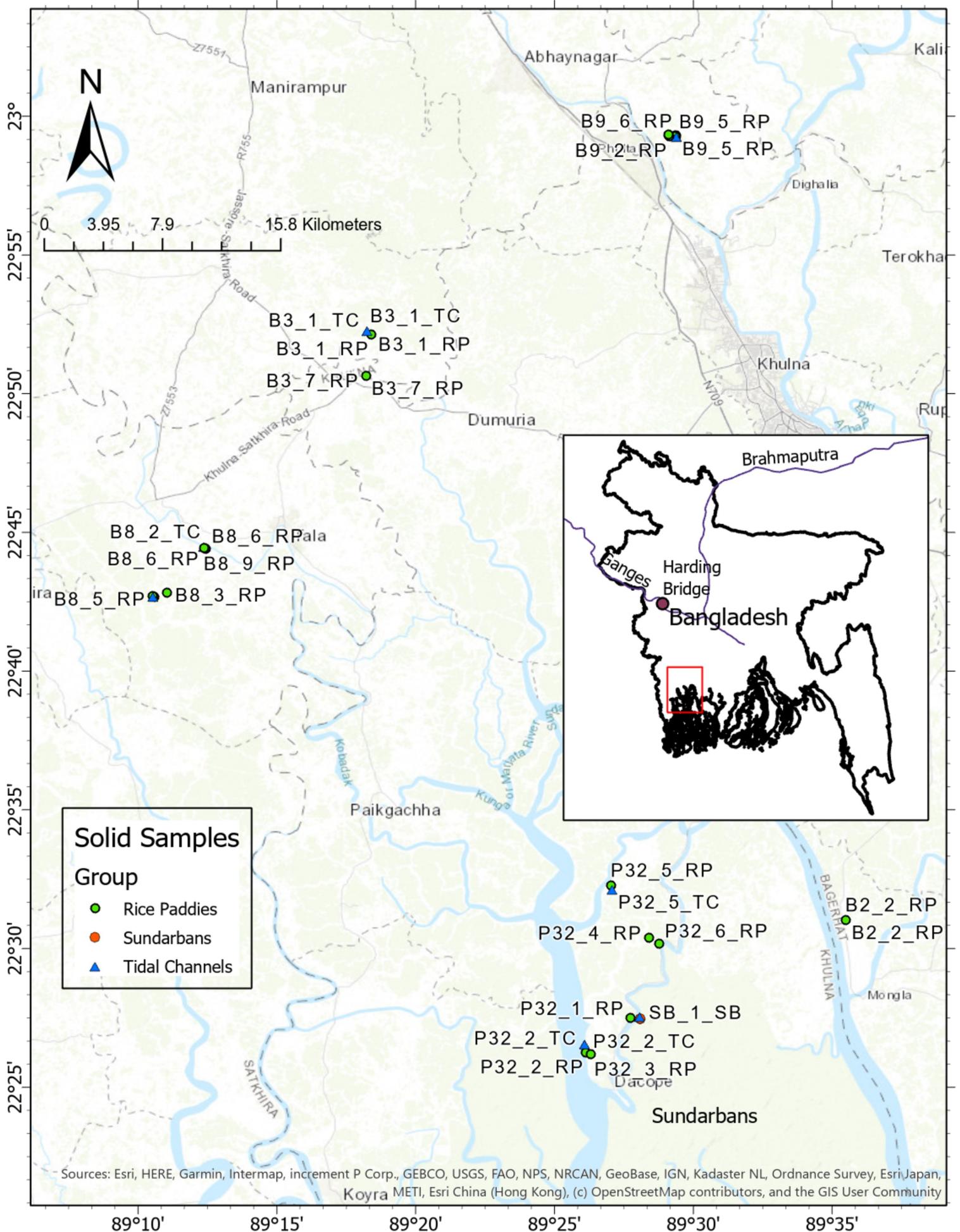


Figure 2.

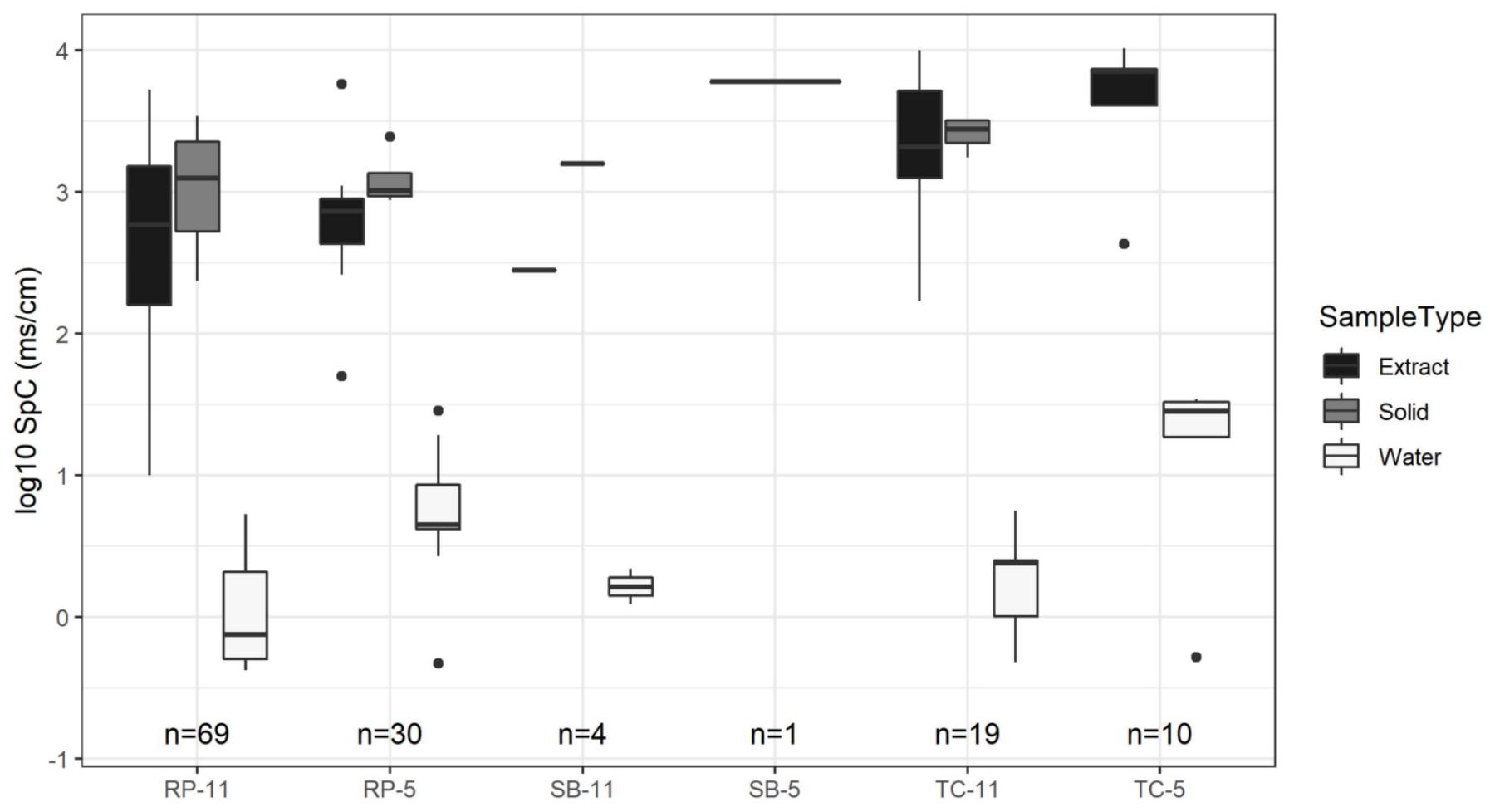
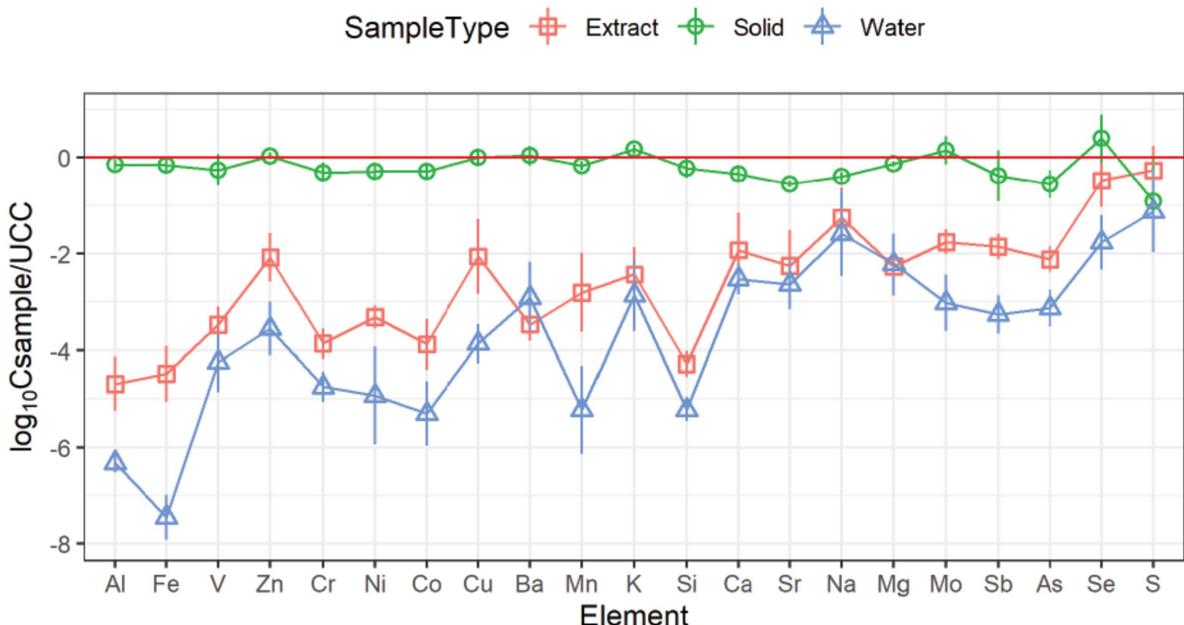


Figure 3.

a)



b)

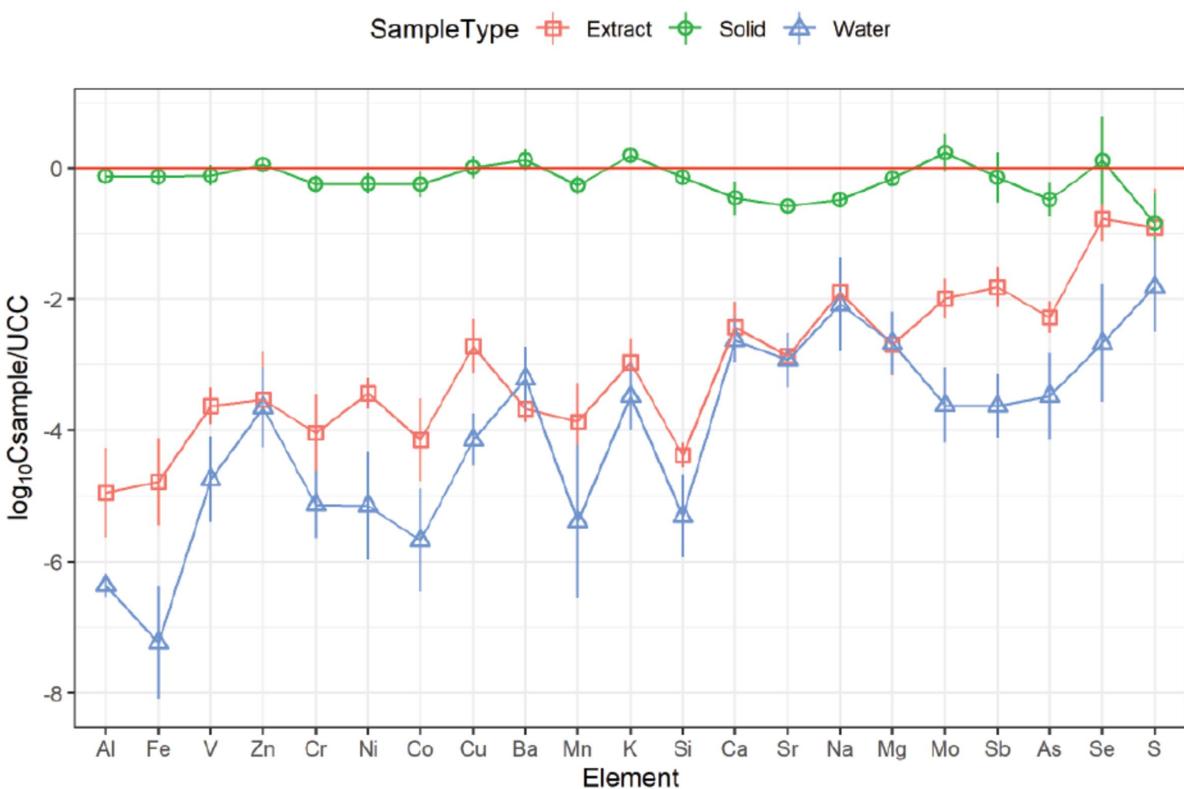


Figure 4.

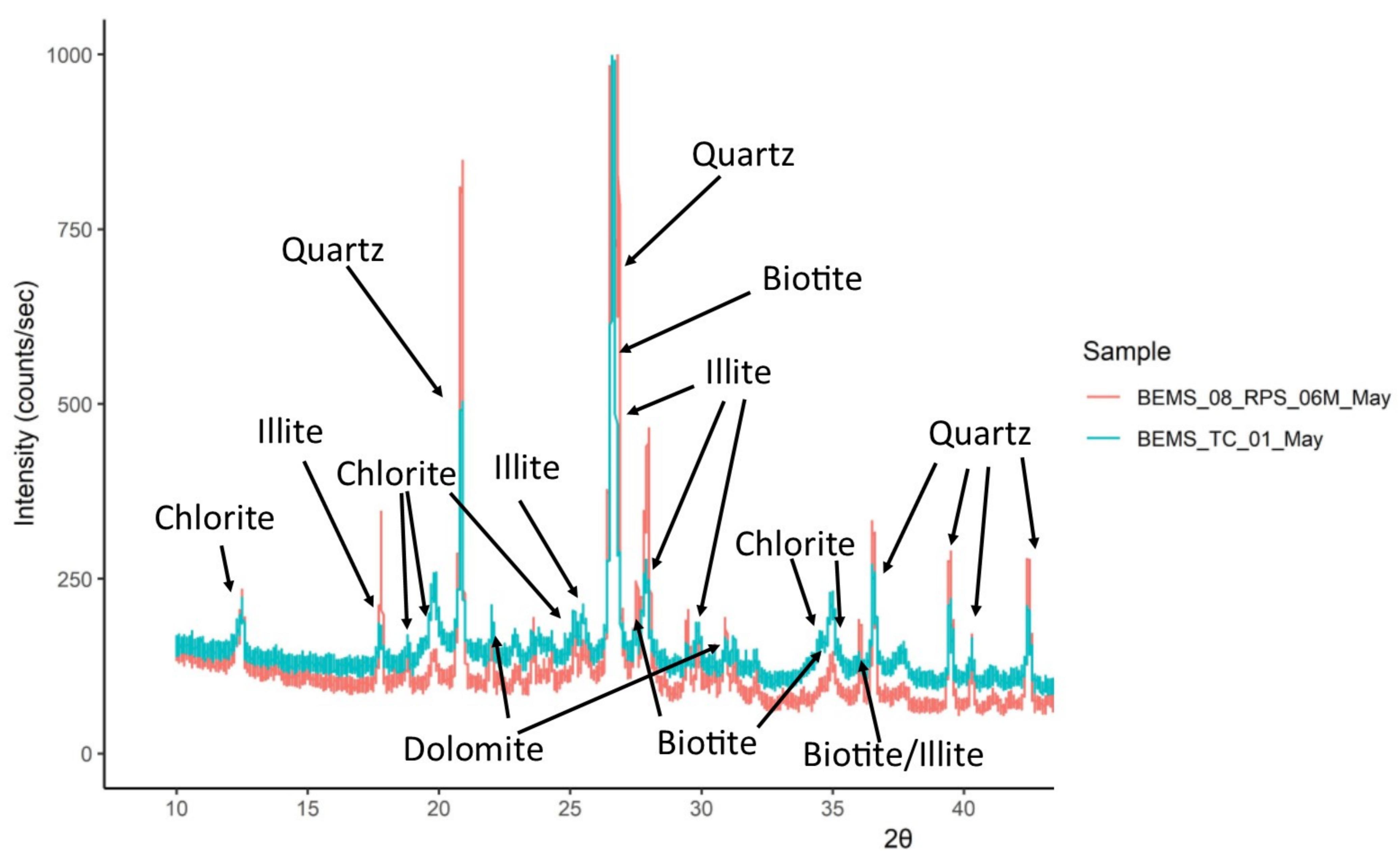


Figure 5.

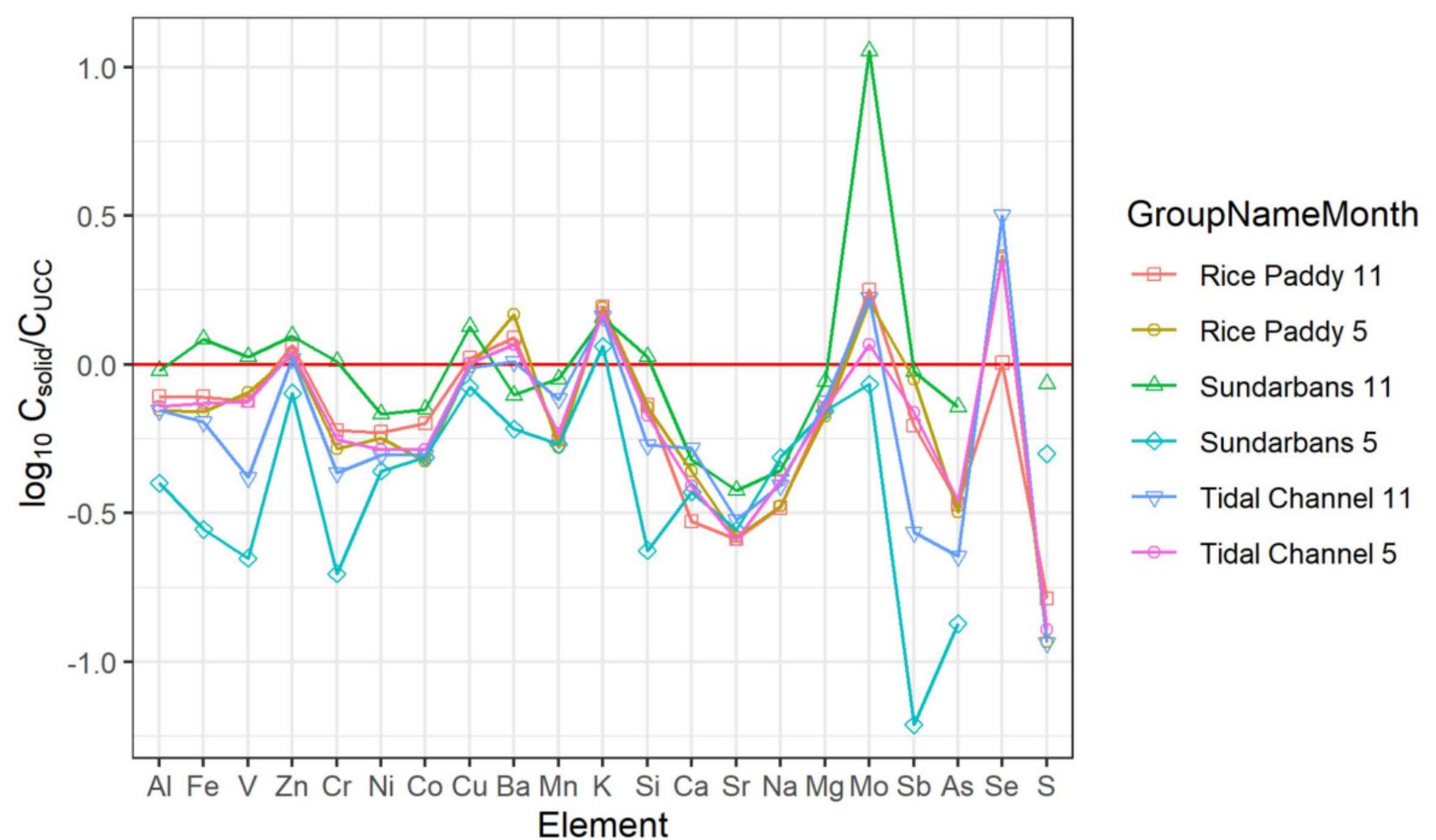


Figure 6.

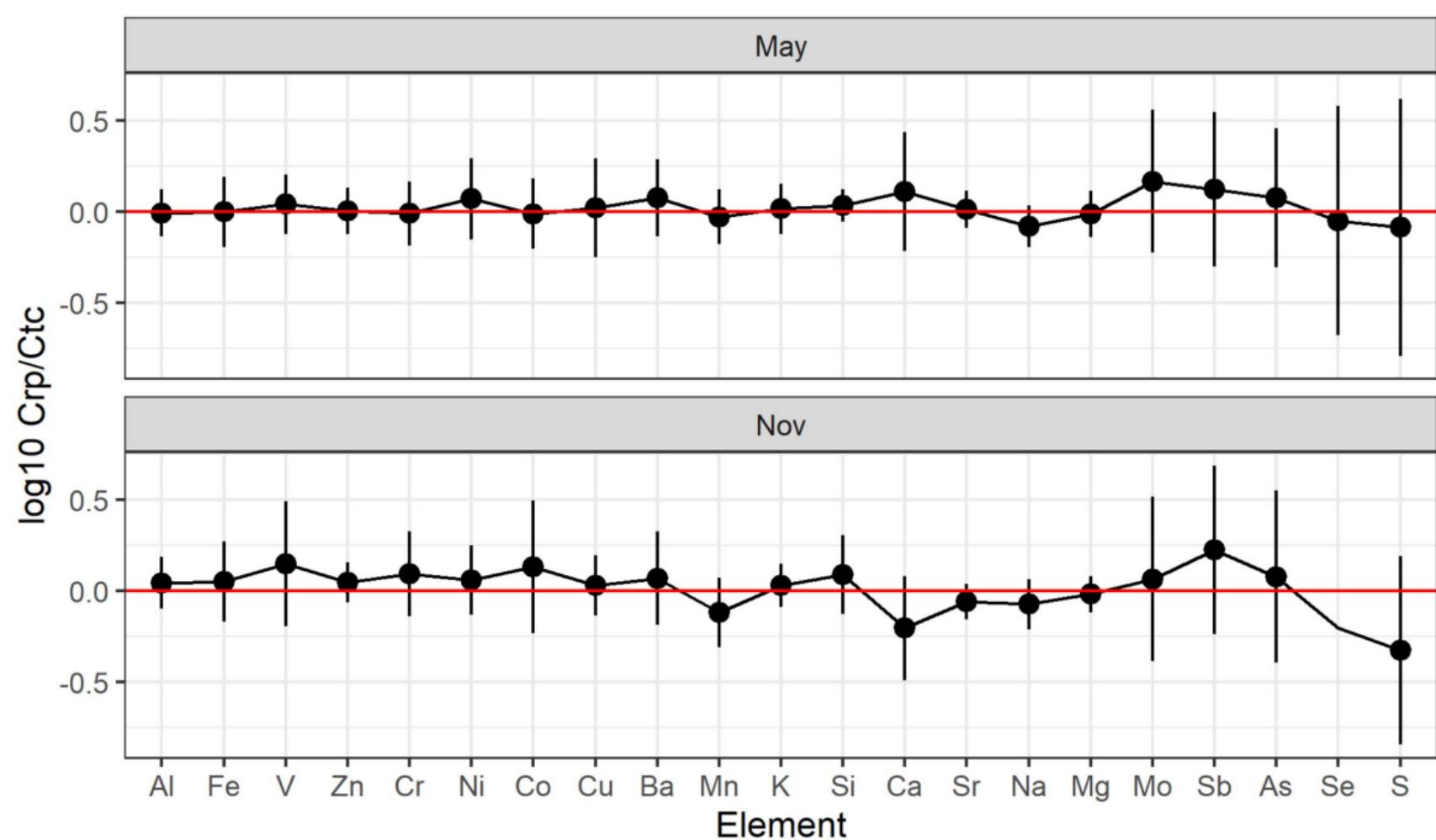


Figure 7.

Dtype  Solid/Extract  Solid/Water

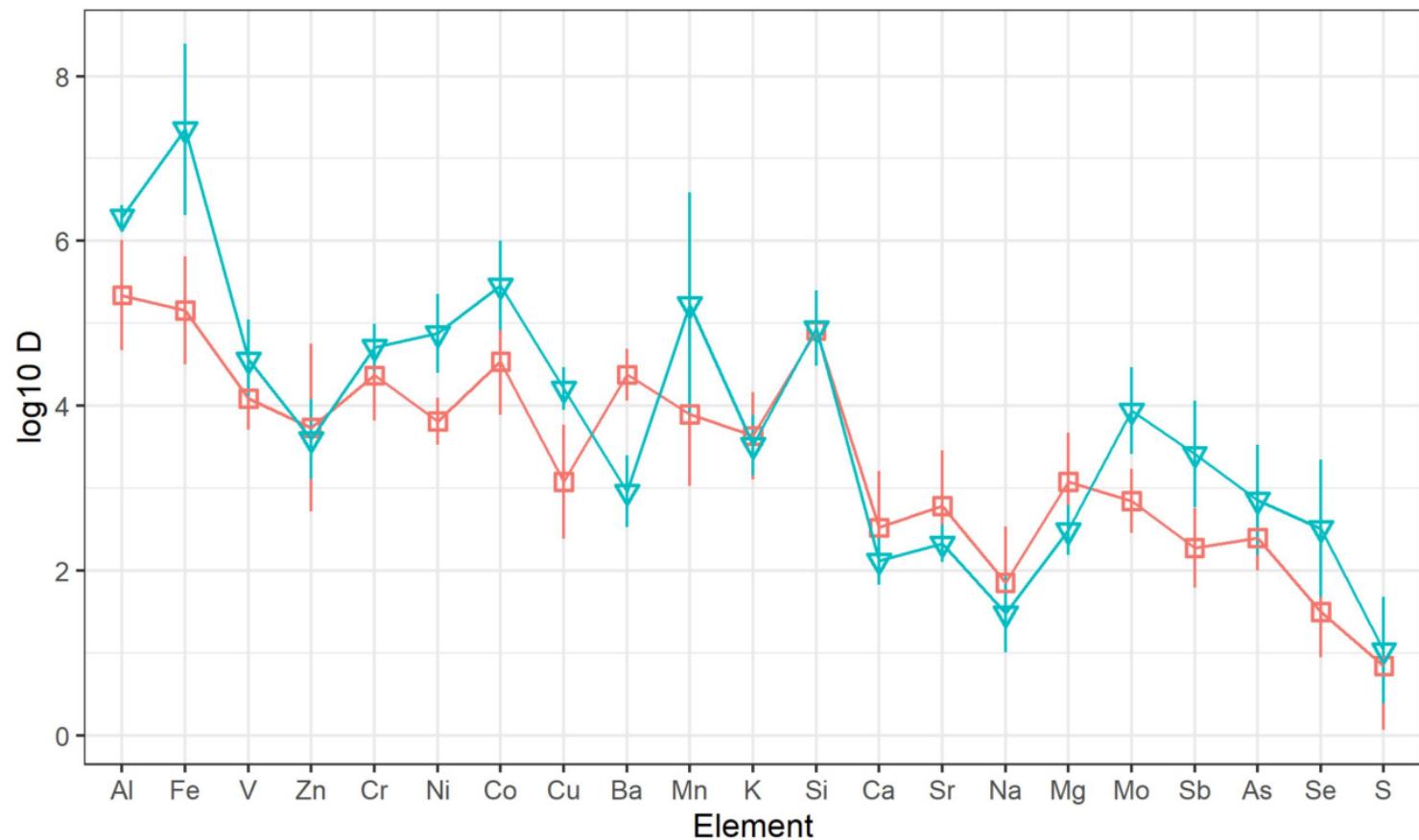


Figure 8.

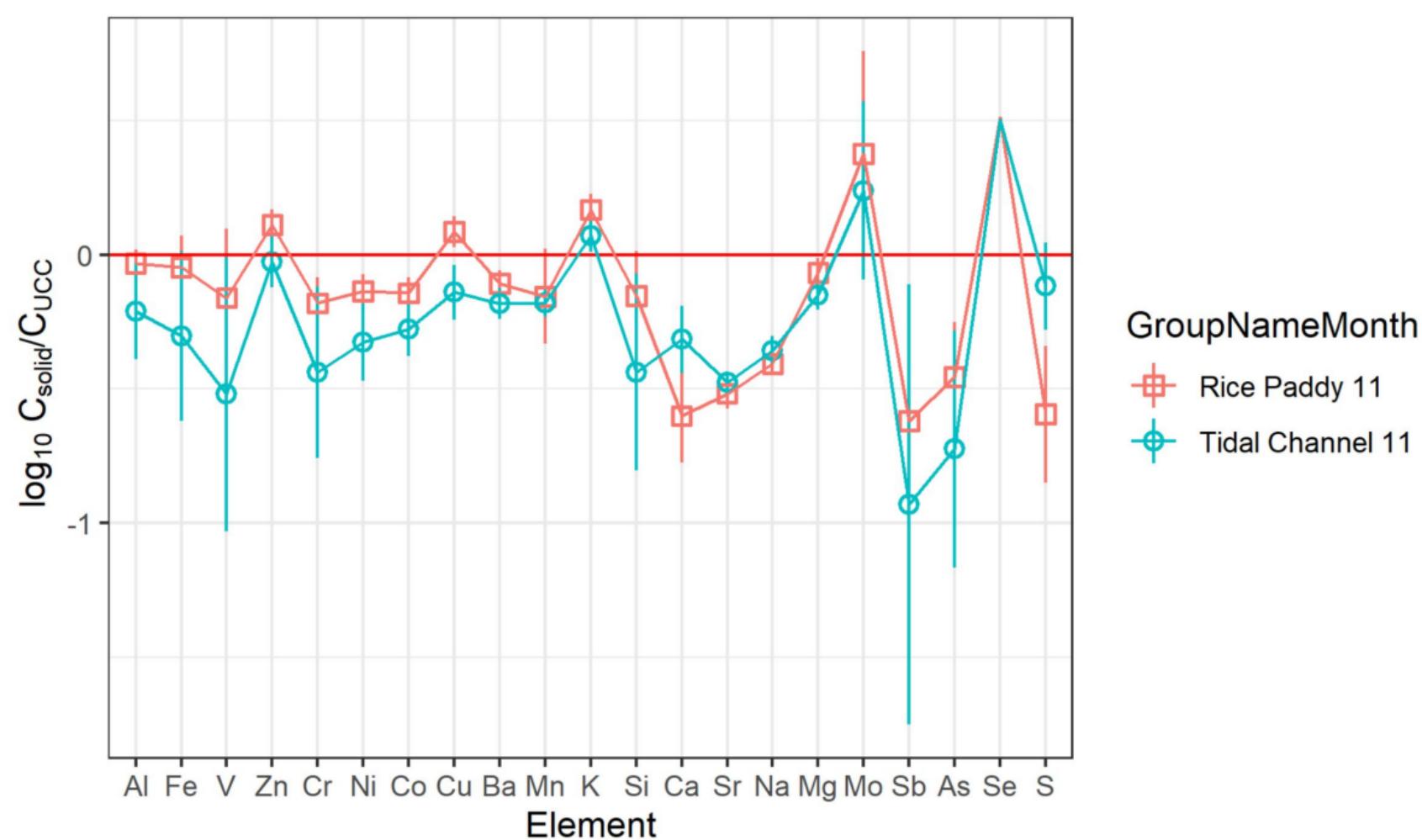


Figure 9.

Location  Downstream  Upstream

