

Shotgun metagenomics of soil invertebrate communities reflects taxonomy, biomass and reference genome properties

Alexandra Schmidt^{1,2,3,7*}, Clément Schneider^{3,4}, Peter Decker^{3,5}, Karin Hohberg^{3,4}, Jörg Römbke⁶, Ricarda Lehmitz^{3,4}, Miklós Bálint^{1,3,8*}

¹ Senckenberg Biodiversity Climate Research Center, Senckenberganlage 25, 60325 Frankfurt am Main, Germany

² Biology Department, J.W. Goethe University, Max-von-Laue-Straße 15, 60438 Frankfurt am Main, Germany

³ Loewe Center for Translational Biodiversity Genomics (LOEWE-TBG), Senckenberganlage 25, 60325 Frankfurt am Main, Germany

⁴ Soil Zoology Department, Senckenberg Museum of Natural History Görlitz, Am Museum 1, 02826 Görlitz, Germany

⁵ Blumenstr. 5, 02826 Görlitz, Germany

⁶ ECT Oekotoxikologie GmbH, Böttgerstraße 2 – 14, 65439 Flörsheim am Main, Germany

⁷ Limnological Institute (Environmental Genomics), University of Konstanz, Mainaustraße 252, 78464 Konstanz, Germany

⁸ Institute for Insect Biotechnology, Justus Liebig University, Heinrich-Buff-Ring 26, 35392 Gießen, Germany

* corresponding authors: alexandra.schmidt@uni-konstanz.de, miklos.balint@senckenberg.de

Running headline: Metagenomics of soil invertebrates

Abstract

1. Metagenomics - shotgun sequencing of all DNA fragments from a community DNA extract - is routinely used to describe the composition, structure and function of microorganism communities. Advances in DNA sequencing and the availability of genome databases increasingly allow the use of shotgun metagenomics on eukaryotic communities. Metagenomics offers major advances in the recovery of biomass relationships in a sample, in comparison to taxonomic marker gene based approaches (metabarcoding). However, little is known about the factors which influence metagenomics data from eukaryotic communities, such as differences among organism groups, the properties of reference genomes and genome assemblies.
2. We evaluated how shotgun metagenomics records composition and biomass in artificial soil invertebrate communities. We generated mock communities of controlled biomass ratios from 28 species from all major soil mesofauna groups: mites, springtails, nematodes, tardigrades and potworms. We shotgun-sequenced these communities and taxonomically assigned them with a database of over 270 soil invertebrate genomes.
3. We recovered 90% of the species, and observed relatively high false positive detection rates. We found strong differences in reads assigned to different taxa, with some groups (e.g. springtails) consistently attracting more hits than others (e.g. enchytraeids). Original biomass could be predicted from read counts after considering these taxon-specific differences. Species with larger genomes, and with more complete assemblies consistently attracted more reads than species with smaller genomes. The GC content of the genome assemblies had no effect on the biomass-read relationships.
4. The results show considerable differences in taxon recovery and taxon specificity of biomass recovery from metagenomic sequence data. The properties of reference genomes and genome assemblies also influence biomass recovery, and they should be considered in metagenomic studies of eukaryotes. We provide a roadmap for investigating factors which influence metagenomics-based eukaryotic community reconstructions. Understanding these factors is timely as accessibility of DNA sequencing, and

45 momentum for reference genomes projects show a future where the taxonomic assignment of DNA from
46 any community sample becomes a reality.

47 **Keywords**

48 shotgun metagenomics, invertebrates, genome completeness, eukaryotes, genome size, biomonitoring,
49 taxonomic bias, species composition

50

51 Introduction

52 Biodiversity research, and particularly the investigation of hard-to-observe ecological communities increasingly
53 relies on DNA- and RNA-based tools ([Taberlet, Bonin, Zinger, & Coissac, 2018](#)). If preconditions are met, e.g.
54 nucleotide sequence databases exist ([Margaryan et al., 2021](#)) with curated taxonomic links ([Schenk, Hohberg,
55 Helder, Ristau, & Traunspurger, 2017](#)), and experimental designs are robust ([Zinger et al., 2019](#)), these
56 approaches can provide much needed data on soil invertebrate diversity.

57 There are two main approaches to the molecular biomonitoring of communities: metabarcoding and
58 metagenomics. Metabarcoding uses high-throughput-sequences of taxonomic marker genes (“barcodes”) which
59 are PCR-amplified from a community DNA extract. Metabarcoding is becoming a standard tool in biodiversity
60 research. Its use is supported by several years of research in distinct organisms groups ([Taberlet et al., 2018](#)),
61 and the availability of barcode databases ([Hebert, Cywinska, Ball, & deWaard, 2003](#); [Nilsson et al., 2019](#)).
62 However, metabarcoding has an important long-known drawback: it relies on the amplification of a marker gene
63 ([Taberlet, Coissac, Pompanon, Brochmann, & Willerslev, 2012](#)). This can result in biases in species recovery
64 from the resulting sequence data, and in taxon-related distortions of the original biomass - sequencing read
65 relationships ([Piñol, Senar, & Symondson, 2019](#)). However, the amplification step solves two important issues:
66 one can effectively target the taxonomic groups of interest (e.g. insects), and avoid others (e.g. microorganisms)
67 and small or rare organisms with low amounts of DNA can still be recorded. Metagenomics randomly sequences
68 all DNA fragments from a community DNA extract, generally without enrichment of certain parts of the genome.
69 It is more quantitative than metabarcoding, since it skips the potentially biased PCR amplification step of
70 taxonomic marker genes ([Bista et al., 2018](#)). A random selection of DNA fragments is sequenced from the DNA
71 extracts, resulting in a less biased representation of the community in the sequence data. The taxonomic
72 assignment of metagenomic sequences needs genome databases, and consequently, metagenomics is more
73 frequently applied on microbial communities, where more complete genomic resources are available ([Parks et
74 al., 2020](#)). There are several approaches to circumvent this limitation, from mitogenomes ([Arribas et al., 2020](#))
75 to shallow genome sequencing ([Bohmann, Mirarab, Bafna, & Gilbert, 2020](#)). As genome sequencing

76 technologies mature, the generation of reference genomes for all eukaryotes receives increasing attention
77 [\(Lewin et al., 2018\)](#). However, the technical issues affecting metagenomics are much less investigated than
78 issues affecting metabarcoding, at least for eukaryotes.

79 Here we evaluate how well metagenomics reflects composition and biomass in artificially composed (mock)
80 communities of soil invertebrates. We use a large collection of soil invertebrate genomes to taxonomically assign
81 metagenomic reads. We investigate the effects of metagenomic classification thresholds on correct and false
82 identification. We evaluate the relationship between biomass and reads, and how this relationship is influenced
83 by taxonomy and by the properties of the genome assemblies used for taxonomic assignments.

84 Material & Methods

85 Mock community construction

86 We constructed mock communities from 28 soil invertebrate species from six major taxonomic groups at the
 87 Senckenberg Museum of Natural History Görlitz. Specimens were either freshly collected and stored in 96%
 88 undenatured ethanol (Collembola, Gamasida, Oribatida), or they came from breeding cultures (Enchytraeidae,
 89 Nematoda, Tardigrada). Four different mock types were designed (Table 1). We varied the total body volume
 90 (the sum of body volumes of all individuals) across the four mock communities. The mocks contained very small
 91 species (Nematoda) with average body volumes per species of $0.10-0.15 \times 10^{-6} \mu\text{m}^3$, up to large species
 92 (Collembola, Enchytraeidae, Gamasina, Oribatida) with average body volumes per species of $44.1-50.8 \times 10^{-6}$
 93 μm^3 (Table 1). We used body volume as a proxy of biomass, and refer to it as biomass throughout the text. In
 94 the first mock, all species were represented with equal biomass. In the second mock, very small to small species
 95 had more biomass (200-500%) compared to medium and large species. In the third mock a part of very small to
 96 small species (7 of 11) had larger biomass (200-400%) than medium to large species. In the fourth mock most
 97 small species had more biomass than large species, but some medium to large species also had high biomass.
 98 All four mock types were replicated three times.

99 Table 1. Composition of mock communities. For species where different developmental stages were available,
 100 individuals of different sizes were used to achieve the necessary biomass [adults + juveniles, e.g.
 101 *Paramacrobiotus richtersi* in mock 1: 4 + 1].

Taxon	mean body length [μm]	body volume [$10^{-6} \mu\text{m}^3$]	Number of individuals			
			mock 1	mock 2	mock 3	mock 4
Tardigrada						
<i>Paramacrobiotus richtersi</i> (Murray, 1911)	700	12.1	4+1	9	0+9	2+5
Nematoda						
<i>Acrobeloides nanus</i> (de Man, 1880)	340	0.15	355	1775	1420	710
<i>Panagrolaimus detritophagus</i> Fuchs, 1930	380	0.10	521	1562	1562	521
<i>Panagrellus redivivus</i> (Linnaeus, 1767)	620	0.28	190	570	380	190

<i>Poikilolaimus oxycerca</i> (de Man, 1895)	930	0.98	54	162	54	162
Collembola						
<i>Sphaeridia pumilis</i> (Krausbauer, 1898)	300	5.7	9	9	37	9
<i>Proisotoma minuta</i> (Tullberg, 1871)	880	11.0	5	4	5	5
<i>Podura aquatica</i> Linnaeus, 1758	560	13.9	4	12	8	8
<i>Desoria trispinata</i> (MacGillivray, 1896)	1090	17.3	3	6	6	6
<i>Isotomurus plumosus</i> Bagnall, 1940	1250	31.0	2	2	2	2
<i>Deuterosminthurus bicinctus</i> (Koch, 1840)	730	36.1	1+1	1+1	1+1	1+1
<i>Sinella curviseta</i> Brook, 1882	1090	44.1	1	1	4	4
<i>Folsomia fimetaria</i> (Linnaeus, 1758)	1400	53.2	1	1	2	3
Oribatida						
<i>Tectocepheus velatus</i> (Michael, 1880)	240	4.8	11	33	11	22
<i>Minunthozetes semirufus</i> (C. L. Koch, 1841)	280	5.6	9	28	19	10
<i>Pantelozetes paolii</i> (Oudemans, 1913)	340	12.9	4	12	4	8
<i>Zygoribatula exilis</i> (Nicolet, 1855)	360	13.7	4	12	8	12
<i>Chamobates voigtsi</i> (Oudemans, 1902)	300	15.9	3	3	7	3
<i>Atropacarus striculus</i> (C. L. Koch, 1835)	440	27.1	2	2	2	2
<i>Liebstadia similis</i> (Michael, 1888)	470	35.5	2	1	5	3
<i>Eupelops occultus</i> (C. L. Koch, 1835)	410	46.5	1	1	1	3
<i>Oribatella quadricornuta</i> (Michael, 1880)	560	50.8	1	1	2	2
Gamasida						
<i>Gaeolaelaps aculeifer</i> (Canestrini, 1883)	700	22.0	2+1	5	2+1	5+6
Enchytraeidae						
<i>Enchytraeus bulbosus</i> Nielsen & Christensen, 1963	4000					
<i>Enchytraeus albidus</i> Henle, 1837	2500					
<i>Enchytraeus luxuriosus</i> Schmelz & Collado, 1999	10500					
<i>Enchytraeus bigeminus</i> Nielsen & Christensen, 1963	6500					
<i>Enchytraeus crypticus</i> Westheide & Graefe, 1992	7500					

fragments

102

103

104

105

We used different formulas for body volume approximation. For Collembola we estimated body volumes as ellipsoid volumes ($V(\mu\text{m}^3) = 1.33 \times \pi \times a \times b \times c \times 10^{-6}$, where a, b, c are axis lengths in μm). For Oribatida, Gamasida, and Enchytraeidae we estimated body volumes as cylinder volumes ($V(\mu\text{m}^3) = \pi \times L \times r^2 \times 10^{-6}$, where

L is height and r is radius), for Tardigrada $V(\mu\text{m}^3) = L \times d^2 \times 0.785 \times 10^{-6}$ (Hallas & Yeates, 1972), and for Nematoda $V(\mu\text{m}^3) = L \times d^2 \times 0.577 \times 10^{-6}$ was used (Andrássy, 1956). Average sizes of mite, nematode, tardigrade and springtail specimens were measured in the populations used for mock community construction, as body sizes can vary among specimens depending on life stage and other factors.

We used the tardigrade culture *Paramacrobiotus richtersi* (Murray, 1911) strain Hohberg-99 and the following cultures of nematodes: *Acrobeloides nanus* (de Man, 1880) strain Hohberg-99, *Panagrolaimus detritophagus* Fuchs, 1930 strain Hohberg-07, *Panagrellus redivivus* (Linnaeus, 1767) strain König-18 and *Poikilolaimus oxycerca* (de Man, 1895) strain Hohberg-01. Thousands of nematode specimens were extracted through sieves and milk filters from the culture plates into tap water. Nematode numbers and mean body volumes within the four stock solutions were then calculated by counting individuals of aliquots and measuring body length and width of 20 specimens per aliquot. After counting, we evaporated the water from each stock solution and added 96% ethanol. For the mock communities we added a calculated part of each of the stock solutions, holding the respective nematode volume, i.e. number x mean body volume. As enchytraeids are large compared to the other invertebrates, we used only body fragments. Tardigrades, collembolans and mites were individually counted into the mock communities. In order to achieve the needed biomass of the respective mock type, differently sized individuals (adults and juveniles) were used. All mock community samples were stored in 2 ml Eppendorf tubes in 96% undenatured ethanol at -20 °C until sequencing.

Laboratory work and sequencing

Before performing the DNA extraction, ethanol was evaporated in a SpeedVac Concentrator Plus (Eppendorf, Hamburg, Germany) to avoid losing material. This is especially important for potentially floating Nematoda and Tardigrada specimens. DNA was extracted with DNeasy Blood and Tissue kit (Qiagen, Hilden, Germany). We included a negative control into the extractions to investigate possible cross-sample contamination. DNA concentration was measured on NanoDrop (Thermo Fisher Scientific, Waltham MA, USA) and Qubit™ (with the dsDNA BR Assay Kit (Thermo Fisher Scientific, Waltham MA, USA). Fragment length was checked on TapeStation 2200 (Agilent Technologies, Santa Clara CA, USA). Libraries were prepared with the NEB Next® Ultra™ DNA Library Prep Kit (New England Biolabs, Ipswich MA, USA) and sequenced on an Illumina NovoSeq

6000 PE150 platform at Novogene (Hong Kong, China). Sequencing depth was 20 gigabase per mock community, and 1 gigabase for the negative control (2 x 150 bp, paired-end).

Bioinformatics & data processing

Sequences were trimmed and quality checked with Autotrim v0.6.1 ([Waldvogel et al., 2018](#)). Autotrim relies on Trimmomatic ([Bolger, Lohse, & Usadel, 2014](#)), FastQC ([Andrews, 2017/2021](#)) and MultiQC ([Ewels, Magnusson, Lundin, & Källner, 2016](#)). It removes Illumina sequencing adapters, performs a quality control of the reads, and combines all information into a single report. Taxonomic classification was performed with Kraken2 v2.0.8 ([Wood, Lu, & Langmead, 2019](#)) against a designated soil invertebrate genome database (GenBank Bioproject PRJNA758215). This database contains short-read assemblies of over 250 species (FigShare doi: 10.6084/m9.figshare.16922890, Supplemental Table 1), including all species used for the mock communities. Before conducting metagenomic classification, the reference genomes were used to build a Kraken2 database with the default k-mer size (k=35). Taxonomic identification of reads was performed on 21 classification thresholds (between 0.0 to 1.0, at 0.05 increments). At each classification threshold, we accounted for possible contamination by extracting the hits of each taxon found in the negative control from the hits of that taxon in every mock community. We plotted correctly identified taxa, false negatives and false positives against the Kraken2 classification threshold, and selected the best-performing assignments for further analysis.

Data analysis

Data analysis was conducted with R v3.6.1 in RStudio (Version 1.2.1335), with data formatted with tidyverse ([Wickham et al., 2019](#)). Graphs and plots were generated by using the package ggplot2 ([Wickham, 2016](#)). Unclassified reads, and classified reads representing less than 0.01 percent of the sample were removed from data. We evaluated false negatives and false positives at all 21 Kraken2 classification thresholds (FigShare doi: 10.6084/m9.figshare.16922890).

We predicted read abundances with the total number of sequences obtained for each mock library with a generalized linear model. Initial independent variables were sequencing success, taxon group (Collembola, Enchytraeidae, Nematoda, Oribatida, Gamasida, Tardigrada), mock species biomasses, genome completeness

157 (measured recovered complete Benchmarking Universal Single-Copy Orthologs, complete BUSCOs [\(Simão,](#)
158 [Waterhouse, Ioannidis, Kriventseva, & Zdobnov, 2015\)](#)), GC content and genome sizes. We estimated genome
159 sizes with ModEst, a new method which performs very well in comparison with flow cytometry measurements
160 (Pfenninger, Schönnenbeck, & Schell, 2021). First we performed a combinatorial model selection with MuMIn
161 (Burnham & Anderson, 2003). The best performing model based of quasi-AIC scores can be written up as hits
162 ~ biomass + taxon_group + missing_buscos + genome_size. The final model was fitted with quasipoisson
163 distribution to account for overdispersion. All predictors were scaled. Genome sizes were log-normalized before
164 scaling. We evaluated the relative importance of the predictors by calculating model-specific variable importance
165 scores in the R package vip [\(Greenwell & Boehmke, 2020\)](#).

166 We evaluated the correspondence between community composition captured by metagenomic reads and
167 original biomass composition with redundancy analyses in vegan [\(Oksanen et al., 2019\)](#). We tested
168 metagenomic hit model statistical significance with an ANOVA-like permutation test for redundancy analysis
169 [\(Legendre & Legendre, 2012\)](#).

170 Results

171 The sequencing resulted in ~69 million paired-end reads on average per mock community replicate, with a
172 standard deviation of ~1.5 million reads. Raw sequencing results are available on the European Nucleotide
173 Archive (accession number: PRJEB45431). About ten million reads were recorded in the negative control. Of
174 the reads passing quality filtering, ~95 million were assigned to taxa at a 0.95 classification threshold (Table 1).
175 The number of correctly classified species remained stable across all classification thresholds (Fig. 1). We
176 retained results at 0.95 as a trade-off for correct and false classifications. Of the 28 species from the mock
177 community, 27 were correctly identified at most classification thresholds (Fig 1). However, the number of false
178 positive classifications strongly decreased at more stringent thresholds, from 181 to 11. The number of false
179 negative classifications remained low, stable and consistent - a single species (an oribatid mite: *Atropacarus*
180 *striculus*) was missed at most classification thresholds. Missing this species was due to the stringency of the
181 bioinformatic sequence processing: the species yielded very few sequencing reads which were then discarded
182 during data filtering.

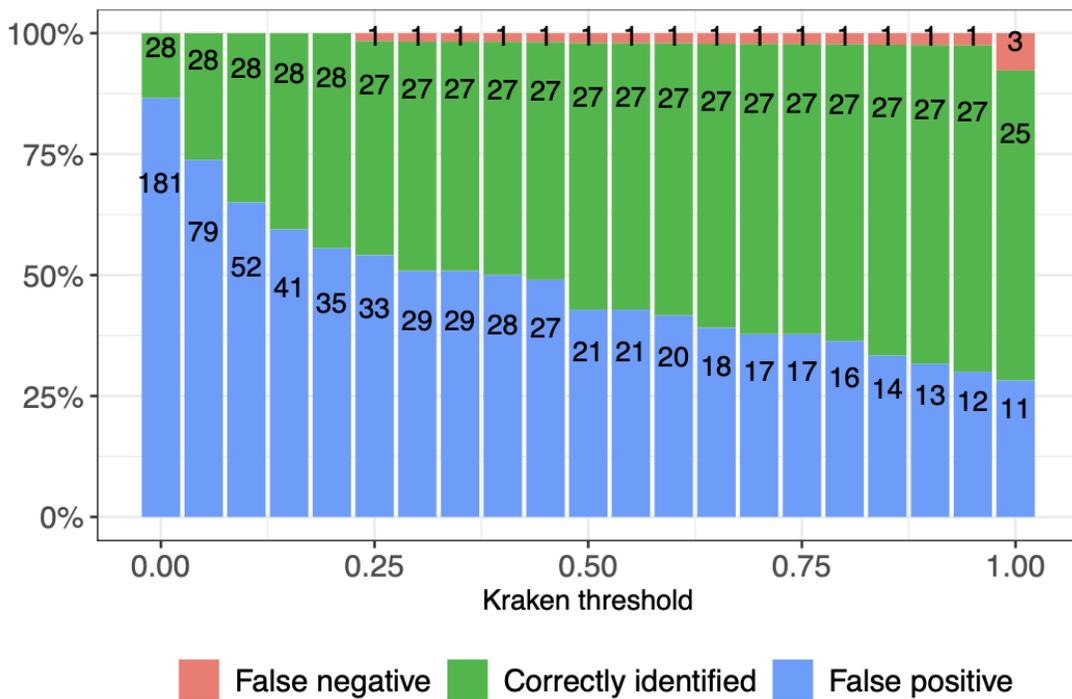


Fig. 1. Species identification success along different Kraken2 classification thresholds. Numbers over bars represent the actual numbers of correctly identified species, and false negative and false positive identifications.

183

184

185

186

187

188

189

190

191

192

193

194

195

196

Some species consistently yielded more reads, regardless of their biomass ratios in the mocks (Fig. 2a). Sequencing depth differences among mock libraries and the GC content of the genomes had little predictive effect on assigned sequencing reads, so they were discarded during model selection. The final model (Fig. 2b, Table 2) showed that metagenomic sequencing success differed across the taxon groups. Compared to reads assigned to Collembola, assignment success to Tardigrada and Nematoda was slightly, but statistically insignificantly lower, while assignment success to Oribatida and Nematoda was statistically significantly lower (Table 2). Biomass of species was positively related to assigned metagenomic reads in all groups. Genome completeness had a statistically significant positive effect on metagenomic read assignment: overall more reads were assigned to taxa with more complete genomes, although this differed across taxon groups. Genome size had a statistically significant positive effect on metagenomic read assignment: more reads were assigned to taxa with larger genomes, regardless of the taxon group. Taxon groups were the most important predictors in the model (Fig. 2c). Replicates of the four mock community types were statistically significantly grouped together in the redundancy analysis ($df = 3, F = 3.863, p < 0.001$, Fig. 3).

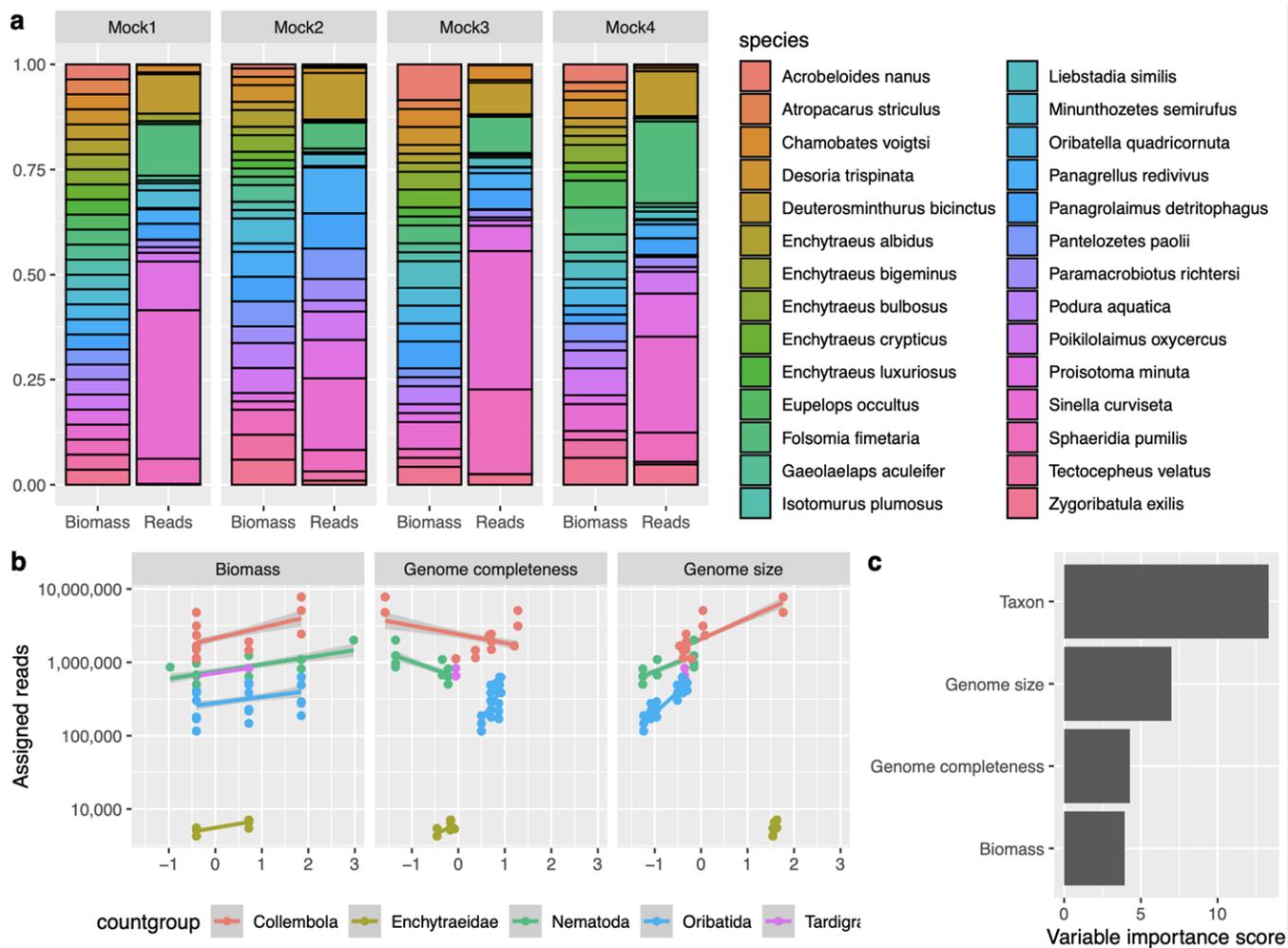


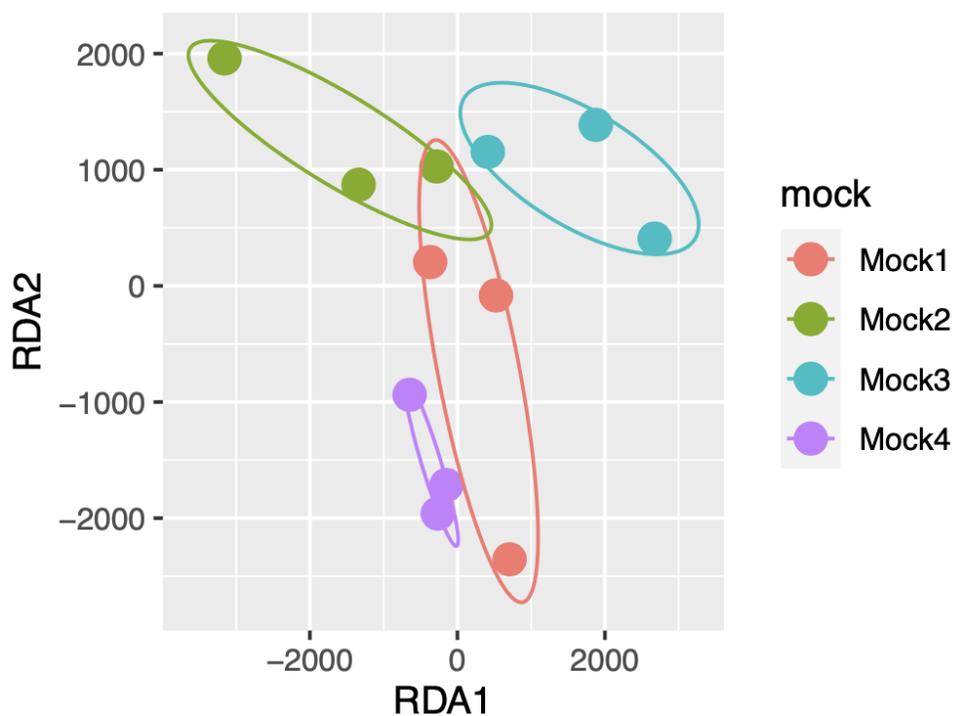
Fig. 2. a) Biomass ratios of taxa and sequencing reads assigned to these taxa in four mock communities. b) GLM-predicted effects of biomass, genome completeness and genome size on taxonomically assigned

197

198 Table 2. Model-predicted biomass, taxon group, genome completeness, genome size effects on assigned
 199 metagenomic read numbers. All predictors were scaled before model fitting. Genome size was log-normalized
 200 before scaling. Collembola served as a model intercept.

	Estimate	Standard error	t	<i>p</i>
(Intercept)	14.286	0.106	134.852	0.000
Biomass	0.215	0.055	3.942	0.000
Enchytraeidae	-7.380	1.787	-4.129	0.000
Nematoda	0.516	0.339	1.521	0.129
Oribatida	-1.374	0.207	-6.632	0.000
Tardigrada	-0.372	0.356	-1.043	0.298
Genome completeness	0.555	0.130	4.283	0.000
Genome size	1.165	0.167	6.991	0.000

201



202 **Fig. 3.** Redundancy analysis ordination of mock community replicates along
 203 the taxonomically assigned metagenomic reads.

202

203

204 Discussion

205 We performed a shotgun metagenomic experiment on soil invertebrate mock communities of known composition.
206 We assigned metagenomic reads to a genome database of soil invertebrates. We investigated how
207 metagenomic reads record the presence of taxa in the mocks, whether read numbers reflect biomass, and how
208 taxonomic, genome, and assembly properties influence biomass - read relationships.

209 Almost all species (27/28) were consistently detected at most classification thresholds. The single false negative
210 species (*A. striculus*) was also detected with very low read numbers, and it was missed only because of stringent
211 quality filtering. The number of false positives was high at low classification thresholds, and rapidly dropped at
212 higher thresholds (Fig. 1). Eleven false positive assignments were retained even at the highest classification
213 threshold. Possible explanations include contamination and bioinformatic issues. Cross-contamination is
214 sometimes observed in mock metagenomes (Bista et al., 2018) but it cannot cause false positives here as all
215 species were present in all mocks. Gut content may also result in the detection of unexpected taxa (Paula et al.,
216 2016). However, most species used in these mocks are not predators. The predatory tardigrade *P. richtersi* was
217 exclusively feeding on a nematode species which was also present in all the mock communities (*A. nanus*). The
218 most likely explanation is related to some aspects of the metagenomic read assignment. The first candidate is
219 the assignment algorithm itself, although comparisons show that Kraken is conservative (Harbert, 2018).
220 Assignment of reads to closely related taxa is an unlikely cause since eight of the 12 false positive species (at
221 0.95 classification threshold) had no genus-level relatives in the mocks. Unmasked repeats might also
222 erroneously attract reads during the assignment. Eukaryotes are rich in low complexity regions, and cross-
223 assignment of these regions might be a considerable source of false positives in all eukaryotic metagenomes
224 (Clarke et al., 2018). The effects of repeat regions in eukaryotic metagenomics assignments should be
225 evaluated, although repeat identification is not trivial, especially for understudied taxa (Clarke et al., 2019).

226 The relationship between sequencing reads and the initial biomass of organisms is a central topic in the DNA-
227 based analysis of community composition. In theory, more shotgun metagenomics reads should be assigned to
228 species which are represented with higher biomass in a sample. However, this relationship might still be

229 influenced by several factors. Here we investigated taxonomic effects, the impact of genome completeness,
230 genome size, and GC content. We found that read counts were most strongly influenced by taxonomy, followed
231 by genome size, genome completeness and biomass (Fig. 2c). We found no statistically significant effects of GC
232 content on read assignment, although this was expected based on previous results with bacterial metagenomes
233 (Browne et al., 2020).

234 There were consistently more reads assigned to some taxonomic groups than to others (Fig. 2b, Table 2). The
235 impact of taxonomy on sequencing reads recovery seems to be systemic, with some species having many reads
236 in all mocks, some species having only few reads (Fig. 2a), and one species was even missed due to the stringent
237 filtering (Fig. 1). Species represented with low biomass in mocks were already found to result in false negatives
238 in metagenomics (Bista 2018), and *A. striculus* was indeed represented with a relatively low biomass in the
239 mocks. However, low biomass alone does not explain the strong taxon effect on read assignment. We suspect
240 that the most important cause for the strong taxon effects is likely caused by differences in DNA yields among
241 different taxa (Sato et al., 2019; Schiebelhut, Abboud, Daglio, Swift, & Dawson, 2017; Tourlousse et al., 2021).
242 Some taxa, e.g. oribatid mites are very hardy, and their cuticles might present obstacles to tissue homogenisation
243 during DNA extractions. Indeed, the single false negative species was an oribatid mite. Cells of different taxa
244 might react differently to extraction (Costea et al., 2017; Morgan, Darling, & Eisen, 2010), with some species
245 consistently yielding lower quality DNA in lower quantities (or no DNA at all) than others (Schiebelhut et al.,
246 2017). However, differential DNA extraction efficiency does not explain why soft-bodied enchytraeids yield
247 considerably less DNA than all other taxa (Fig. 2b). Differences in DNA content relative to body size (or biomass)
248 might be responsible for this: some taxa may contain higher amounts of DNA per unit biomass than others. The
249 association of DNA content with body size can be positive or negative depending on the organism group
250 (Gregory, 2001).

251 Strong taxonomic effects on biomass-read relationships are interesting not only for metagenomic, but also for
252 metabarcoding studies. It is generally assumed that primer mismatch is the most important source of
253 taxonomically biased biomass-read relationships in metabarcoding (Collins et al., 2019; Lamb et al., 2019; Piñol
254 et al., 2019). Our results suggest that taxon-specific differences in DNA extraction efficiency and/or DNA content

255 might also play a role in taxonomic bias. However, recognizing this bias is difficult in metabarcoding: both primer
256 bias, and factors influencing extraction DNA yields are likely phylogenetically conserved. Parallel metabarcoding
257 and metagenomics studies on the same mock communities are necessary to evaluate the relative importance of
258 primer bias versus DNA yield in biomass - read relationships (see e.g. (Bista et al., 2018).

259 Despite considerable taxonomic effects, biomass was a statistically significant predictor of reads (Fig. 2a, Table
260 2). This is in line with other metagenomic mock community studies on multicellular eukaryotes, such as benthic
261 invertebrates (Bista et al., 2018) and pollen samples (Peel et al., 2019). The biomass effect on reads, although
262 considerably smaller than taxon effects (Fig. 2c), was still sufficient to reflect compositional differences among
263 the four mock types (Fig. 3). This confirms the suitability of shotgun metagenomics for a semi-quantitative
264 comparison of soil invertebrate communities.

265 We found that reference genome properties influence taxonomic assignments and read-biomass relationships,
266 and that these need to be considered in metagenomic studies on eukaryotes. We showed that reference
267 genomes size influences metagenomic assignments, with larger genomes attracting more reads than smaller
268 genomes (Fig. 2b). This is known from microbial studies where it was shown that average genome size of a
269 microbial community influences metagenomics results (Beszteri, Temperton, Frickenhaus, & Giovannoni, 2010).
270 We found that genome completeness recorded as BUSCO scores may also influence metagenomic
271 assignments, with more complete genomes attracting more reads. This suggests that reference genome
272 assembly properties should also be considered in metagenomic assignments, even though previous findings
273 show that even low coverage reference genomes can perform well (Sarmashghi, Bohmann, P. Gilbert, Bafna, &
274 Mirarab, 2019). GC content of genomes might also influence metagenomic assignments (Browne et al., 2020),
275 although in our case this effect was limited (Table 2).

276 Our results outline a roadmap for future shotgun metagenomic work on metazoan mock communities. In the wet
277 lab, DNA extraction needs to be optimized and likely adapted to taxa of interest. Differences in DNA content per
278 unit biomass among and within major taxon groups should be evaluated and corrected for. In bioinformatics,
279 assignment algorithms should be evaluated, adapted and developed with eukaryotes in mind. The performance

of distinct genomic regions (i.e. conventional marker genes, mitogenomes, coding regions, ultraconserved regions, repeat elements) should be evaluated, especially with respect to false positive detections. Genome databases will likely remain incomplete for some time. An important direction is to evaluate how incomplete databases (i.e. databases not containing the target species, but congeners or even less related species) perform in taxonomic assignments.

Conclusion

Metagenomics is a promising alternative to metabarcoding also for eukaryotic communities. Although theory suggests that metagenomic reads should well represent biomass relationships in communities, differences among organisms related to DNA extraction efficiency and genome properties have strong influences on the biomass - read relationships. These effects need to be further investigated and quantified in parallel metabarcoding - metagenomic experiments. The effects of taxonomy, genome and assembly properties should be considered in analyses. Generalized linear models provide an excellent opportunity for this. With affordable sequencing and increasingly accessible eukaryotic reference genomes metagenomics is becoming a viable alternative to metabarcoding for describing community composition and structure.

294 Acknowledgements

295 The present study is a result of the LOEWE Centre for Translational Biodiversity Genomics (LOEWE-TBG) and
296 it was funded through the programme "LOEWE – Landes-Offensive zur Entwicklung Wissenschaftlich-
297 ökonomischer Exzellenz" of Hesse's Ministry of Higher Education, Research, and the Arts. MB and CS
298 acknowledge support from the German Research Foundation (DFG project BA 4843/4-1). For the verification of
299 taxon material, we thank Axel Christian and Ulrich Burkhardt.

300 Author contributions

301 MB, KH, RL conceived the ideas and designed methodology. RL, CS, KH, JR provided the animals and ensured
302 correct taxonomic identification. AS, CS, PD, KH, JR, RL collected the data. AS processed the data. AS and MB
303 analysed the data. MB led the writing of the manuscript. All authors contributed critically to the drafts and gave
304 final approval for publication.

305 Data availability

306 Sequence data is available in GenBank (PRJNA758215). R scripts and inputs are available in FigShare (doi:
307 10.6084/m9.figshare.16922890).

References

- Andrássy, I. (1956). Rauminhalts- und Gewichtsbestimmung der Fadenwürmer (Nematoden). *Acta Zoologica*, 2, 1–15.
- Andrews, S. (2021). *S-andrews/FastQC* [Java]. Retrieved from <https://github.com/s-andrews/FastQC> (Original work published 2017)
- Arribas, P., Andújar, C., Moraza, M. L., Linard, B., Emerson, B. C., & Vogler, A. P. (2020). Mitochondrial Metagenomics Reveals the Ancient Origin and Phylodiversity of Soil Mites and Provides a Phylogeny of the Acari. *Molecular Biology and Evolution*, 37(3), 683–694. doi: 10.1093/molbev/msz255
- Beszteri, B., Temperton, B., Frickenhaus, S., & Giovannoni, S. J. (2010). Average genome size: A potential source of bias in comparative metagenomics. *The ISME Journal*, 4(8), 1075–1077. doi: 10.1038/ismej.2010.29
- Bista, I., Carvalho, G. R., Tang, M., Walsh, K., Zhou, X., Hajibabaei, M., ... Creer, S. (2018). Performance of amplicon and shotgun sequencing for accurate biomass estimation in invertebrate community samples. *Molecular Ecology Resources*, 18(5), 1020–1034. doi: <https://doi.org/10.1111/1755-0998.12888>
- Bohmann, K., Mirarab, S., Bafna, V., & Gilbert, M. T. P. (2020). Beyond DNA barcoding: The unrealized potential of genome skim data in sample identification. *Molecular Ecology*, 29(14), 2521–2534. doi: <https://doi.org/10.1111/mec.15507>
- Bolger, A. M., Lohse, M., & Usadel, B. (2014). Trimmomatic: A flexible trimmer for Illumina sequence data. *Bioinformatics*, 30(15), 2114–2120. doi: 10.1093/bioinformatics/btu170
- Browne, P. D., Nielsen, T. K., Kot, W., Aggerholm, A., Gilbert, M. T. P., Puetz, L., ... Hansen, L. H. (2020). GC bias affects genomic and metagenomic reconstructions, underrepresenting GC-poor organisms. *GigaScience*, 9(2). doi: 10.1093/gigascience/giaa008
- Burnham, K. P., & Anderson, D. R. (2003). *Model Selection and Multimodel Inference: A Practical Information-Theoretic Approach*. Springer Science & Business Media.
- Clarke, E. L., Lauder, A. P., Hofstaedter, C. E., Hwang, Y., Fitzgerald, A. S., Imai, I., ... Collman, R. G. (2018). Microbial Lineages in Sarcoidosis. A Metagenomic Analysis Tailored for Low-Microbial Content

3334 Samples. *American Journal of Respiratory and Critical Care Medicine*, 197(2), 225–234. doi:
3335 10.1164/rccm.201705-0891OC

3336 Clarke, E. L., Taylor, L. J., Zhao, C., Connell, A., Lee, J.-J., Fett, B., ... Bittinger, K. (2019). Sunbeam: An
3337 extensible pipeline for analyzing metagenomic sequencing experiments. *Microbiome*, 7(1), 46. doi:
3338 10.1186/s40168-019-0658-x

3339 Collins, R. A., Bakker, J., Wangensteen, O. S., Soto, A. Z., Corrigan, L., Sims, D. W., ... Mariani, S. (2019).
3340 Non-specific amplification compromises environmental DNA metabarcoding with COI. *Methods in
3341 Ecology and Evolution*, 10(11), 1985–2001. doi: <https://doi.org/10.1111/2041-210X.13276>

3342 Costea, P. I., Zeller, G., Sunagawa, S., Pelletier, E., Alberti, A., Levenez, F., ... Bork, P. (2017). Towards
3343 standards for human fecal sample processing in metagenomic studies. *Nature Biotechnology*, 35(11),
3344 1069–1076. doi: 10.1038/nbt.3960

3345 Ewels, P., Magnusson, M., Lundin, S., & Källér, M. (2016). MultiQC: Summarize analysis results for multiple
3346 tools and samples in a single report. *Bioinformatics*, 32(19), 3047–3048. doi:
3347 10.1093/bioinformatics/btw354

3348 Greenwell, B., M., & Boehmke, B., C. (2020). Variable Importance Plots—An Introduction to the vip Package.
3349 *The R Journal*, 12(1), 343. doi: 10.32614/RJ-2020-013

3350 Gregory, T. R. (2001). Coincidence, coevolution, or causation? DNA content, cellsize, and the C-value enigma.
3351 *Biological Reviews*, 76(1), 65–101. doi: 10.1111/j.1469-185X.2000.tb00059.x

3352 Hallas, T. E., & Yeates, G. W. (1972). Tardigrada of the soil and litter of a Danish beech forest. *Pedobiologia*.
3353 Retrieved from
3354 [https://scholar.google.com/scholar_lookup?title=Tardigrada+of+the+soil+and+litter+of+a+Danish+beech
3355 h+forest&author=Hallas%2C+T.E.&publication_year=1972](https://scholar.google.com/scholar_lookup?title=Tardigrada+of+the+soil+and+litter+of+a+Danish+beech+forest&author=Hallas%2C+T.E.&publication_year=1972)

3356 Harbert, R. S. (2018). Algorithms and strategies in short-read shotgun metagenomic reconstruction of plant
3357 communities. *Applications in Plant Sciences*, 6(3), e1034. doi: <https://doi.org/10.1002/aps3.1034>

3358 Hebert, P. D. N., Cywinska, A., Ball, S. L., & deWaard, J. R. (2003). Biological identifications through DNA
3359 barcodes. *Proceedings. Biological Sciences*, 270(1512), 313–321. doi: 10.1098/rspb.2002.2218

3360 Lamb, P. D., Hunter, E., Pinnegar, J. K., Creer, S., Davies, R. G., & Taylor, M. I. (2019). How quantitative is

361 metabarcoding: A meta-analytical approach. *Molecular Ecology*, 28(2), 420–430. doi:
362 <https://doi.org/10.1111/mec.14920>

363 Legendre, P., & Legendre, L. F. J. (2012). *Numerical Ecology*. Elsevier.

364 Lewin, H. A., Robinson, G. E., Kress, W. J., Baker, W. J., Coddington, J., Crandall, K. A., ... Zhang, G. (2018).
365 Earth BioGenome Project: Sequencing life for the future of life. *Proceedings of the National Academy of*
366 *Sciences*, 115(17), 4325–4333. doi: 10.1073/pnas.1720115115

367 Margaryan, A., Noer, C. L., Richter, S. R., Restrup, M. E., Bülow-Hansen, J. L., Leerhøi, F., ... Bohmann, K.
368 (2021). Mitochondrial genomes of Danish vertebrate species generated for the national DNA reference
369 database, DNAMark. *Environmental DNA*, 3(2), 472–480. doi: 10.1002/edn3.138

370 Morgan, J. L., Darling, A. E., & Eisen, J. A. (2010). Metagenomic Sequencing of an In Vitro-Simulated
371 Microbial Community. *PLOS ONE*, 5(4), e10209. doi: 10.1371/journal.pone.0010209

372 Nilsson, R. H., Larsson, K.-H., Taylor, A. F. S., Bengtsson-Palme, J., Jeppesen, T. S., Schigel, D., ...
373 Abarenkov, K. (2019). The UNITE database for molecular identification of fungi: Handling dark taxa and
374 parallel taxonomic classifications. *Nucleic Acids Research*, 47(D1), D259–D264. doi:
375 10.1093/nar/gky1022

376 Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., ... Wagner, H. (2019). *vegan*:
377 *Community Ecology Package*. Retrieved from <https://CRAN.R-project.org/package=vegan>

378 Parks, D. H., Chuvochina, M., Chaumeil, P.-A., Rinke, C., Mussig, A. J., & Hugenholtz, P. (2020). A complete
379 domain-to-species taxonomy for Bacteria and Archaea. *Nature Biotechnology*, 38(9), 1079–1086. doi:
380 10.1038/s41587-020-0501-8

381 Paula, D. P., Linard, B., Crampton-Platt, A., Srivathsan, A., Timmermans, M. J. T. N., Sujii, E. R., ... Vogler, A.
382 P. (2016). Uncovering Trophic Interactions in Arthropod Predators through DNA Shotgun-Sequencing
383 of Gut Contents. *PLOS ONE*, 11(9), e0161841. doi: 10.1371/journal.pone.0161841

384 Peel, N., Dicks, L. V., Clark, M. D., Heavens, D., Percival-Alwyn, L., Cooper, C., ... Yu, D. W. (2019). Semi-
385 quantitative characterisation of mixed pollen samples using MinION sequencing and Reverse
386 Metagenomics (RevMet). *Methods in Ecology and Evolution*, 10(10), 1690–1701. doi: 10.1111/2041-
387 210X.13265

- 388 Pfenninger, M., Schönnenbeck, P., & Schell, T. (2021). *Precise estimation of genome size from NGS data* (p.
389 2021.05.18.444645). doi: 10.1101/2021.05.18.444645
- 390 Piñol, J., Senar, M. A., & Symondson, W. O. C. (2019). The choice of universal primers and the characteristics
391 of the species mixture determine when DNA metabarcoding can be quantitative. *Molecular Ecology*,
392 28(2), 407–419. doi: 10.1111/mec.14776
- 393 Sarmashghi, S., Bohmann, K., P. Gilbert, M. T., Bafna, V., & Mirarab, S. (2019). Skmer: Assembly-free and
394 alignment-free sample identification using genome skims. *Genome Biology*, 20(1), 34. doi:
395 10.1186/s13059-019-1632-4
- 396 Sato, M. P., Ogura, Y., Nakamura, K., Nishida, R., Gotoh, Y., Hayashi, M., ... Hayashi, T. (2019). Comparison
397 of the sequencing bias of currently available library preparation kits for Illumina sequencing of bacterial
398 genomes and metagenomes. *DNA Research*, 26(5), 391–398. doi: 10.1093/dnares/dsz017
- 399 Schenk, J., Hohberg, K., Helder, J., Ristau, K., & Traunspurger, W. (2017). The D3-D5 region of large subunit
400 ribosomal DNA provides good resolution of German limnic and terrestrial nematode communities.
401 *Nematology*, 19(7), 821–837. doi: 10.1163/15685411-00003089
- 402 Schiebelhut, L. M., Abboud, S. S., Daglio, L. E. G., Swift, H. F., & Dawson, M. N. (2017). A comparison of DNA
403 extraction methods for high-throughput DNA analyses. *Molecular Ecology Resources*, 17(4), 721–729.
404 doi: <https://doi.org/10.1111/1755-0998.12620>
- 405 Simão, F. A., Waterhouse, R. M., Ioannidis, P., Kriventseva, E. V., & Zdobnov, E. M. (2015). BUSCO:
406 Assessing genome assembly and annotation completeness with single-copy orthologs. *Bioinformatics*,
407 31(19), 3210–3212. doi: 10.1093/bioinformatics/btv351
- 408 Taberlet, P., Bonin, A., Zinger, L., & Coissac, E. (2018). *Environmental DNA: For Biodiversity Research and*
409 *Monitoring*. Oxford, New York: Oxford University Press.
- 410 Taberlet, P., Coissac, E., Pompanon, F., Brochmann, C., & Willerslev, E. (2012). Towards next-generation
411 biodiversity assessment using DNA metabarcoding. *Molecular Ecology*, 21(8), 2045–2050. doi:
412 10.1111/j.1365-294X.2012.05470.x
- 413 Tourlousse, D. M., Narita, K., Miura, T., Sakamoto, M., Ohashi, A., Shiina, K., ... Terauchi, J. (2021). Validation
414 and standardization of DNA extraction and library construction methods for metagenomics-based

415 human fecal microbiome measurements. *Microbiome*, 9(1), 95. doi: 10.1186/s40168-021-01048-3

416 Waldvogel, A.-M., Wieser, A., Schell, T., Patel, S., Schmidt, H., Hankeln, T., ... Pfenninger, M. (2018). The
417 genomic footprint of climate adaptation in *Chironomus riparius*. *Molecular Ecology*, 27(6), 1439–1456.
418 doi: 10.1111/mec.14543

419 Wickham, H. (2016). *ggplot2: Elegant Graphics for Data Analysis* (2nd ed.). Springer International Publishing.
420 doi: 10.1007/978-3-319-24277-4

421 Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L., François, R., ... Yutani, H. (2019). Welcome to
422 the Tidyverse. *Journal of Open Source Software*, 4(43), 1686. doi: 10.21105/joss.01686

423 Wood, D. E., Lu, J., & Langmead, B. (2019). Improved metagenomic analysis with Kraken 2. *Genome Biology*,
424 20(1), 257. doi: 10.1186/s13059-019-1891-0

425 Zinger, L., Bonin, A., Alsos, I. G., Bálint, M., Bik, H., Boyer, F., ... Taberlet, P. (2019). DNA metabarcoding—
426 Need for robust experimental designs to draw sound ecological conclusions. *Molecular Ecology*, 28(8),
427 1857–1862. doi: 10.1111/mec.15060