Discontinuity in fluvial plastic transport increased by floating vegetation

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Abstract

Understanding plastic mobility in rivers is crucial in estimating plastic emissions into the oceans. Most studies have so far considered fluvial plastic transport as a uniform process, with stream discharge and plastic concentrations as the main variables necessary to quantify plastic transport. Decelerating (e.g.: trapping effects) and accelerating effects (e.g.: increased water flows) on plastic transport are poorly understood, despite growing evidence that such mechanisms affect riverine plastic mobility. In this observation-based study, we explored the roles of an invasive floating plant species (i.e. water hyacinths) as a major disruptor of plastic transport. The different functions of aquatic vegetation in trapping and transporting plastics play a key part in our evolving understanding of how plastic moves in rivers. We collected a one-year dataset on plastic transport, densities and hyacinth abundance in the Saigon river, Vietnam, using both a visual counting method and UAV imagery analysis. We found that hyacinths trap the majority of floating plastic observed ($^{60\%}$), and plastic densities within patches are ten times higher than otherwise found at the river surface. At a monthly and seasonal scale, high hyacinth coverage coincides with peaks in both plastic transport and densities over the dry season (Dec-May) in the Saigon river. We also investigated the large-scale mechanisms governing plant-plastic-water interactions through a conceptual model based on our observations and available literature. Distinguishing total and net plastic transport is crucial to consider fluctuations in freshwater discharge, tidal dynamics and trapping effects caused by the interactions with aquatic vegetation and/or other sinks.

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¹⁰ Key Points:

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11	•	Water hyacinths are major plastic sinks, with plastic densities up to ten times higher
12		than at the river surface.
13	•	Plastic transport, plastic densities and hyacinth abundance are closely linked, with
14		timing and location of accumulation coinciding.
15	•	Hyacinth coverage and plastic densities are affected by fluctuations in river dis-
16		charge which in turn impact plastic transport seasonality.

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17 Abstract

¹⁸ Understanding plastic mobility in rivers is crucial in estimating plastic emissions into ¹⁹ the oceans. Most studies have so far considered fluvial plastic transport as a uniform ²⁰ process, with stream discharge and plastic concentrations as the main variables necessary ²¹ to quantify plastic transport. Decelerating (e.g.: trapping effects) and accelerating effects ²² (e.g.: increased water flows) on plastic transport are poorly understood, despite growing ²³ evidence that such mechanisms affect riverine plastic mobility.

In this observation-based study, we explored the roles of an invasive floating plant species (i.e. water hyacinths) as a major disruptor of plastic transport. The different functions of aquatic vegetation in trapping and transporting plastics play a key part in our evolving understanding of how plastic moves in rivers. We collected a one-year dataset on plastic transport, densities and hyacinth abundance in the Saigon river, Vietnam, using both a visual counting method and UAV imagery analysis.

We found that hyacinths trap the majority of floating plastic observed ($\sim 60\%$), and plastic densities within patches are ten times higher than otherwise found at the river surface. At a monthly and seasonal scale, high hyacinth coverage coincides with peaks in both plastic transport and densities over the dry season (Dec-May) in the Saigon river.

We also investigated the large-scale mechanisms governing plant-plastic-water interactions through a conceptual model based on our observations and available literature. Distinguishing total and net plastic transport is crucial to consider fluctuations in freshwater discharge, tidal dynamics and trapping effects caused by the interactions with aquatic vegetation and/or other sinks.

39 1 Introduction

Plastic pollution poses a series of threats to global ecosystems, including aquatic systems 40 such as rivers. High levels of plastic pollution in rivers can reduce availability of potable 41 freshwater, cause damage to urban infrastructure, and potentially harm the local fauna 42 (van Emmerik & Schwarz, 2020). Rivers are considered the main pathways for land-based 43 plastic emissions into the oceans (Meijer et al., 2021). In addition, rivers can also retain 44 plastics for decades, if not longer (Tramoy et al., 2020). Understanding plastic mobility in 45 rivers is therefore crucial for risk assessments for riverine ecosystems under variable plastic 46 concentrations, and for accurate estimations of emissions into the oceans. 47

Rivers have long been considered as simple conduits for plastic transport to the sea. 48 Many studies portrayed the plastic journey in rivers to be a continuous trajectory of particles 49 through a uniform medium that offers little to no resistance to its final export into coastal 50 waters. As a result, plastic transport in rivers is often quantified as a direct function of 51 plastic concentrations in the water and river discharge (Schmidt et al., 2017; van Emmerik 52 et al., 2018; Haberstroh et al., 2021). However, recent scientific advances have shed light on 53 the discontinuous dynamics at play in fluvial plastic transport; at both temporal and spatial 54 scales. Temporally, plastic transport rates have been observed to follow seasonal patterns 55 and transport in various rivers (van Emmerik et al., 2019; van Emmerik, de Lange, et al., 56 2022), at times linked to seasonal variation in freshwater discharge. In addition, extreme 57 discharge events such as floods lead to disproportionally increased plastic transport rates 58 (Hurley et al., 2018; Roebroek et al., 2021; van Emmerik, Frings, et al., 2022). Spatially, 59 changes in river shape such as meander bends and the presence or absence of physical 60 barriers can lead to varying trapping rates, which affects plastic propagation in the water 61 (Newbound, 2021). Physical traps or barriers include infrastructure such as damns, groynes, 62 bridges and weirs, as well as bank and aquatic vegetation. These impediments can physically 63 retain plastic items temporarily or even permanently (Cesarini & Scalici, 2022; Schreyers, 64 van Emmerik, Luan Nguyen, Castrop, et al., 2021; Skalska et al., 2020). In addition, 65

varying plastic concentrations caused by human behaviours along the river (plastic leakage
and removal) contributes to spatially varying plastic transport rates. These discontinuities
likely lead to accelerating or decelerating effects of plastic distribution and propagation in
the water, similarly to what is observed for other floating debris such as wood (Wohl, 2017;
Wohl & Scott, 2017). As such, these discontinuities challenge the common assumption of a
uniform and unidirectional effect of river discharge on plastic mobility.

Aquatic vegetation can disrupt plastic mobility in rivers physically, spatially and tem-72 porally, and could therefore generate discontinuous effects in fluvial plastic transport. Veg-73 74 etation can trap plastic items, therefore leading to deposition and transport mechanisms that are affected by water-plant-plastic interactions (*physical discontinuity*). Vegetation 75 coverage varies due to the seasonal cycle, which, in turn, leads to higher or lower plastic 76 retention rates depending on the period of the year considered (temporal discontinuity). 77 Small scale variations in vegetation abundance along and/or across a given river might also 78 alter both plastic transport and deposition rates (*spatial discontinuity*). Here, we explore 79 the discontinuous nature of fluvial plastic transport by focusing on the role of an aquatic 80 vegetation species (e.g.: water hyacinths, *Eichhornia crassipes*) in trapping plastics in the 81 Saigon River. Hyacinths function as a major aggregator of floating macroplastics in trop-82 ical rivers and can, therefore, act as a dominant control factor of fluvial plastic transport 83 (Schreyers et al., 2021a; Schreyers et al., 2021b). These invasive aquatic species are now 84 present in most tropical lakes and rivers worldwide (CABI, 2020; Thamaga & Dube, 2019), 85 and their coverage of water surfaces can double in within one to two weeks due to their 86 rapid growth rate (Ouma et al., 2005). As a surface plant species, hyacinth float in patches 87 of varying sizes and densities. Their drift patterns are passive, and spatial distributions 88 are influenced by factors such as currents and wind. In low flow conditions, hyacinths can 89 rapidly blanket a large portion of the waterway. Kleinschroth et al. (Kleinschroth et al., 90 2021) found that for small reservoirs, peaks in hyacinth coverage often exceeded 80% of 91 the total reservoir area. Conversely, in more active systems like rivers, hyacinth coverage 92 tends to be lower due to the transport of the plants with water flow, but can still reach 93 up to 25% of the river surface (Janssens et al., 2022). Previous field-based studies have 94 successfully shown that hyacinths play a crucial role in fluvial plastic transport, however, 95 these observations were conducted over a short measurement period (6 weeks) and at only 96 one location. This study provides a much-needed more comprehensive understanding of how 97 hyacinth abundance alters fluvial plastic transport over both time and space. 98

For the present study, we monitored hyacinth coverage, plastic transport and plastic 99 densities in the Saigon river, Vietnam, over one year. The Saigon river has one of the 100 highest plastic transport rates in the world and is severely impacted by hyacinths invasion 101 (van Calcar & van Emmerik, 2019; Janssens et al., 2022). We hypothesize that hyacinths 102 function as a major temporary sink for riverine plastics and that therefore temporal peaks 103 and spatial accumulation zones in hyacinth coverage generally coincide with high plastic 104 loads. We first established the overall role of hyacinths as temporary traps for plastic items 105 (section 3.1). We then investigated the evolution of the measured metrics (e.g.: hyacinth 106 coverage, plastic transport and densities) at various temporal scales (seasonally, monthly and 107 daily) to characterize synchronous or asynchronous trends in transport and accumulation 108 (section 3.2). In addition, we analyzed how these variables are spatially distributed in the 109 river system, between upstream and downstream locations along the river and across the 110 river channel (section 3.3). The first part of this study focuses on quantifying hyacinth's role 111 as a temporary and mobile sink of floating plastic based on our field observations (section 3. 112 Results and Discussion). In the second part, we further expand on the interactions between 113 plastic-plant-water at a system scale (section 4. Synthesis and Conceptual model). We first 114 summarize our main findings which identified different modes of plastic transport in the river 115 in relation to hyacinth coverage (section 4.1). We present a conceptual model based on these 116 observational findings and our broader understanding of the fluvial system investigated, to 117 explain spatio-temporal variations in plastic transport (section 4.2). We thus synthesize 118 the discontinuous effects induced by hyacinth abundance on plastic transport (section 4.3) 119

and finally identify next steps in future research effort that seek to understand large-scale plastic transport and deposition processes in fluvial systems (section 4.4). The outcomes of this study are useful for scientists seeking to understand large-scale fluvial plastic transport and deposition mechanisms. In addition, river plastic monitoring and reduction strategies might seek to opportunistically use (temporary) sinks because of their role in aggregating large quantities in floating plastics.

¹²⁶ 2 Data and Methods

127 **2.1 Study area**

We measured plastic transport, hyacinth abundance and plastic densities between De-128 cember 12, 2020 and January 15, 2022 at the Saigon river, Vietnam (Fig. 1 and Table 1). 129 The Saigon river originates in Cambodia and flows into the Dau Tieng reservoir, approx-130 imately 120 km north from Ho Chi Minh City (Nguyen et al., 2020). The river crosses 131 agricultural areas of paddy rice and rubber plantation before entering the city. South of the 132 city, the Saigon river confluences with the Dong Nai river. There, the Dong-Nai-Saigon river 133 system branches into several channels that meanders in the Can Gio mangrove forest before 134 entering the East Sea (Dijksma et al., 2010). The Saigon river is subject to asymmetrical 135 semi-diurnal tidal cycle. Because of the tidal influence, the net river discharge is considered 136 relatively low and subject to seasonal variations between the dry and wet seasons (monthly 137 averages vary between -80 and $320 \text{ m}^3/\text{s}$) (Camenen et al., 2021). In addition, the Saigon 138 river is considered one of the most plastic polluted rivers worldwide, with transport rates 139 within the order of 10^4 items/hour (van Calcar & van Emmerik, 2019). Hyacinth invasions 140 are also particularly severe in this river, with peak coverage reaching up to 14% of the river 141 surface (Janssens et al., 2022). 142

This study focuses on floating macroplastic (>0.5 cm of size) density and transport, 143 hereafter referred to as plastic. Plastic transport was measured at two locations in Ho Chi 144 Minh City (Fig. 1). The first site (L1) is located north of the city (10.89025, 106.69209) 145 and the second (L2) in its southern part (latitude: 10.785984; longitude: 106.718332). The 146 two monitored sites approximately 30 km apart. At Ho Chi Minh City, the Saigon river 147 progresses from north to south, therefore enabling to compare upstream and downstream 148 plastic transport values within the urban area. Plastic transport was measured using the 149 visual counting method for floating bridges from bridges (section 2.2), and hyacinth abun-150 dance and plastic density were measured using Unmanned Aerial Vehicle (UAV) imagery 151 analysis (section 2.3 and 2.4). Flying at the downstream site was deemed unfeasible for long-152 term monitoring, due to the proximity of a military site. For this reason, UAV surveys were 153 only conducted at L1. UAV images were taken across the river channel, with a frequency of 154 one to four flights per measurement day. Each flight consisted of two overpasses across the 155 Saigon river, with a range of 41 to 65 images taken per flight. UAV surveys were carried 156 at a constant elevation of approximately 10 m above the water level. More information 157 on the UAV surveys is available in Supporting Information (Extended Methods). Table 2 158 summarizes the measurement frequencies per month at each location. Data gaps are no-159 ticeable for certain months: no data could be collected for any of the variables investigated 160 in August and September 2021. Due to the COVID-19 pandemic, a strict confinement was 161 mandated in Ho Chi Minh City, thus not allowing observers to leave their houses. A larger 162 data gap is noticeable for hyacinth abundance and plastic densities, with no measurements 163 conducted in April, July and October 2021. The gap during the month of April was due 164 to the unavailability of the observer conducting the UAV flights. The missing data from 165 July and October 2021 was also caused by COVID-19 restrictions. In those months, the 166 government did not allow inhabitants to cross the border between two different provinces, 167 thus not enabling access to the UAV flying site at L1 (a few hundred meters upstream of 168 where the visual counting measurements were conducted). 169

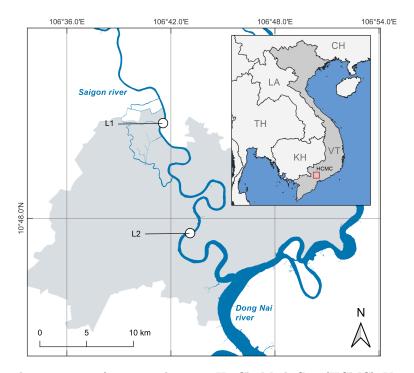


Figure 1: Localization map of monitored sites in Ho Chi Minh City (HCMC), Vietnam and measurement frequency at each location.

Table 1: Measurement frequency at each location. Total refers to the total number of UAV images analyzed in the case of hyacinth abundance and plastic densities. For plastic transport, it refers to the total number of observations, with one observation corresponding to a measurement per observation segment.

	Measurement locations								
	L	1	L2						
	Total	Daily	Total	Daily					
Plastic transport	900	49	1,272	51					
Hyacinth abundance	3,544	29	N/A	N/A					
Plastic densities	2,360	29	N/A	N/A					

Table 2: UAV images and plastic transport measurement frequency per month. The values here refer to the total number of UAV images for hyacinth abundance and plastic density. For plastic transport, the reported values correspond to the total number of observations. Blank cells indicate that no observations were conducted for that period.

	Number of measurements by month													
	Dec 20	Jan 21	Feb 21	Mar 21	Apr 21	May 21	Jun 21	Jul 21	Aug 21	Sep 21	Oct 21	Nov 21	Dec 21	Jan 22
Plastic transport (L1)	54	108	72	126	126	90	54	36			18	90	90	36
Plastic transport (L2)	84	144	83	168	168	120	72	46			89	110	110	44
Hyacinth abundance (L1)	142	536	141	935		407	186					550	363	284
Plastic densities (L1)	105	388	108	391		376	95					435	192	274

¹⁷⁰ 2.2 Floating plastic transport

Plastic transport were estimated using the visual counting method, developed by (González-171 Fernández & Hanke, 2017) and now widely used in observational studies on macroplastic 172 transport (González-Fernández et al., 2021; van Calcar & van Emmerik, 2019). All floating 173 macroplastic and macrolitter items (>0.5 cm) floating at the river surface were counted 174 during a determined time frame at each observation segment. Several observation segments 175 were determined per measurement location, to account for the spatial variability in plas-176 tic transport across the river width (van Emmerik et al., 2018). The number of segments 177 178 depends on the river width of the measurement location. Nine observation segments were selected at L1 (upstream site, river width of 200 m) and twelve at L2 (downstream site, river 179 width of 300 m), enabling to cover respectively 68% and 60%. At each observation segment, 180 two types of observation were conducted: counting of *entrapped* macroplastic and macrolit-181 ter, i.e.: items entrapped in hyacinth patches and counting of *free-floating* macroplastic and 182 macrolitter, i.e.: items freely floating at the water surface. 183

The mean plastic transport observation F [items/hour] for observation point i was calculated using the following equation:

$$F_{i} = \frac{N_{t,i}}{t_{t,i}} + \frac{N_{f,i}}{t_{f,i}}$$
(1)

Here, N_t is the plastic count of items [items] trapped in hyacinths and N_f plastic count of freefloating items [items] for observation point *i* during observation t_t and t_f [min], respectively. This distinction between trapped items and free-floating items enables to calculate the ratio of total trapped items over the total count of items, which is reported as a percentage [%]. The total floating plastic transport F_{total} [items/hour] was calculated using the following equation, derived from van Emmerik, de Lange, et al. (2022):

$$F_{total} = \sum_{i=1}^{n} \frac{F_i}{w_i} \cdot W \tag{2}$$

Here, w_i is the observation segment width [m], W the total river width [m]. The observation track width w_i [m] was estimated at 15 m for both measurement locations. We extrapolated floating plastic transport at an annual scale, considering both the mean and median F_{total} for all measurements done over the monitored period, thus calculating both the mean and median annual item transport [million items/year]. We also expressed floating plastic transport in terms of mass transport [tons/year], using the following equation (Vriend et al., 2020):

$$M = F_{total} \cdot \overline{m} \tag{3}$$

Here, \overline{m} expresses either the mean or the median mass per plastic item. We used both mean and median mass because other studies found that plastic transport estimates vary greatly depending on mass statistics (van Emmerik, de Lange, et al., 2022). We used the mass statistics from van Emmerik et al. (2019), who collected and weighted 3,022 items over 45 measurement days at the Saigon river. The mean mass was approximately 10 grams and the median mass 4.3 grams.

¹⁹⁰ 2.3 Hyacinth abundance

Hyacinth patches were detected using UAV imagery analysis. We used a color filtering 191 approach which enables to separate floating vegetation content from other elements present 192 at the river surface (e.g. water, banks, boats, wooden debris, floating items). This approach 193 leverages the color characteristics of active vegetation in the visible range to distinguish it 194 from other materials. A total of 3,562 UAV images was collected throughout the measure-195 ment period. To characterize hyacinth abundance, 3,544 images were ultimately processed. 196 A few images (n = 18) were discarded because these were blurry, taken with a side-angle or 197 due to the presence of boats which interfered with the hyacinth detection. Image processing 198

was done using the Open CV 4.5.4.60 library in Python 3.9.7. In addition to the color 199 filtering, we performed morphological operations over the images, involving noise reduction 200 and dilation to close small gaps. These operations and related parameters are detailed in 201 Supplementary Material (Extended Method). A minimum threshold area ($\geq =0.1 \text{ m}^2$) was also defined to filter out individual leaves and branches. All these operation parameters 203 were defined by trial and error through visual inspection, which was performed through 204 a subset of the total UAV image dataset. Trial and error sought to maximize detection 205 and minimize false positives as well as accurately detect the edges of the hyacinth patches. 206 Physical sampling of the patches to estimate plastic densities was not deemed feasible for 207 long-term monitoring, given that the patches typically move within minutes. More details 208 on the processing steps performed and their validation can be found in the Supporting Infor-209 mation (Extended Method). Fig.2 provides an example of hyacinth detection for one UAV 210 image. 211

We quantify hyacinth abundance in terms of coverage and count of patches. Hyacinth coverage $[km^2/km^2]$ was calculated as the total area covered by hyacinth over the total river area considered. The count of patches [#] is expressed as the number of total patches found per measurement unit. For both variables, four measurement units/scales were retained: image, flight, day and month. We include statistics on the mean size of hyacinth patches $[m^2]$ in section 2.3.

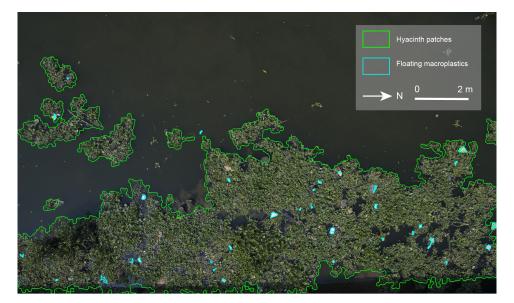


Figure 2: Example of processed UAV image [from 2 February 2021] with floating macroplastics and hyacinth patches identified.

218 2.4 Plastic densities

Plastic densities at the river surface and within hyacinth patches were also quantified 219 using UAV imagery analysis. The approach chosen is similar to the one described for hy-220 acinth detection in the previous section. The detection of floating plastic relied also on a 221 color filtering operation, which filtered pixels of white and light grey color. This approach 222 does not enable to detect all floating macroplastic and macrolitter items, which can be of 223 varying colors, and opacity and transparency levels. However, our visual assessment on 224 the entire dataset led to the conclusion that the majority ($\sim 70-90\%$) of macroplastic and 225 macrolitter items were of this color range. This is consistent with previous studies that quan-226

tified macrolitter composition in the Saigon river and demonstrated the high proportion of 227 items such as expanded polystyrene (food packaging, insulation foam), polystyrene (plastic 228 cups and cutlery) and soft polyolefins (plastic bags and foils) (Schreyers, van Emmerik, 229 Luan Nguyen, Castrop, et al., 2021; van Emmerik et al., 2019). Because of this limitation, 230 our estimates of plastic densities should be considered conservative. In addition to the color 231 filtering, morphological operations were also applied to the UAV imagery dataset, i.e. noise 232 reduction with Gaussian filtering and closing of gaps. Overall, processing steps for plastic 233 detection were less computer-intensive than for hyacinth patch detection, mainly due to the 234 smaller size of the objects of interest and the broader homogeneity of items compared to 235 hyacinth patches (edges were more easily distinguished for these anthropogenic items than 236 for the rather loose patches). Additional details on the processing operations and their 237 parameters can be found in the Supplementary Material (Extended Method). An example 238 of plastic detection for one UAV image can be seen on Fig. 1. 239

Plastic detection could only be implemented after manually removing (by cropping) the area affected by sun glint from each image. Sun glint pixels have the same color characteristics as the detected plastics. Cropping was therefore necessary to avoid false positive detection. Given that many images had a very large glint area, many were completely discarded for plastic detection (n = 1,202). More information on these aspects can be found in the Supporting Information (Extended Methods).

We calculated two types of plastic densities: river surface plastic density, expressing the number of items over the total river area considered and hyacinth plastic density [items/km²], expressing the number of items over the total hyacinth area considered. Plastic densities were expressed both as items densities [items/km²] and mass densities [kg/km²]. For mass densities, we used both the mean and median mass values per plastic item derived from van Emmerik et al. (2019), as described in section 2.2.

2.5 Additional data

252

To better understand plastic and hyacinth abundance in the Saigon river in relation to 253 hydrological processes and their seasonality, we used available data on rainfall and freshwa-254 ter discharge at the Saigon river. Rainfall and freshwater discharge are measured daily and 255 the resulting datasets are openly and freely available on the website of the Ho Chi Minh 256 City Irrigation Service Management company (http://www.dichvuthuyloi.com.vn/vn/tin-257 tuc/thong-tin-ve-tinh-hinh-dien-bien-khi-tuong-thuy-van-719/). We extracted all available 258 daily data on rainfall and freshwater discharge at the Saigon river for the year 2021, 259 corresponding to the measurement period for plastic transport, hyacinth abundance and 260 plastic densities. We used the rainfall data measured at the station Mac Dinh Chi, lo-261 cated in the first district of Ho Chi Minh City (latitude: 10.784223242113756; longitude: 262 106.69904438238632), as this is the closest rainfall measurement station from our measure-263 ment sites. River discharge is not measured within Ho Chi Minh City. River discharge 264 is measured in the Tây Ninh province, in the upstream area of the Saigon river and mea-265 surements correspond to the Dau Tieng reservoir inflow into the Saigon river. Monthly 266 cumulative rainfall [mm] and mean freshwater discharge $[m^3/s]$ were calculated based on 267 the above-mentioned rainfall and discharge data and are presented in Fig. S2. 268

269 2.6 Statistical analysis

The variables presented in the previous sections were aggregated at various temporal scales to identify temporal trends. We aggregated values by seasons, with the dry season spanning from December to May and the wet season from June to November, as rainfall and water flow seasonality are key components of the hydrological regime of the Saigon river (Camenen et al., 2021). To test whether the mean ranks of hyacinth coverage, plastic densities and plastic transport are significantly different between dry and wet seasons we used the Kruskal-Wallis test, which does not assume a normal distribution of the data. For the

daily and monthly aggregation levels we tested the Spearman correlations between pairs of 277 variables. The spatial distribution of plastic densities, plastic transport and hyacinth cover-278 age across the river was also investigated. The averaged cross-sectional spatial distribution 279 was calculated based on daily means for the metrics considered. We tested the similarity 280 in spatial distribution also using Spearman correlations. We characterized different regimes 281 (see Results and Discussion, section 3.2) of plastic transport and hyacinth coverage. For 282 this, we used the median values to distinguish between high and low categories of transport 283 and coverage values. 284

285

3 Results and Discussion

286

3.1 Plastic density in hyacinths ten times higher than at river surface

On average, between 55% and 65% of floating macroplastic is being transported by 287 hyacinth patches, depending on the location and the flow direction considered (L1, landward: 288 65%, seaward: 55%; L2, landward: 56%, seaward: 57%). We found that hyacinths cover an 289 average of 6% of the river surface, therefore indicating that patches trap much more floating 290 debris than could be hypothesized solely based on their relative coverage of the river surface. 291 This is confirmed by the discrepancies observed between river surface and hyacinth plastic 292 densities, with the latter being approximately one order of magnitude higher than the former 293 (mean river surface plastic density: $2.5 \cdot 10^4$ items/km² and mean hyacinth plastic density: 294 $2.1 \cdot 10^5$ items/km²) (Table 3). These results confirm that hyacinths act as physical traps 295 for floating plastics. Plastic transport in fluvial systems affected by hyacinth invasion are 296 therefore not only influenced by the two-way interactions between water and particles, but 297 are also likely affected by the movement of hyacinth at the water surface and changes in 298 patch coverage. These include the growth and reduction of individual patches, as well as 299 the aggregation and separation of patches among themselves. 300

Plastic item transport was estimated on average between 109 and 372 million items/year, 301 for L1 and L2 respectively (Table 3), approximately two orders of magnitude higher than the 302 top plastic polluted rivers in Europe (González-Fernández et al., 2021). Mean and median 303 plastic mass transport estimates vary by a factor of approximately two (Table 3), depending 304 on whether a mean or median mass per item was considered. This highlights the uncer-305 tainties associated with estimating plastic mass transport values. In addition, our estimates 306 focus on the total plastic transport (i.e. the total volume of plastic being transported in the 307 river, irrespective of the flow direction). Given that the Saigon river is strongly affected by 308 tidal dynamics, a distinction between total and net plastic transport (i.e. the total volume of 309 outgoing plastic) should be made in further studies and will be further discussed in section 310 4 (Synthesis and Outlook). 311

Mean item plastic densities at the river surface are 36 times higher $(2.5 \cdot 10^4 \text{ items/km}^2)$ 312 than those found in the Great Pacific Garbage Patch (GPGP) (6.9 ·10² items/km²) (Lebreton 313 et al., 2018). The average plastic mass densities found at the river surface (102-250 $\mathrm{kg/km^2}$ 314 for mean and median mass densities, respectively) are 3 to 6 times higher those observed in 315 the GPGP (mean mass density: 42 kg/km^2), a likely result of the heavier items found in the 316 ocean compared to river plastic. The highest plastic density found in our observations (4.7 317 $\cdot 10^5$ items/km²) is 190 times higher than the top density for the GPCP (2.4 $\cdot 10^3$ items/km²) 318 (Lebreton et al., 2018) and was measured for plastic trapped within hyacinths. Overall, this 319 comparison between river and ocean plastic densities supports the hypothesis that most 320 plastics is retained in rivers and not emitted into the oceans (van Emmerik, Mellink, et al., 321 2022). We also show that within rivers, aquatic floating vegetation such as hyacinths act 322 as physical traps of floating plastics, accumulating even higher densities of plastics than 323 otherwise found at the river surface. 324

Table 3: Floating transport and plastic densities estimates. We here report absolute values for floating plastic transport, irrespective of the flow direction.

			Floating	transport	;					
	Item tr	ansport	Mass transport							
	[:+	. /	Mean ma	ass/item	Median	mass/item				
	Items	s/year]	[tonnes	s/year]	[tonn	es/year]				
Location(s)	Median	Mean	Median	Mean	Median	Mean				
L1	90	109	903	1098	386	469				
L2	243	372	2447	3740	1045	1598				
	River surface plastic density									
	Item o	lensity	Mass density							
	[:+	$[km^2]$	Mean mas	ss density	Median mass density					
	Items	s/km j	[kg/l	km ²]	$[kg/km^2]$					
Location(s)	Median	Mean	Median	Mean	Median	Mean				
L1	$2.4 \cdot 10^4$	$2.5 \cdot 10^4$	239	250	102	107				
		H	Iyacinth p	lastic dens	sity					
	Item o	lensity		Mass	density					
	[:+	/121	Mean mas	ss density	Median r	nass density				
	litems	[items/km ²] [kg/km ²]				$/\mathrm{km}^2$]				
Location(s)	Median	Mean	Median	Mean	Median	Mean				
L1	$1.8 \cdot 10^{5}$	$2.1 \cdot 10^{5}$	1830	2107	782	900				

325 326

3.2 Temporal variability in hyacinth abundance and plastic accumulation and transport

All variables related to hyacinth abundance, plastic densities and transport have a 327 clear seasonality, with higher hyacinth and plastic loads during the dry season (Dec-May), 328 compared to the wet season (Jun-Nov) (Fig. 3). Only for the river surface plastic density 329 no significant statistical difference was found between dry and wet seasons (p-value=0.14); 330 however, the mean river surface plastic density was 1.3 times higher during the dry season 331 compared to the wet season (mean river surface plastic density for the dry and wet seasons, 332 respectively: $2.8 \cdot 10^4$ items/km² and $2.1 \cdot 10^4$ items/km²). Plastic transport variables (Fig. 333 3 E-H) have stronger significant values compared to metrics related to hyacinth abundance 334 and plastic densities, especially for the site L2 (downstream location). This study moni-335 tored hyacinth coverage at one location over the river (L1, upstream location), but results 336 are consistent with other studies that considered a larger geographic area. Janssens et al. 337 (2022) characterized hyacinth abundance over a larger portion (115 km of river length and 338 12,64 km²) of the Saigon river and showed that the dry season corresponds to higher wa-339 ter hyacinth abundance. Hyacinth coverage is the variable with the strongest correlation 340 with plastic transport (Spearman $\rho=0.86$, p-value < 0.05 for both L1 and L2) at a monthly 341 scale (Table 4). Plastic densities were not found to be significantly correlated with plastic 342 transport at a monthly scale. However, the Spearman correlation coefficients were found 343 to be quite high and p-values close to significance level (all p-values ≤ 0.2 and $\rho \geq 0.46$), 344 suggesting that such a relation might exist but is not highlighted with the current data 345 at a monthly scale, probably due to the relatively short time-series. Plastic densities were 346 found to be significantly correlated with the number of hyacinth patches (Spearman $\rho=0.82$) 347 p-value < 0.05 and Spearman ρ =0.68, p-value < 0.1 for hyacinth and river surface plastic den-348 sity, respectively) but not with hyacinth coverage at a monthly scale (p-value>0.1). This 349 highlights that high hyacinth plastic density values typically coincide with a high number 350 of patches, but not necessarily with large hyacinth coverage. 351

The monthly time-series provide a more detailed view of the seasonal cycle in hyacinth coverage, plastic loads and transport throughout the year (Fig.4). The peak in plastic transport occurs between March and May (Fig.4 A-B): March for the seaward transport at the downstream site, May at the upstream site and April for landward transport at both locations. The highest plastic densities at the river surface and within hyacinths

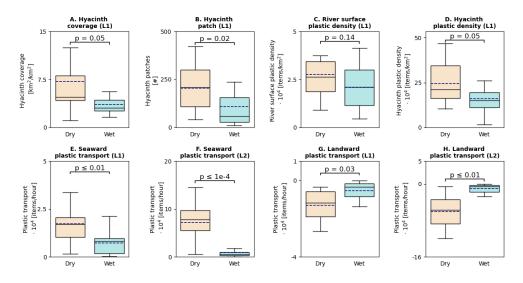


Figure 3: Seasonality at the Saigon river for A. Hyacinth coverage (L1) B. Hyacinth patch (L2). River surface plastic density (L1) D. Hyacinth plastic density (L1) E. Seaward plastic transport L2 F. Seaward plastic transport (L2). G. Landward plastic transport (L1). H. Landward plastic transport (L1). The blue dotted line indicates median values. Statistical differences between the dry (Dec-May) and wet (Jun-Nov) seasons were tested using the Krustal-Wallis test. p-values are indicated on top of each pair of boxplots. Values are considered statistically significant for p-value ≤ 0.05 .

Table 4: Spearman correlation coefficients between hyacinth abundance, plastic densities and transport variables. Variables were aggregated at both monthly and daily scales. Values marked with * indicate p-value<0.1, **< 0.05, ***< 0.01. The absence of sign indicates p>0.1

		h coverage	Hyacinth patch			
	$[\mathrm{km}^2/\mathrm{I}]$	m^{2}] (L1)	[#] (L1)			
	Monthly	Daily	Monthly	Daily		
River surface plastic density [items/km ²] (L1)	0.64	0.36^{*}	0.68*	0.02		
Hyacinth plastic density [items/km ²] (L1)	0.32	-0.29	0.82**	0.41^{**}		
Plastic transport [items/hours] (L1)	0.86^{**}	0.11	0.64	0.47**		
Plastic transport [items/hour] (L2)	0.86**	0.08	0.43	0.38*		
		plastic density		lastic density m^{21} (I 1)		
	[items/	km^{2}] (L1)	[items/l	m^{2}] (L1)		
II - dath						
Hyacinth coverage $[km^2/km^2]$ (L1)	[items/	km^{2}] (L1)	[items/l	m^{2}] (L1)		
	[items/ Monthly	km ²] (L1) Daily	[items/l Monthly	cm ²] (L1) Daily		
$[\mathrm{km}^2/\mathrm{km}^2]$ (L1) Hyacinth patch	[items/ Monthly 0.64	km ²] (L1) Daily 0.36*	[items/l Monthly 0.32	xm ²] (L1) Daily -0.29		
[km ² /km ²] (L1) Hyacinth patch [#] (L1) Plastic transport	[items/ Monthly 0.64 0.68*	$\frac{\mathrm{km}^2] (\mathrm{L1})}{\mathrm{Daily}}$ 0.36* 0.02	[items/l Monthly 0.32 0.82**	$\frac{\text{cm}^2 \text{[(L1)]}}{\text{Daily}}$ -0.29 0.41**		

are registered during the month of February. This also corresponds to the month with the 357 highest number of patches. Hyacinth coverage, on the other hand, is at its highest in March. 358 It should be noted however, that variables for plastic densities and hyacinth abundance were 359 not monitored during the month of April. Janssens et al. (2022) estimated hyacinth coverage 360 over three years at the Saigon river, using satellite imagery. The time-series analysis showed 361 that peaks in hyacinth typically occur between the end of February until the end of April. 362 May and June mark the decline in all the variables studied. These months correspond 363 to the start of the wet season over the Saigon river. For the year 2021, an increase in 364 discharge and rainfall was observed starting from April and intensified from June onward 365 (Supporting Information, Fig. S2). Few data were available between June and October, 366 thus limiting our understanding of the full cycle of plastic loads over the wet season and the 367 start of the post-monsoon season (Nov-Dec). van Emmerik et al. (2019)) observed a peak 368 in plastic transport in September and October, based on observations conducted in 2018. 369 Such a peak was not observed in the present study, despite the absence of data in August 370 and September. The following months (Oct-Dec) generally correspond to an increase in 371 all studied variables compared to the previous months (Jun-Sep). Overall, the monthly 372 variations in plastic transport, densities and hyacinth coverage show similar trends but are 373 not strictly synchronous. The noted discrepancies could result from gaps in data collection. 374 However, they could also indicate a temporal lag between the different processes of plastic 375 accumulation and transport. 376

At at a daily scale, hyacinth coverage and plastic transport are not significantly cor-377 related for both upstream and downstream locations (p-value>0.01) (Table 4). No signifi-378 cant correlations were found between river surface plastic density and plastic transport for 379 daily values either. Positive and statistically significant correlations were however found 380 for other variable combinations. Hyacinth plastic density (L1: Spearman $\rho=0.54$, pj0.01) 381 and hyacinth patch quantities (L1: Spearman $\rho=0.47$, p-value < 0.05, L2: Spearman $\rho=0.38$, 382 p-value < 0.01) have significant and positive relations with plastic transport for one or both 383 monitored locations at a daily scale. One reason for the absence of correlation at daily scale 384 between hyacinth coverage and plastic transport might be related to a temporal lag in the 385 processes of hyacinth abundance and plastic transport. Fig.5 A and B detail the time-series 386 of plastic transport, hyacinth coverage and river surface plastic density at L1 for two periods 387 (March and May-June 2021). Both time-series clearly show first a peak in plastic transport, 388 followed a few days later by an increase in hyacinth abundance and plastic densities (hy-389 acinth coverage and river surface plastic density). In March, the peak in hyacinth coverage 390 and plastic densities is asynchronous, with hyacinth coverage increasing 5 days before the 391 highest river plastic density is observed. This is not the case for the period of May-June, 392 where the peaks are registered on the same day. A likely explanation for this time lag be-393 tween the transport and accumulation processes pertains to the succession of mobilization 394 and retention processes. We hypothesize that high river discharge first mobilizes floating materials (including plastic and hyacinths), which get transported within the river system. 396 Then, reduced water flows (probably due to tidal dynamics and/or seasonality in the net 397 discharge) can cause a decrease in observed plastic transport for the same considered loca-398 tion. Simultaneously, low flow velocities cause the accumulation of plastic and hyacinths in 399 certain parts of the river channel, for instance on its lateral sections. At L2 (downstream 400 location), additional plastic inputs from the HCMC canals could also contribute to increased 401 plastic densities in low flow conditions. Plastic densities and hyacinth abundance increase 402 on the lateral sections of the river; until an increase again in discharge flushes the deposited 403 debris again. 404

Overall, plastic transport, plastic densities and hyacinth abundance are closely linked. With few exceptions, all the variables studied show a correlation with plastic transport either at a daily or monthly scale. For certain variables (e.g.: hyacinth coverage and river surface plastic density), the temporal lag observed in transport and accumulation processes demonstrates that plastic transport is best predicted when considering a wider time-frame than the daily scale. Satellite images are not available at a daily resolution with sufficiently high spatial resolution to detect hyacinths in rivers. Hyacinth coverage can be estimated
with freely available satellite imagery every 5 to 7 days (Janssens et al., 2022) for the same
location. This allows to build reliable monthly hyacinth coverage estimates, making it a
suitable proxy for plastic transport and accumulation in the Saigon river. The current
observations indicate that monthly means in hyacinth coverage can be a good predictor of
plastic transport.

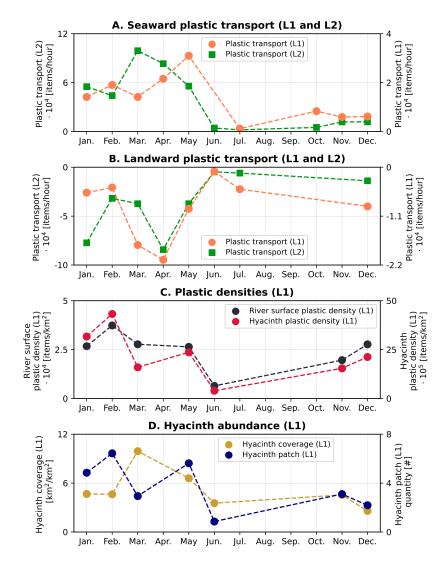


Figure 4: Monthly averages of variables related to plastic transport (A-B), plastic densities (C) and hyacinth abundance (D)

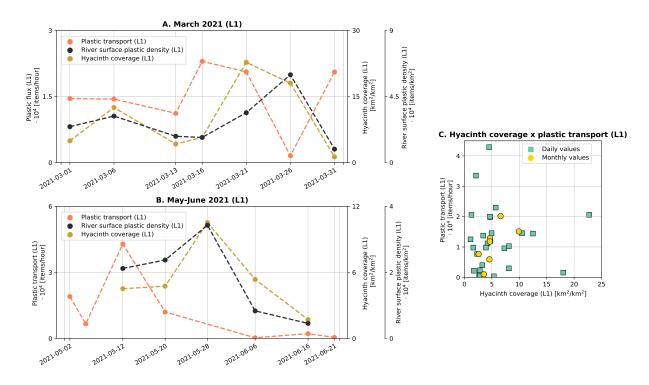


Figure 5: Observed daily values in hyacinth coverage, plastic transport and river surface plastic density at L1. A. Detailed time-series for the month of March 2021. B. Detailed time-series for the period of May-June 2021. C. Hyacinth coverage versus plastic transport at L1, daily and monthly mean values (Spearman $\rho = 0.11$ and 0.86, respectively, p-values>0.1 and <0.05).

3.3 Spatial variability in hyacinth abundance, plastic densities and plastic transport

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Plastic transport are approximately 3 to 4 times higher at L2 (downstream) than at 419 L1 (upstream). On average, the seaward transport is estimated at $4.4 \cdot 10^4$ items/hour for 420 L2 and $1.4 \cdot 10^4$ items/hour for L1. The average landward plastic transport is -4.9 $\cdot 10^4$ 421 items/hour for L2 and $-1.0 \cdot 10^4$ items/hour for L1. This difference in plastic transport 422 between locations could be explained by additional quantities of plastic inputted between 423 the monitored locations, a likely factor given that the river passes through Ho Chi Minh 424 City's urban area. In addition, stronger tidal influence at L2 compared to L1 probably 425 limits net discharge and net plastic transport, thus increasing plastic transport found in the 426 water regardless of additional plastic inputs between monitored locations. Our current data 427 did not quantify tidal dynamics and its effects on plastic transport, but lower net plastic 428 transport can be expected at L1 given its more upstream position in the river. 429

Plastic densities were not monitored during this study at L2, but we compared our 430 results for L1 with data from a previous study that reported such values for the same 431 month (May). Similarly to this study, Schreyers, van Emmerik, Luan Nguyen, Phung, 432 et al. (2021) used UAV imagery to estimate river surface plastic density, hyacinth plastic 433 density and hyacinth patch size. These estimates were done only for the month of May 434 2020 at L2, which we compare with the L1 values found for May 2021. At L2, hyacinth 435 plastic density was estimated at $2.1 \cdot 10^6$ items/km². In this study, we found a value of one 436 order of magnitude lower at L1 $(2.4 \cdot 10^5 \text{ items/km}^2, \text{ average for May 2021})$ than L2. River 437 surface plastic density was also found to be higher at L2 $(5.0 \cdot 10^5 \text{ items/km}^2)$ compared 438

Table 5: Plastic transport, densities and hyacinth coverage at L1 and L2. (1) indicates values from Schreyers, van Emmerik, Luan Nguyen, Phung, et al. (2021). (2) indicates values from Janssens et al. (2022). In the latter, hyacinth coverage was monitored along several large reaches of the Saigon river using satellite imagery. Two of the monitored section include L1 and L2. Plastic densities and average hyacinth patch size are reported for the month of May 2021 for this study and May 2020 to allow comparison across studies. Hyacinth coverage values are here reported as the average over a 3-year time-series.

	Seaward plastic transport [items/hour]	Landward plastic transport [items/hour]
L1	$1.4 \cdot 10^4$	-1.0 .104
L2	$5.0 \cdot 10^4$	$-4.5 \cdot 10^4$
	River surface plastic density	Hyacinth plastic density
	$[items/km^2]$	[items/km ²]
L1	$2.5 \cdot 10^4$	$2.2 \cdot 10^5$
L2	$5 \cdot 10^5 (1)$	$2.1 \cdot 10^6 (1)$
	Average patch size	Hyacinth coverage
	$[m^2]$	$[\mathrm{km}^2/\mathrm{km}^2]$
L1	1.5	$1.4 \cdot 10^1 (2)$
L2	0.82(1)	$9.5 \cdot 10^{-2} (2)$

to L1 $(2.6 \cdot 10^4 \text{ items/km}^2)$. The higher plastic densities found at L2 confirm that larger 439 riverine plastic quantities are present downstream. The increase in hyacinth plastic densities 440 downstream can also be partially explained by a decrease in hyacinth coverage between L1 441 and L2. Janssens et al. (2022) estimated hyacinth coverage continuously for three years 442 (2018-2020) over a large portion of the Saigon river, including the two locations of this 443 study. Between 2018 and 2020, on average, the midstream section (where L1 is situated) 444 had approximately 15 times larger hyacinth coverage than the downstream area (where L2 445 is located). In addition to a decrease in hyacinth coverage, hyacinth patches are also of a 446 smaller size downstream than upstream. Schreyers, van Emmerik, Luan Nguyen, Phung, et 447 al. (2021) estimated hyacinth patch average size at L2 at 0.82 m² in May 2020. In this study, 448 we found that hyacinth patches were on average twice as large in size at L1 (size of 1.5 m^2 , 449 average for May 2021). This decrease in hyacinth patch size is likely the result of mechanical 450 break-down due to boat traffic and possibly higher flow velocities (Petrell & Bagnall, 1991). 451 This comparison across studies bears many uncertainties, mainly because it assumes that 452 the temporal variation in hyacinth and plastic densities is negligible between May 2020 and 453 May 2021. Given the high temporal variability in plastic densities observed in this study, 454 and the intrannual variability in hyacinth coverage found in Janssens et al. (2022), such an 455 assumption is probably incorrect. For instance, between 2018-2020, hyacinth coverage was 456 found to vary by as much as a factor of eight for the month of May (Janssens et al., 2022). 457 This factor however, remains much lower than the difference found in hyacinth coverage 458 between L1 and L2 (of a factor of 15). We can therefore reasonably infer that hyacinth 459 coverage decrease and plastic transport and densities increase along the river course still 460 holds. Upstream of Ho Chi Minh City, hyacinth can cover a large extent of the river surface, 461 up to 24% of the river surface (Janssens et al., 2022). As the hyacinth drift downstream 462 of the city, patches get destabilized and break-down into smaller patches. Overall, the 463 hyacinth coverage decreases, covering on average less than 0.1% in its most downstream 464 section. Conversely, the plastic densities at the river surface and within hyacinth are higher 465 downstream than upstream of Ho Chi Minh City. The higher quantities in plastic result in 466 higher plastic transport downstream than upstream of the city. 467

In addition to spatial variation between upstream and downstream locations, the horizontal spatial variability (i.e.: across the river width) is also an important factor to understand the nexus between hyacinth abundance and plastic accumulation and transport

processes. Overall, we did not find that plastic densities, plastic transport and hyacinth 471 abundance all followed a similar horizontal spatial distribution (Fig. 6 A-B). Our findings 472 show that high transport of plastics can coincide with both high hyacinth coverage, which 473 occurs in the lateral reaches of the river; or with low hyacinth coverage in the middle of 474 channel. Our observations suggest that the drivers for these two high transport modes are 475 of different nature. The first is mainly driven by the mobilization of hyacinth patches, the 476 second is more closely tightened to variations in flow velocities and plastic quantities found 477 in the river. 478

479 Hyacinths tend to accumulate on the sides of the river channel, where the flow velocity is lower. Both the coverage and number of patches gradually decrease towards the middle 480 of the river channel (Fig. 6A). River surface plastic density follows a similar distribution 481 (Fig. 6B) and was found to be positively correlated with hyacinth abundance (hyacinth 482 coverage: $\rho=0.84$, p-value<0.01, hyacinth patch: $\rho=0.47$, p-value<0.05). A peak in river 483 surface plastic density was however observed at 80 m from the West bank, in a section 484 of the river with low hyacinth coverage (<4% on average). Hyacinth plastic density and 485 plastic transport, on the other hand, have a more complex and chaotic spatial distribution, 486 with a succession of peaks and drops in values (Fig. 6B). An overall trend is difficult to 487 establish. No strong significant correlation was found between these variables and hyacinth 488 abundance, or among themselves (all $\rho < 0.2$). For plastic transport, two main areas where 489 high plastic transport typically occur can be distinguished. One is at around 25 m from the 490 West riverbank, in an area with generally high hyacinth coverage and high plastic densities. 491 Plastic transport is also relatively high at approximately 120 m from the West riverbank, 492 in an section with low hyacinth coverage. The discrepancies in the spatial distribution of 493 plastic densities is explained by the fact that one considers the river area as its reference, and the other the hyacinth coverage. High hyacinth plastic densities can be observed in 495 areas with low surface plastic densities and hyacinth abundance, notably in the case of high 496 quantities of plastic present in small hyacinth patches. Overall, we can distinguish four 497 modes of transport and accumulation across the river (Fig. 6C). On both lateral sides of 498 the river channel high coverage of hyacinth dominates. This high accumulation is combined 499 with both low and high transport rates. Both hyacinth and plastic tend to accumulate in 500 this area, due to low current velocities. When the current increases, hyacinths get mobilized 501 in batches and important quantities of plastic and hyacinth are then washed out, resulting 502 in high plastic transport. On the lateral reaches of the river, plastic transport is therefore 503 intermittent, alternating periods of low plastic transport and high accumulation (deposition 504 dominated) with 'washed-out' periods (transport dominated). One main unknown is the 505 thresholds in flow conditions (stream velocity and water level) necessary to destabilize these 506 temporary deposition zones. In the middle reach of the river channel, both high and low 507 plastic transport rates can be found as well, hyacinth coverage is generally low. Hyacinth 508 patches do not cover large portion of the river surface there, are highly mobile and generally 509 510 present in small amount.

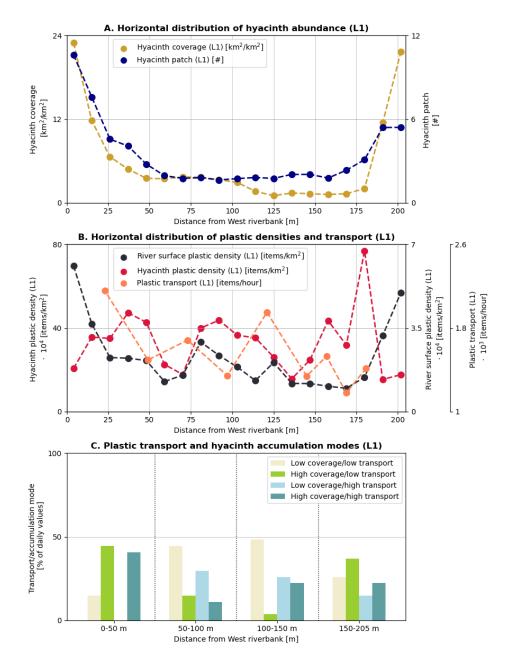


Figure 6: Horizontal distribution of hyacinth abundance (A) and plastic densities and transport (B) Plastic transport / hyacinth accumulation modes (C). Daily values were averaged across the river section.

511 4 Synthesis and Outlook

512 **4.1 Summary**

In this study, we demonstrated the role that hyacinths have in accumulating and transporting floating plastic. We found that $\sim 60\%$ of transported items are trapped within hyacinth patches, and that hyacinth plastic densities are on average one order of magnitude higher than otherwise found at the river surface. In comparison, the highest plastic densities found in the Great Pacific Garbage Patch are 190 times lower. Hyacinths function as

major temporary sinks for floating plastics; however this trapping effect varies greatly both 518 in time and space. Our analysis showed that on a temporal scale, high plastic transport and 519 hyacinth coverage tend to co-occur, especially when considering a monthly to seasonal scale. 520 This is likely the result of a time-lag between plastic transport and hyacinth coverage peak 521 events at a sub-monthly scale. Plastic densities, hyacinth coverage and plastic transport are 522 all higher during the dry season (Dec-May) when compared with the wet season (Jun-Nov) 523 At a spatial scale, we identified different transport modes in relation to hyacinth coverage. 524 Depending on the sections of the river, different mechanisms can explain high plastic trans-525 port rates. In the lateral sections of the river, low surface flow velocities and the abundance 526 of high hyacinth coverage promote the temporary deposition of large quantities of items, 527 with limited transport rates (Fig. 7A). Increased surface flow velocities mobilize in batches 528 of these temporary accumulation zones, leading to high plastic transport rates (Fig. 7B). 529 In the middle of the channel, plastic items are less affected in their trajectories by hyacinth-530 water interactions, and move therefore more freely at the water surface. We hypothesize 531 that the intermittent transport on the lateral reaches of the river is mainly governed by 532 semi-diurnal variations in river flow, caused by tidal dynamics. 533

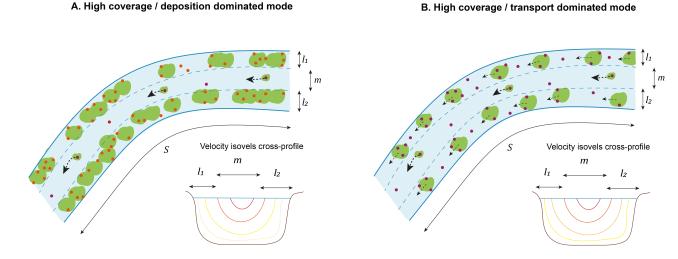


Figure 7: Variation in plastic transport modes on the lateral reaches of the river channel depending on hyacinth coverage. A. Deposition dominated mode, hyacinths and plastic have limited mobility during low flow conditions. B. Transport dominated mode, the hyacinths and plastic are mobilized in batches. S is the total longitudinal section of the river considered, l_1 and l_2 correspond to the lateral sections of the river, and m signifies the middle section. The cross-sectional views schematize the velocity isovels, with lower flow velocity on the lateral section of the river during a deposition dominated mode.

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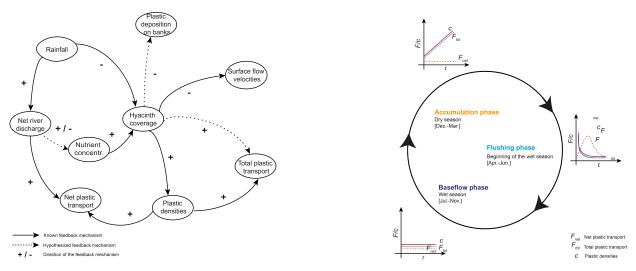
4.2 Conceptual model for plastic-hyacinth interactions

Fluvial plastic transport is affected by several hydro-meteorological and ecological fac-535 tors (Schreyers, van Emmerik, Luan Nguyen, Castrop, et al., 2021; Hurley et al., 2018; 536 Roebroek et al., 2021; van Emmerik, de Lange, et al., 2022) (Fig. 8A), of which hyacinth 537 coverage is a key component for tropical rivers. Low rainfall rates during the dry sea-538 son both limit freshwater discharge (Fig. S2) and net plastic transport, but generate an 539 increase in hyacinth coverage (Camenen et al., 2021; Harun et al., 2021; Janssens et al., 540 2022). This is likely the result of the higher nutrient concentrations found in the water 541 during periods of low net river discharge. In turn, increased hyacinth coverage also alters 542

plastic transport, with high rates of (temporary) deposition within hyacinths. In-stream 543 vegetation (floating and submerged) can function as a resistance force to water flows in 544 certain systems (Wharton et al., 2020; Sand-Jensen, 1998; Cornacchia et al., 2020) and 545 ultimately regulate surface flow velocities and water levels. We hypothesize that because 546 of this influence on the riverine flow dynamics, high hyacinth abundance also affects fluvial 547 plastic transport, by causing a (temporary) trapping of items, thus ultimately decelerating 548 transport transport. During the wet season, the lower coverage of hyacinths results in lower 549 deposition/accumulation rates of plastic items within patches compared to the dry season. 550 The role of hyacinths as aggregators and temporary sinks is therefore more limited during 551 this season. The fate of plastic could be affected by this in two distinct ways. Items could 552 flow more freely at the water surface, probably leading to longer transport trajectories. An-553 other likely scenario is that limited hyacinth coverage facilitates the contact and deposition 554 of plastic in other compartments, such as riparian vegetation or riverbanks. In such a case, 555 the higher hyacinth abundance during the dry season can be considered as a barrier to other 556 accumulation processes. Plastic deposition in these compartments would probably result in 557 longer deposition periods, because they can be considered more stable (e.g. less frequently 558 affected by hydrological dynamics). 559

The above-mentioned interactions between hyacinth coverage, plastic densities and net 560 discharge affect the seasonality in plastic transport (Fig. 8B). We can distinguish three 561 phases in the annual plastic transport cycle: an accumulation phase, a flushing phase and 562 a baseflow phase. The accumulation phase corresponds to the bulk of the dry season (Dec-563 Mar). During this phase, the Saigon's net discharge is low, with even negative net discharge 564 monthly values registered for some years (Camenen et al., 2021). Net discharge estimates 565 were not available for the year 2021. However, measured freshwater discharge and rainfall 566 rates in 2021 also suggest low net discharge rates for the period spanning from December 567 to March (Fig. S2). In this accumulation phase, plastic densities are gradually increasing 568 due to the cumulative effect of additional plastic inputs and limited net downstream plastic 569 transport. Most plastics therefore remain into the river, moving upstream and downstream 570 depending on the flow direction. High total plastic transport rates are observed, mainly 571 governed by the high plastic densities found in the river. A large part of the transported 572 items are most likely not flushed out of the system, because of the relative low net discharge. 573

At the beginning of the wet season (Apr-Jun), the increased net discharge generates a 574 575 flushing effect (Fig. S2). Most items are transported downstream and plastic densities in the river channel therefore decrease. Higher flow velocities destabilize hyacinths, which tend 576 to break-down more easily. Other studies also observed that increased precipitation rates 577 can be associated with the seasonal reduction in hyacinth coverage during the wet season 578 (Janssens et al., 2022; Harun et al., 2021). During the wettest months (Jul-Nov), rainfall 579 levels, freshwater discharge (Fig. S2) and thus net discharge (Camenen et al., 2021) are at 580 their highest. However, plastic transport rates are low during this period, as a result of a 581 drop in plastic densities during the previous flushing phase (baseflow phase). 582



A. Processes and feedback mechanisms governing riverine plastic transport B. Schematic annual cycle in plastic transport

Figure 8: Overview of plastic transport processes in the Saigon river, under the assumption of constant plastic influx. A. Processes and feedback mechanisms governing riverine plastic transport B. Schematic annual cycle in plastic transport.

4.3 Discontinuity in plastic transport

This study confirmed that hyacinths alter fluvial plastic transport by generating dis-584 continuous effects at various levels. The majority of plastics are trapped by hyacinths 585 $(\sim 60\%)$, despite hyacinth only covering $\sim 6\%$ of the river surface, thus confirming the role 586 of hyacinths as major accumulators of plastics (*physical discontinuity*). The fate of plastic 587 transport in rivers is therefore impacted not only by water-particle interactions but also 588 by the interactions with floating vegetation. As a result, the presence of hyacinths gener-589 ates different transport modes, with different accumulation and release dynamics between 590 areas where hyacinth are abundant and less affected areas. For instance, an intermittent 591 transport mode was observed on the lateral sections of the river (spatial discontinuity). In 592 addition, the seasonality in net river discharge similarly affects both hyacinth coverage and 593 plastic transport. In the accumulation phase, high hyacinth coverage alters the temporary 594 deposition and release dynamics of plastics, because items are more often and more likely 595 temporarily deposited and released by hyacinths. Without such large hyacinth coverage, the 596 deposition mechanisms of plastic would likely be dominated by interactions with the banks, 597 with seasonal release timescale, whereas transport mechanisms would be entirely governed 598 by daily flow dynamics (*temporal discontinuity*). 599

4.4 Outlook

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For tropical river systems heavily affected by tidal influence and seasonal variation in 601 river net discharge, such as the Saigon river, distinguishing between net and total plastic 602 transport is essential. Estimating the net plastic transport is however challenging, as it 603 requires to take into account: a) ebb and flood phases during the semi-diurnal tidal cycles; 604 b) neap and spring tides, and c) annual net discharge cycle. The net transport of items over 605 a river longitudinal section can be expressed as the product of seaward and landward plastic 606 transport within a predetermined area. Factors affecting the variation between seaward 607 and landward transport include tidal dynamics, variations in freshwater discharge and the 608

resulting seasonality in net water discharge, as well as deposition mechanisms in other river
 compartments.

Current observation techniques and protocols are limited in time and space, and do not 611 enable accurate quantification of net plastic transport in rivers. The timescales of observa-612 tions are inappropriate to estimate plastic mobility, especially in systems with fluctuating 613 transport regimes such as tidal rivers, and systems heavily affected by temporary deposition 614 mechanisms (for instance due to high floating vegetation presence). For our observations, 615 both landward and seaward plastic transport were considered, but monitoring was not car-616 ried throughout entire tidal cycles and thus we could not accurately quantify net plastic 617 transport. Continuous measurements throughout tidal cycles are needed to further charac-618 terize plastic mobility in rivers. Techniques such as tracking of the mobility of individual 619 particles, for instance with GPS trackers (Ledieu et al., 2022; Newbound, 2021; Tramoy 620 et al., 2020) and continuous transport measurements over tidal cycles could help in better 621 understanding plastic transport mechanisms. 622

Despite characterizing hyacinths as temporary sinks of plastics, we could not quantify 623 deposition times of plastics within hyacinths, nor do we know the hydraulic conditions at 624 which accumulations of hyacinth-plastic are entrained. This aspect is particularly impor-625 tant as it likely determines the alternation between mobile and relatively stable phases of 626 hyacinths-plastic mobility and ultimately affects the timescale of fluvial plastic transport. 627 Furthermore, to better understand the overall role of hyacinths as temporary sinks of plas-628 tics and how this relates to other components of the river system, studies on transport and 629 temporary deposition mechanisms across various riverine compartments are needed. Ex-630 tending direct observations is one way forward, but presents certain challenges. First, it 631 is usually time consuming and can sometimes be costly. Second, isolating the explaining 632 variables is often challenged by the inherent complexity and heterogeneity of the observed 633 systems. Another way forward could involve testing hypothesis on deposition and transport 634 dynamics of plastic within vegetation and other sinks through controlled laboratory exper-635 iments. This could be done for instance by building physical models that test under which 636 hydraulic conditions floating plastics are mobilized and deposited in various river compart-637 ments. Nevertheless, extending field-based research to other tropical systems is a necessary 638 step to further explore the role of hyacinth in trapping and transport plastics. 639

5 Conclusions

Hyacinth function as a major temporary sink for riverine floating plastics. Plastic densities in hyacinths were found to be 10 times higher than at the river surface and \sim 60% of the total transported items were trapped by hyacinth patches. These plant-plastic dynamics are not unique to the main observation location, as similar findings were also found for another location in the Saigon river. This suggests that the results are transferable to other sites within the river, as well as to other fluvial systems invaded by hyacinths.

Temporally, peaks in plastic transport and hyacinth coverage coincide, especially on 647 a monthly to seasonal scale. A time-lag in peak events was observed at a sub-monthly 648 scale. These findings suggests that to a certain extent, hyacinth coverage could be used as 649 a proxy for plastic pollution. In addition, we showed that hyacinths are a key component 650 in explaining plastic transport mechanisms. Peaks in plastic transport are caused either by 651 high transport governed by daily flow dynamics - mainly in the middle of the channel -, or 652 by high accumulation of hyacinth-plastics in the lateral sections of the river, which can be 653 mobilized in batches. 654

We linked hyacinth coverage and plastic accumulation to hydrological factors in a conceptual model, which can be used to explain spatio-temporal variations in plastic transport. A crucial aspect is the distinction between net and total river discharge, which likely drives changes in net/total plastic transport and hyacinth coverage at the river scale. We identified three phases (accumulation, flushing, baseflow phases) throughout the year which explain the annual variation in net and total plastic transport within the river.

Overall, hyacinth abundance in tropical rivers alters floating plastic transport because it interferes with the two-way interaction between water and plastic items. Because they trap the majority of plastic items, the mechanisms driving hyacinth movement and temporary deposition at the river surface also influence plastic propagation in rivers. As major temporary (and mobile) sinks of plastics, hyacinth abundance lead to increased discontinuity in plastic transport.

667 6 Data availability

All the UAV images used in this study are publicly available at https://doi.org/10.4121/21648152.v1.
 All remaining data will be made publicly available upon publication and have been included
 in the submission documents.

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Discontinuity in fluvial plastic transport increased by floating vegetation

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¹⁰ Key Points:

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11	•	Water hyacinths are major plastic sinks, with plastic densities up to ten times higher
12		than at the river surface.
13	•	Plastic transport, plastic densities and hyacinth abundance are closely linked, with
14		timing and location of accumulation coinciding.
15	•	Hyacinth coverage and plastic densities are affected by fluctuations in river dis-
16		charge which in turn impact plastic transport seasonality.

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17 Abstract

¹⁸ Understanding plastic mobility in rivers is crucial in estimating plastic emissions into ¹⁹ the oceans. Most studies have so far considered fluvial plastic transport as a uniform ²⁰ process, with stream discharge and plastic concentrations as the main variables necessary ²¹ to quantify plastic transport. Decelerating (e.g.: trapping effects) and accelerating effects ²² (e.g.: increased water flows) on plastic transport are poorly understood, despite growing ²³ evidence that such mechanisms affect riverine plastic mobility.

In this observation-based study, we explored the roles of an invasive floating plant species (i.e. water hyacinths) as a major disruptor of plastic transport. The different functions of aquatic vegetation in trapping and transporting plastics play a key part in our evolving understanding of how plastic moves in rivers. We collected a one-year dataset on plastic transport, densities and hyacinth abundance in the Saigon river, Vietnam, using both a visual counting method and UAV imagery analysis.

We found that hyacinths trap the majority of floating plastic observed ($\sim 60\%$), and plastic densities within patches are ten times higher than otherwise found at the river surface. At a monthly and seasonal scale, high hyacinth coverage coincides with peaks in both plastic transport and densities over the dry season (Dec-May) in the Saigon river.

We also investigated the large-scale mechanisms governing plant-plastic-water interactions through a conceptual model based on our observations and available literature. Distinguishing total and net plastic transport is crucial to consider fluctuations in freshwater discharge, tidal dynamics and trapping effects caused by the interactions with aquatic vegetation and/or other sinks.

39 1 Introduction

Plastic pollution poses a series of threats to global ecosystems, including aquatic systems 40 such as rivers. High levels of plastic pollution in rivers can reduce availability of potable 41 freshwater, cause damage to urban infrastructure, and potentially harm the local fauna 42 (van Emmerik & Schwarz, 2020). Rivers are considered the main pathways for land-based 43 plastic emissions into the oceans (Meijer et al., 2021). In addition, rivers can also retain 44 plastics for decades, if not longer (Tramoy et al., 2020). Understanding plastic mobility in 45 rivers is therefore crucial for risk assessments for riverine ecosystems under variable plastic 46 concentrations, and for accurate estimations of emissions into the oceans. 47

Rivers have long been considered as simple conduits for plastic transport to the sea. 48 Many studies portrayed the plastic journey in rivers to be a continuous trajectory of particles 49 through a uniform medium that offers little to no resistance to its final export into coastal 50 waters. As a result, plastic transport in rivers is often quantified as a direct function of 51 plastic concentrations in the water and river discharge (Schmidt et al., 2017; van Emmerik 52 et al., 2018; Haberstroh et al., 2021). However, recent scientific advances have shed light on 53 the discontinuous dynamics at play in fluvial plastic transport; at both temporal and spatial 54 scales. Temporally, plastic transport rates have been observed to follow seasonal patterns 55 and transport in various rivers (van Emmerik et al., 2019; van Emmerik, de Lange, et al., 56 2022), at times linked to seasonal variation in freshwater discharge. In addition, extreme 57 discharge events such as floods lead to disproportionally increased plastic transport rates 58 (Hurley et al., 2018; Roebroek et al., 2021; van Emmerik, Frings, et al., 2022). Spatially, 59 changes in river shape such as meander bends and the presence or absence of physical 60 barriers can lead to varying trapping rates, which affects plastic propagation in the water 61 (Newbound, 2021). Physical traps or barriers include infrastructure such as damns, groynes, 62 bridges and weirs, as well as bank and aquatic vegetation. These impediments can physically 63 retain plastic items temporarily or even permanently (Cesarini & Scalici, 2022; Schreyers, 64 van Emmerik, Luan Nguyen, Castrop, et al., 2021; Skalska et al., 2020). In addition, 65

varying plastic concentrations caused by human behaviours along the river (plastic leakage
and removal) contributes to spatially varying plastic transport rates. These discontinuities
likely lead to accelerating or decelerating effects of plastic distribution and propagation in
the water, similarly to what is observed for other floating debris such as wood (Wohl, 2017;
Wohl & Scott, 2017). As such, these discontinuities challenge the common assumption of a
uniform and unidirectional effect of river discharge on plastic mobility.

Aquatic vegetation can disrupt plastic mobility in rivers physically, spatially and tem-72 porally, and could therefore generate discontinuous effects in fluvial plastic transport. Veg-73 74 etation can trap plastic items, therefore leading to deposition and transport mechanisms that are affected by water-plant-plastic interactions (*physical discontinuity*). Vegetation 75 coverage varies due to the seasonal cycle, which, in turn, leads to higher or lower plastic 76 retention rates depending on the period of the year considered (temporal discontinuity). 77 Small scale variations in vegetation abundance along and/or across a given river might also 78 alter both plastic transport and deposition rates (*spatial discontinuity*). Here, we explore 79 the discontinuous nature of fluvial plastic transport by focusing on the role of an aquatic 80 vegetation species (e.g.: water hyacinths, *Eichhornia crassipes*) in trapping plastics in the 81 Saigon River. Hyacinths function as a major aggregator of floating macroplastics in trop-82 ical rivers and can, therefore, act as a dominant control factor of fluvial plastic transport 83 (Schreyers et al., 2021a; Schreyers et al., 2021b). These invasive aquatic species are now 84 present in most tropical lakes and rivers worldwide (CABI, 2020; Thamaga & Dube, 2019), 85 and their coverage of water surfaces can double in within one to two weeks due to their 86 rapid growth rate (Ouma et al., 2005). As a surface plant species, hyacinth float in patches 87 of varying sizes and densities. Their drift patterns are passive, and spatial distributions 88 are influenced by factors such as currents and wind. In low flow conditions, hyacinths can 89 rapidly blanket a large portion of the waterway. Kleinschroth et al. (Kleinschroth et al., 90 2021) found that for small reservoirs, peaks in hyacinth coverage often exceeded 80% of 91 the total reservoir area. Conversely, in more active systems like rivers, hyacinth coverage 92 tends to be lower due to the transport of the plants with water flow, but can still reach 93 up to 25% of the river surface (Janssens et al., 2022). Previous field-based studies have 94 successfully shown that hyacinths play a crucial role in fluvial plastic transport, however, 95 these observations were conducted over a short measurement period (6 weeks) and at only 96 one location. This study provides a much-needed more comprehensive understanding of how 97 hyacinth abundance alters fluvial plastic transport over both time and space. 98

For the present study, we monitored hyacinth coverage, plastic transport and plastic 99 densities in the Saigon river, Vietnam, over one year. The Saigon river has one of the 100 highest plastic transport rates in the world and is severely impacted by hyacinths invasion 101 (van Calcar & van Emmerik, 2019; Janssens et al., 2022). We hypothesize that hyacinths 102 function as a major temporary sink for riverine plastics and that therefore temporal peaks 103 and spatial accumulation zones in hyacinth coverage generally coincide with high plastic 104 loads. We first established the overall role of hyacinths as temporary traps for plastic items 105 (section 3.1). We then investigated the evolution of the measured metrics (e.g.: hyacinth 106 coverage, plastic transport and densities) at various temporal scales (seasonally, monthly and 107 daily) to characterize synchronous or asynchronous trends in transport and accumulation 108 (section 3.2). In addition, we analyzed how these variables are spatially distributed in the 109 river system, between upstream and downstream locations along the river and across the 110 river channel (section 3.3). The first part of this study focuses on quantifying hyacinth's role 111 as a temporary and mobile sink of floating plastic based on our field observations (section 3. 112 Results and Discussion). In the second part, we further expand on the interactions between 113 plastic-plant-water at a system scale (section 4. Synthesis and Conceptual model). We first 114 summarize our main findings which identified different modes of plastic transport in the river 115 in relation to hyacinth coverage (section 4.1). We present a conceptual model based on these 116 observational findings and our broader understanding of the fluvial system investigated, to 117 explain spatio-temporal variations in plastic transport (section 4.2). We thus synthesize 118 the discontinuous effects induced by hyacinth abundance on plastic transport (section 4.3) 119

and finally identify next steps in future research effort that seek to understand large-scale plastic transport and deposition processes in fluvial systems (section 4.4). The outcomes of this study are useful for scientists seeking to understand large-scale fluvial plastic transport and deposition mechanisms. In addition, river plastic monitoring and reduction strategies might seek to opportunistically use (temporary) sinks because of their role in aggregating large quantities in floating plastics.

¹²⁶ 2 Data and Methods

127 **2.1 Study area**

We measured plastic transport, hyacinth abundance and plastic densities between De-128 cember 12, 2020 and January 15, 2022 at the Saigon river, Vietnam (Fig. 1 and Table 1). 129 The Saigon river originates in Cambodia and flows into the Dau Tieng reservoir, approx-130 imately 120 km north from Ho Chi Minh City (Nguyen et al., 2020). The river crosses 131 agricultural areas of paddy rice and rubber plantation before entering the city. South of the 132 city, the Saigon river confluences with the Dong Nai river. There, the Dong-Nai-Saigon river 133 system branches into several channels that meanders in the Can Gio mangrove forest before 134 entering the East Sea (Dijksma et al., 2010). The Saigon river is subject to asymmetrical 135 semi-diurnal tidal cycle. Because of the tidal influence, the net river discharge is considered 136 relatively low and subject to seasonal variations between the dry and wet seasons (monthly 137 averages vary between -80 and $320 \text{ m}^3/\text{s}$) (Camenen et al., 2021). In addition, the Saigon 138 river is considered one of the most plastic polluted rivers worldwide, with transport rates 139 within the order of 10^4 items/hour (van Calcar & van Emmerik, 2019). Hyacinth invasions 140 are also particularly severe in this river, with peak coverage reaching up to 14% of the river 141 surface (Janssens et al., 2022). 142

This study focuses on floating macroplastic (>0.5 cm of size) density and transport, 143 hereafter referred to as plastic. Plastic transport was measured at two locations in Ho Chi 144 Minh City (Fig. 1). The first site (L1) is located north of the city (10.89025, 106.69209) 145 and the second (L2) in its southern part (latitude: 10.785984; longitude: 106.718332). The 146 two monitored sites approximately 30 km apart. At Ho Chi Minh City, the Saigon river 147 progresses from north to south, therefore enabling to compare upstream and downstream 148 plastic transport values within the urban area. Plastic transport was measured using the 149 visual counting method for floating bridges from bridges (section 2.2), and hyacinth abun-150 dance and plastic density were measured using Unmanned Aerial Vehicle (UAV) imagery 151 analysis (section 2.3 and 2.4). Flying at the downstream site was deemed unfeasible for long-152 term monitoring, due to the proximity of a military site. For this reason, UAV surveys were 153 only conducted at L1. UAV images were taken across the river channel, with a frequency of 154 one to four flights per measurement day. Each flight consisted of two overpasses across the 155 Saigon river, with a range of 41 to 65 images taken per flight. UAV surveys were carried 156 at a constant elevation of approximately 10 m above the water level. More information 157 on the UAV surveys is available in Supporting Information (Extended Methods). Table 2 158 summarizes the measurement frequencies per month at each location. Data gaps are no-159 ticeable for certain months: no data could be collected for any of the variables investigated 160 in August and September 2021. Due to the COVID-19 pandemic, a strict confinement was 161 mandated in Ho Chi Minh City, thus not allowing observers to leave their houses. A larger 162 data gap is noticeable for hyacinth abundance and plastic densities, with no measurements 163 conducted in April, July and October 2021. The gap during the month of April was due 164 to the unavailability of the observer conducting the UAV flights. The missing data from 165 July and October 2021 was also caused by COVID-19 restrictions. In those months, the 166 government did not allow inhabitants to cross the border between two different provinces, 167 thus not enabling access to the UAV flying site at L1 (a few hundred meters upstream of 168 where the visual counting measurements were conducted). 169

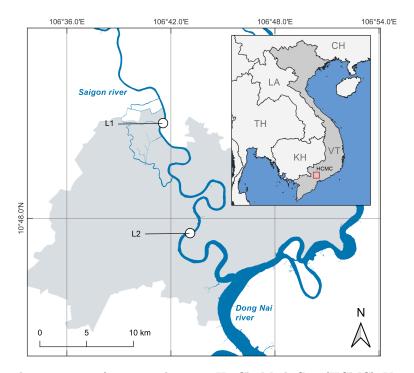


Figure 1: Localization map of monitored sites in Ho Chi Minh City (HCMC), Vietnam and measurement frequency at each location.

Table 1: Measurement frequency at each location. Total refers to the total number of UAV images analyzed in the case of hyacinth abundance and plastic densities. For plastic transport, it refers to the total number of observations, with one observation corresponding to a measurement per observation segment.

	Measurement locations								
	L	1	L2						
	Total	Daily	Total	Daily					
Plastic transport	900	49	1,272	51					
Hyacinth abundance	3,544	29	N/A	N/A					
Plastic densities	2,360	29	N/A	N/A					

Table 2: UAV images and plastic transport measurement frequency per month. The values here refer to the total number of UAV images for hyacinth abundance and plastic density. For plastic transport, the reported values correspond to the total number of observations. Blank cells indicate that no observations were conducted for that period.

	Number of measurements by month													
	Dec 20	Jan 21	Feb 21	Mar 21	Apr 21	May 21	Jun 21	Jul 21	Aug 21	Sep 21	Oct 21	Nov 21	Dec 21	Jan 22
Plastic transport (L1)	54	108	72	126	126	90	54	36			18	90	90	36
Plastic transport (L2)	84	144	83	168	168	120	72	46			89	110	110	44
Hyacinth abundance (L1)	142	536	141	935		407	186					550	363	284
Plastic densities (L1)	105	388	108	391		376	95					435	192	274

¹⁷⁰ 2.2 Floating plastic transport

Plastic transport were estimated using the visual counting method, developed by (González-171 Fernández & Hanke, 2017) and now widely used in observational studies on macroplastic 172 transport (González-Fernández et al., 2021; van Calcar & van Emmerik, 2019). All floating 173 macroplastic and macrolitter items (>0.5 cm) floating at the river surface were counted 174 during a determined time frame at each observation segment. Several observation segments 175 were determined per measurement location, to account for the spatial variability in plas-176 tic transport across the river width (van Emmerik et al., 2018). The number of segments 177 178 depends on the river width of the measurement location. Nine observation segments were selected at L1 (upstream site, river width of 200 m) and twelve at L2 (downstream site, river 179 width of 300 m), enabling to cover respectively 68% and 60%. At each observation segment, 180 two types of observation were conducted: counting of *entrapped* macroplastic and macrolit-181 ter, i.e.: items entrapped in hyacinth patches and counting of *free-floating* macroplastic and 182 macrolitter, i.e.: items freely floating at the water surface. 183

The mean plastic transport observation F [items/hour] for observation point i was calculated using the following equation:

$$F_{i} = \frac{N_{t,i}}{t_{t,i}} + \frac{N_{f,i}}{t_{f,i}}$$
(1)

Here, N_t is the plastic count of items [items] trapped in hyacinths and N_f plastic count of freefloating items [items] for observation point *i* during observation t_t and t_f [min], respectively. This distinction between trapped items and free-floating items enables to calculate the ratio of total trapped items over the total count of items, which is reported as a percentage [%]. The total floating plastic transport F_{total} [items/hour] was calculated using the following equation, derived from van Emmerik, de Lange, et al. (2022):

$$F_{total} = \sum_{i=1}^{n} \frac{F_i}{w_i} \cdot W \tag{2}$$

Here, w_i is the observation segment width [m], W the total river width [m]. The observation track width w_i [m] was estimated at 15 m for both measurement locations. We extrapolated floating plastic transport at an annual scale, considering both the mean and median F_{total} for all measurements done over the monitored period, thus calculating both the mean and median annual item transport [million items/year]. We also expressed floating plastic transport in terms of mass transport [tons/year], using the following equation (Vriend et al., 2020):

$$M = F_{total} \cdot \overline{m} \tag{3}$$

Here, \overline{m} expresses either the mean or the median mass per plastic item. We used both mean and median mass because other studies found that plastic transport estimates vary greatly depending on mass statistics (van Emmerik, de Lange, et al., 2022). We used the mass statistics from van Emmerik et al. (2019), who collected and weighted 3,022 items over 45 measurement days at the Saigon river. The mean mass was approximately 10 grams and the median mass 4.3 grams.

¹⁹⁰ 2.3 Hyacinth abundance

Hyacinth patches were detected using UAV imagery analysis. We used a color filtering 191 approach which enables to separate floating vegetation content from other elements present 192 at the river surface (e.g. water, banks, boats, wooden debris, floating items). This approach 193 leverages the color characteristics of active vegetation in the visible range to distinguish it 194 from other materials. A total of 3,562 UAV images was collected throughout the measure-195 ment period. To characterize hyacinth abundance, 3,544 images were ultimately processed. 196 A few images (n = 18) were discarded because these were blurry, taken with a side-angle or 197 due to the presence of boats which interfered with the hyacinth detection. Image processing 198

was done using the Open CV 4.5.4.60 library in Python 3.9.7. In addition to the color 199 filtering, we performed morphological operations over the images, involving noise reduction 200 and dilation to close small gaps. These operations and related parameters are detailed in 201 Supplementary Material (Extended Method). A minimum threshold area ($\geq =0.1 \text{ m}^2$) was also defined to filter out individual leaves and branches. All these operation parameters 203 were defined by trial and error through visual inspection, which was performed through 204 a subset of the total UAV image dataset. Trial and error sought to maximize detection 205 and minimize false positives as well as accurately detect the edges of the hyacinth patches. 206 Physical sampling of the patches to estimate plastic densities was not deemed feasible for 207 long-term monitoring, given that the patches typically move within minutes. More details 208 on the processing steps performed and their validation can be found in the Supporting Infor-209 mation (Extended Method). Fig.2 provides an example of hyacinth detection for one UAV 210 image. 211

We quantify hyacinth abundance in terms of coverage and count of patches. Hyacinth coverage $[km^2/km^2]$ was calculated as the total area covered by hyacinth over the total river area considered. The count of patches [#] is expressed as the number of total patches found per measurement unit. For both variables, four measurement units/scales were retained: image, flight, day and month. We include statistics on the mean size of hyacinth patches $[m^2]$ in section 2.3.

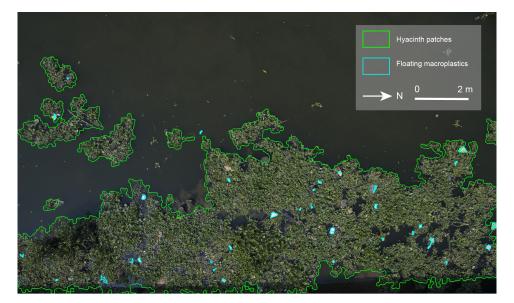


Figure 2: Example of processed UAV image [from 2 February 2021] with floating macroplastics and hyacinth patches identified.

218 2.4 Plastic densities

Plastic densities at the river surface and within hyacinth patches were also quantified 219 using UAV imagery analysis. The approach chosen is similar to the one described for hy-220 acinth detection in the previous section. The detection of floating plastic relied also on a 221 color filtering operation, which filtered pixels of white and light grey color. This approach 222 does not enable to detect all floating macroplastic and macrolitter items, which can be of 223 varying colors, and opacity and transparency levels. However, our visual assessment on 224 the entire dataset led to the conclusion that the majority ($\sim 70-90\%$) of macroplastic and 225 macrolitter items were of this color range. This is consistent with previous studies that quan-226

tified macrolitter composition in the Saigon river and demonstrated the high proportion of 227 items such as expanded polystyrene (food packaging, insulation foam), polystyrene (plastic 228 cups and cutlery) and soft polyolefins (plastic bags and foils) (Schreyers, van Emmerik, 229 Luan Nguyen, Castrop, et al., 2021; van Emmerik et al., 2019). Because of this limitation, 230 our estimates of plastic densities should be considered conservative. In addition to the color 231 filtering, morphological operations were also applied to the UAV imagery dataset, i.e. noise 232 reduction with Gaussian filtering and closing of gaps. Overall, processing steps for plastic 233 detection were less computer-intensive than for hyacinth patch detection, mainly due to the 234 smaller size of the objects of interest and the broader homogeneity of items compared to 235 hyacinth patches (edges were more easily distinguished for these anthropogenic items than 236 for the rather loose patches). Additional details on the processing operations and their 237 parameters can be found in the Supplementary Material (Extended Method). An example 238 of plastic detection for one UAV image can be seen on Fig. 1. 239

Plastic detection could only be implemented after manually removing (by cropping) the area affected by sun glint from each image. Sun glint pixels have the same color characteristics as the detected plastics. Cropping was therefore necessary to avoid false positive detection. Given that many images had a very large glint area, many were completely discarded for plastic detection (n = 1,202). More information on these aspects can be found in the Supporting Information (Extended Methods).

We calculated two types of plastic densities: river surface plastic density, expressing the number of items over the total river area considered and hyacinth plastic density [items/km²], expressing the number of items over the total hyacinth area considered. Plastic densities were expressed both as items densities [items/km²] and mass densities [kg/km²]. For mass densities, we used both the mean and median mass values per plastic item derived from van Emmerik et al. (2019), as described in section 2.2.

2.5 Additional data

252

To better understand plastic and hyacinth abundance in the Saigon river in relation to 253 hydrological processes and their seasonality, we used available data on rainfall and freshwa-254 ter discharge at the Saigon river. Rainfall and freshwater discharge are measured daily and 255 the resulting datasets are openly and freely available on the website of the Ho Chi Minh 256 City Irrigation Service Management company (http://www.dichvuthuyloi.com.vn/vn/tin-257 tuc/thong-tin-ve-tinh-hinh-dien-bien-khi-tuong-thuy-van-719/). We extracted all available 258 daily data on rainfall and freshwater discharge at the Saigon river for the year 2021, 259 corresponding to the measurement period for plastic transport, hyacinth abundance and 260 plastic densities. We used the rainfall data measured at the station Mac Dinh Chi, lo-261 cated in the first district of Ho Chi Minh City (latitude: 10.784223242113756; longitude: 262 106.69904438238632), as this is the closest rainfall measurement station from our measure-263 ment sites. River discharge is not measured within Ho Chi Minh City. River discharge 264 is measured in the Tây Ninh province, in the upstream area of the Saigon river and mea-265 surements correspond to the Dau Tieng reservoir inflow into the Saigon river. Monthly 266 cumulative rainfall [mm] and mean freshwater discharge $[m^3/s]$ were calculated based on 267 the above-mentioned rainfall and discharge data and are presented in Fig. S2. 268

269 2.6 Statistical analysis

The variables presented in the previous sections were aggregated at various temporal scales to identify temporal trends. We aggregated values by seasons, with the dry season spanning from December to May and the wet season from June to November, as rainfall and water flow seasonality are key components of the hydrological regime of the Saigon river (Camenen et al., 2021). To test whether the mean ranks of hyacinth coverage, plastic densities and plastic transport are significantly different between dry and wet seasons we used the Kruskal-Wallis test, which does not assume a normal distribution of the data. For the

daily and monthly aggregation levels we tested the Spearman correlations between pairs of 277 variables. The spatial distribution of plastic densities, plastic transport and hyacinth cover-278 age across the river was also investigated. The averaged cross-sectional spatial distribution 279 was calculated based on daily means for the metrics considered. We tested the similarity 280 in spatial distribution also using Spearman correlations. We characterized different regimes 281 (see Results and Discussion, section 3.2) of plastic transport and hyacinth coverage. For 282 this, we used the median values to distinguish between high and low categories of transport 283 and coverage values. 284

285

3 Results and Discussion

286

3.1 Plastic density in hyacinths ten times higher than at river surface

On average, between 55% and 65% of floating macroplastic is being transported by 287 hyacinth patches, depending on the location and the flow direction considered (L1, landward: 288 65%, seaward: 55%; L2, landward: 56%, seaward: 57%). We found that hyacinths cover an 289 average of 6% of the river surface, therefore indicating that patches trap much more floating 290 debris than could be hypothesized solely based on their relative coverage of the river surface. 291 This is confirmed by the discrepancies observed between river surface and hyacinth plastic 292 densities, with the latter being approximately one order of magnitude higher than the former 293 (mean river surface plastic density: $2.5 \cdot 10^4$ items/km² and mean hyacinth plastic density: 294 $2.1 \cdot 10^5$ items/km²) (Table 3). These results confirm that hyacinths act as physical traps 295 for floating plastics. Plastic transport in fluvial systems affected by hyacinth invasion are 296 therefore not only influenced by the two-way interactions between water and particles, but 297 are also likely affected by the movement of hyacinth at the water surface and changes in 298 patch coverage. These include the growth and reduction of individual patches, as well as 299 the aggregation and separation of patches among themselves. 300

Plastic item transport was estimated on average between 109 and 372 million items/year, 301 for L1 and L2 respectively (Table 3), approximately two orders of magnitude higher than the 302 top plastic polluted rivers in Europe (González-Fernández et al., 2021). Mean and median 303 plastic mass transport estimates vary by a factor of approximately two (Table 3), depending 304 on whether a mean or median mass per item was considered. This highlights the uncer-305 tainties associated with estimating plastic mass transport values. In addition, our estimates 306 focus on the total plastic transport (i.e. the total volume of plastic being transported in the 307 river, irrespective of the flow direction). Given that the Saigon river is strongly affected by 308 tidal dynamics, a distinction between total and net plastic transport (i.e. the total volume of 309 outgoing plastic) should be made in further studies and will be further discussed in section 310 4 (Synthesis and Outlook). 311

Mean item plastic densities at the river surface are 36 times higher $(2.5 \cdot 10^4 \text{ items/km}^2)$ 312 than those found in the Great Pacific Garbage Patch (GPGP) (6.9 ·10² items/km²) (Lebreton 313 et al., 2018). The average plastic mass densities found at the river surface (102-250 $\mathrm{kg/km^2}$ 314 for mean and median mass densities, respectively) are 3 to 6 times higher those observed in 315 the GPGP (mean mass density: 42 kg/km^2), a likely result of the heavier items found in the 316 ocean compared to river plastic. The highest plastic density found in our observations (4.7 317 $\cdot 10^5$ items/km²) is 190 times higher than the top density for the GPCP (2.4 $\cdot 10^3$ items/km²) 318 (Lebreton et al., 2018) and was measured for plastic trapped within hyacinths. Overall, this 319 comparison between river and ocean plastic densities supports the hypothesis that most 320 plastics is retained in rivers and not emitted into the oceans (van Emmerik, Mellink, et al., 321 2022). We also show that within rivers, aquatic floating vegetation such as hyacinths act 322 as physical traps of floating plastics, accumulating even higher densities of plastics than 323 otherwise found at the river surface. 324

Table 3: Floating transport and plastic densities estimates. We here report absolute values for floating plastic transport, irrespective of the flow direction.

	Floating transport							
	Item tr	ansport	Mass transport					
	[items/year]		Mean mass/item		Median mass/item			
			[tonnes/year]		[tonnes/year]			
Location(s)	Median	Mean	Median	Mean	Median	Mean		
L1	90	109	903	1098	386	469		
L2	243	372	2447	3740	1045	1598		
	River surface plastic density							
	Item o	lensity	density					
	$[items/km^2]$		Mean mass density		Median mass density			
			$[kg/km^2]$		$[kg/km^2]$			
Location(s)	Median	Mean	Median	Mean	Median	Mean		
L1	$2.4 \cdot 10^4$	$2.5 \cdot 10^4$	239	250	102	107		
	Hyacinth plastic density							
	Item density		Mass density					
	[items/km ²]		Mean mass density		Median mass density			
			$[kg/km^2]$		$[kg/km^2]$			
Location(s)	Median	Mean	Median	Mean	Median	Mean		
L1	$1.8 \cdot 10^{5}$	$2.1 \cdot 10^{5}$	1830	2107	782	900		

325 326

3.2 Temporal variability in hyacinth abundance and plastic accumulation and transport

All variables related to hyacinth abundance, plastic densities and transport have a 327 clear seasonality, with higher hyacinth and plastic loads during the dry season (Dec-May), 328 compared to the wet season (Jun-Nov) (Fig. 3). Only for the river surface plastic density 329 no significant statistical difference was found between dry and wet seasons (p-value=0.14); 330 however, the mean river surface plastic density was 1.3 times higher during the dry season 331 compared to the wet season (mean river surface plastic density for the dry and wet seasons, 332 respectively: $2.8 \cdot 10^4$ items/km² and $2.1 \cdot 10^4$ items/km²). Plastic transport variables (Fig. 333 3 E-H) have stronger significant values compared to metrics related to hyacinth abundance 334 and plastic densities, especially for the site L2 (downstream location). This study moni-335 tored hyacinth coverage at one location over the river (L1, upstream location), but results 336 are consistent with other studies that considered a larger geographic area. Janssens et al. 337 (2022) characterized hyacinth abundance over a larger portion (115 km of river length and 338 12,64 km²) of the Saigon river and showed that the dry season corresponds to higher wa-339 ter hyacinth abundance. Hyacinth coverage is the variable with the strongest correlation 340 with plastic transport (Spearman $\rho=0.86$, p-value < 0.05 for both L1 and L2) at a monthly 341 scale (Table 4). Plastic densities were not found to be significantly correlated with plastic 342 transport at a monthly scale. However, the Spearman correlation coefficients were found 343 to be quite high and p-values close to significance level (all p-values ≤ 0.2 and $\rho \geq 0.46$), 344 suggesting that such a relation might exist but is not highlighted with the current data 345 at a monthly scale, probably due to the relatively short time-series. Plastic densities were 346 found to be significantly correlated with the number of hyacinth patches (Spearman $\rho=0.82$, 347 p-value < 0.05 and Spearman ρ =0.68, p-value < 0.1 for hyacinth and river surface plastic den-348 sity, respectively) but not with hyacinth coverage at a monthly scale (p-value>0.1). This 349 highlights that high hyacinth plastic density values typically coincide with a high number 350 of patches, but not necessarily with large hyacinth coverage. 351

The monthly time-series provide a more detailed view of the seasonal cycle in hyacinth coverage, plastic loads and transport throughout the year (Fig.4). The peak in plastic transport occurs between March and May (Fig.4 A-B): March for the seaward transport at the downstream site, May at the upstream site and April for landward transport at both locations. The highest plastic densities at the river surface and within hyacinths

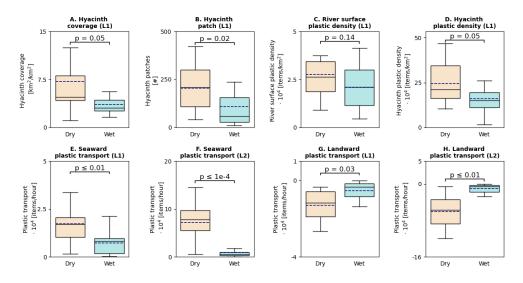


Figure 3: Seasonality at the Saigon river for A. Hyacinth coverage (L1) B. Hyacinth patch (L2). River surface plastic density (L1) D. Hyacinth plastic density (L1) E. Seaward plastic transport L2 F. Seaward plastic transport (L2). G. Landward plastic transport (L1). H. Landward plastic transport (L1). The blue dotted line indicates median values. Statistical differences between the dry (Dec-May) and wet (Jun-Nov) seasons were tested using the Krustal-Wallis test. p-values are indicated on top of each pair of boxplots. Values are considered statistically significant for p-value ≤ 0.05 .

Table 4: Spearman correlation coefficients between hyacinth abundance, plastic densities and transport variables. Variables were aggregated at both monthly and daily scales. Values marked with * indicate p-value<0.1, **< 0.05, ***< 0.01. The absence of sign indicates p>0.1

		h coverage	Hyacinth patch	
	$[\mathrm{km}^2/\mathrm{I}]$	m^{2} (L1)	[#] (L1)	
	Monthly	Daily	Monthly	Daily
River surface plastic density [items/km ²] (L1)	0.64	0.36^{*}	0.68*	0.02
Hyacinth plastic density [items/km ²] (L1)	0.32	-0.29	0.82**	0.41**
Plastic transport [items/hours] (L1)	0.86**	0.11	0.64	0.47**
Plastic transport [items/hour] (L2)	0.86**	0.08	0.43	0.38*
	River surface plastic density [items/km ²] (L1)		Hyacinth plastic density [items/km ²] (L1)	
			ι /	1()
TT : (1	Monthly	Daily	Monthly	Daily
Hyacinth coverage [km ² /km ²] (L1)	0.64	0.36^{*}	0.32	-0.29
Hyacinth patch [#] (L1)	0.68*	0.02	0.82**	0.41^{**}
Plastic transport [items/hours] (L1)	0.46	0.08	0.57	0.54^{***}
Plastic transport				

are registered during the month of February. This also corresponds to the month with the 357 highest number of patches. Hyacinth coverage, on the other hand, is at its highest in March. 358 It should be noted however, that variables for plastic densities and hyacinth abundance were 359 not monitored during the month of April. Janssens et al. (2022) estimated hyacinth coverage 360 over three years at the Saigon river, using satellite imagery. The time-series analysis showed 361 that peaks in hyacinth typically occur between the end of February until the end of April. 362 May and June mark the decline in all the variables studied. These months correspond 363 to the start of the wet season over the Saigon river. For the year 2021, an increase in 364 discharge and rainfall was observed starting from April and intensified from June onward 365 (Supporting Information, Fig. S2). Few data were available between June and October, 366 thus limiting our understanding of the full cycle of plastic loads over the wet season and the 367 start of the post-monsoon season (Nov-Dec). van Emmerik et al. (2019)) observed a peak 368 in plastic transport in September and October, based on observations conducted in 2018. 369 Such a peak was not observed in the present study, despite the absence of data in August 370 and September. The following months (Oct-Dec) generally correspond to an increase in 371 all studied variables compared to the previous months (Jun-Sep). Overall, the monthly 372 variations in plastic transport, densities and hyacinth coverage show similar trends but are 373 not strictly synchronous. The noted discrepancies could result from gaps in data collection. 374 However, they could also indicate a temporal lag between the different processes of plastic 375 accumulation and transport. 376

At at a daily scale, hyacinth coverage and plastic transport are not significantly cor-377 related for both upstream and downstream locations (p-value>0.01) (Table 4). No signifi-378 cant correlations were found between river surface plastic density and plastic transport for 379 daily values either. Positive and statistically significant correlations were however found 380 for other variable combinations. Hyacinth plastic density (L1: Spearman $\rho=0.54$, pj0.01) 381 and hyacinth patch quantities (L1: Spearman $\rho=0.47$, p-value < 0.05, L2: Spearman $\rho=0.38$, 382 p-value < 0.01) have significant and positive relations with plastic transport for one or both 383 monitored locations at a daily scale. One reason for the absence of correlation at daily scale 384 between hyacinth coverage and plastic transport might be related to a temporal lag in the 385 processes of hyacinth abundance and plastic transport. Fig.5 A and B detail the time-series 386 of plastic transport, hyacinth coverage and river surface plastic density at L1 for two periods 387 (March and May-June 2021). Both time-series clearly show first a peak in plastic transport, 388 followed a few days later by an increase in hyacinth abundance and plastic densities (hy-389 acinth coverage and river surface plastic density). In March, the peak in hyacinth coverage 390 and plastic densities is asynchronous, with hyacinth coverage increasing 5 days before the 391 highest river plastic density is observed. This is not the case for the period of May-June, 392 where the peaks are registered on the same day. A likely explanation for this time lag be-393 tween the transport and accumulation processes pertains to the succession of mobilization 394 and retention processes. We hypothesize that high river discharge first mobilizes floating materials (including plastic and hyacinths), which get transported within the river system. 396 Then, reduced water flows (probably due to tidal dynamics and/or seasonality in the net 397 discharge) can cause a decrease in observed plastic transport for the same considered loca-398 tion. Simultaneously, low flow velocities cause the accumulation of plastic and hyacinths in 399 certain parts of the river channel, for instance on its lateral sections. At L2 (downstream 400 location), additional plastic inputs from the HCMC canals could also contribute to increased 401 plastic densities in low flow conditions. Plastic densities and hyacinth abundance increase 402 on the lateral sections of the river; until an increase again in discharge flushes the deposited 403 debris again. 404

Overall, plastic transport, plastic densities and hyacinth abundance are closely linked. With few exceptions, all the variables studied show a correlation with plastic transport either at a daily or monthly scale. For certain variables (e.g.: hyacinth coverage and river surface plastic density), the temporal lag observed in transport and accumulation processes demonstrates that plastic transport is best predicted when considering a wider time-frame than the daily scale. Satellite images are not available at a daily resolution with sufficiently high spatial resolution to detect hyacinths in rivers. Hyacinth coverage can be estimated
with freely available satellite imagery every 5 to 7 days (Janssens et al., 2022) for the same
location. This allows to build reliable monthly hyacinth coverage estimates, making it a
suitable proxy for plastic transport and accumulation in the Saigon river. The current
observations indicate that monthly means in hyacinth coverage can be a good predictor of
plastic transport.

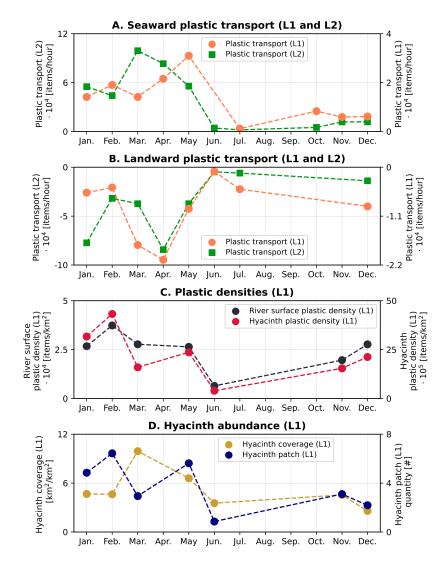


Figure 4: Monthly averages of variables related to plastic transport (A-B), plastic densities (C) and hyacinth abundance (D)

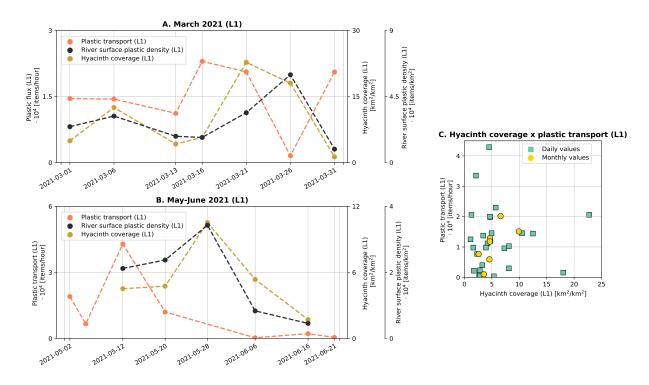


Figure 5: Observed daily values in hyacinth coverage, plastic transport and river surface plastic density at L1. A. Detailed time-series for the month of March 2021. B. Detailed time-series for the period of May-June 2021. C. Hyacinth coverage versus plastic transport at L1, daily and monthly mean values (Spearman $\rho = 0.11$ and 0.86, respectively, p-values>0.1 and <0.05).

3.3 Spatial variability in hyacinth abundance, plastic densities and plastic transport

417

418

Plastic transport are approximately 3 to 4 times higher at L2 (downstream) than at 419 L1 (upstream). On average, the seaward transport is estimated at $4.4 \cdot 10^4$ items/hour for 420 L2 and $1.4 \cdot 10^4$ items/hour for L1. The average landward plastic transport is -4.9 $\cdot 10^4$ 421 items/hour for L2 and $-1.0 \cdot 10^4$ items/hour for L1. This difference in plastic transport 422 between locations could be explained by additional quantities of plastic inputted between 423 the monitored locations, a likely factor given that the river passes through Ho Chi Minh 424 City's urban area. In addition, stronger tidal influence at L2 compared to L1 probably 425 limits net discharge and net plastic transport, thus increasing plastic transport found in the 426 water regardless of additional plastic inputs between monitored locations. Our current data 427 did not quantify tidal dynamics and its effects on plastic transport, but lower net plastic 428 transport can be expected at L1 given its more upstream position in the river. 429

Plastic densities were not monitored during this study at L2, but we compared our 430 results for L1 with data from a previous study that reported such values for the same 431 month (May). Similarly to this study, Schreyers, van Emmerik, Luan Nguyen, Phung, 432 et al. (2021) used UAV imagery to estimate river surface plastic density, hyacinth plastic 433 density and hyacinth patch size. These estimates were done only for the month of May 434 2020 at L2, which we compare with the L1 values found for May 2021. At L2, hyacinth 435 plastic density was estimated at $2.1 \cdot 10^6$ items/km². In this study, we found a value of one 436 order of magnitude lower at L1 $(2.4 \cdot 10^5 \text{ items/km}^2, \text{ average for May 2021})$ than L2. River 437 surface plastic density was also found to be higher at L2 $(5.0 \cdot 10^5 \text{ items/km}^2)$ compared 438

Table 5: Plastic transport, densities and hyacinth coverage at L1 and L2. (1) indicates values from Schreyers, van Emmerik, Luan Nguyen, Phung, et al. (2021). (2) indicates values from Janssens et al. (2022). In the latter, hyacinth coverage was monitored along several large reaches of the Saigon river using satellite imagery. Two of the monitored section include L1 and L2. Plastic densities and average hyacinth patch size are reported for the month of May 2021 for this study and May 2020 to allow comparison across studies. Hyacinth coverage values are here reported as the average over a 3-year time-series.

	Seaward plastic transport [items/hour]	Landward plastic transport [items/hour]
L1	$1.4 \cdot 10^4$	-1.0 .104
L2	$5.0 \cdot 10^4$	$-4.5 \cdot 10^4$
	River surface plastic density	Hyacinth plastic density
	$[items/km^2]$	[items/km ²]
L1	$2.5 \cdot 10^4$	$2.2 \cdot 10^5$
L2	$5 \cdot 10^5 (1)$	$2.1 \cdot 10^6 (1)$
	Average patch size	Hyacinth coverage
	$[m^2]$	$[\mathrm{km}^2/\mathrm{km}^2]$
L1	1.5	$1.4 \cdot 10^1 (2)$
L2	0.82(1)	$9.5 \cdot 10^{-2} (2)$

to L1 $(2.6 \cdot 10^4 \text{ items/km}^2)$. The higher plastic densities found at L2 confirm that larger 439 riverine plastic quantities are present downstream. The increase in hyacinth plastic densities 440 downstream can also be partially explained by a decrease in hyacinth coverage between L1 441 and L2. Janssens et al. (2022) estimated hyacinth coverage continuously for three years 442 (2018-2020) over a large portion of the Saigon river, including the two locations of this 443 study. Between 2018 and 2020, on average, the midstream section (where L1 is situated) 444 had approximately 15 times larger hyacinth coverage than the downstream area (where L2 445 is located). In addition to a decrease in hyacinth coverage, hyacinth patches are also of a 446 smaller size downstream than upstream. Schreyers, van Emmerik, Luan Nguyen, Phung, et 447 al. (2021) estimated hyacinth patch average size at L2 at 0.82 m² in May 2020. In this study, 448 we found that hyacinth patches were on average twice as large in size at L1 (size of 1.5 m^2 , 449 average for May 2021). This decrease in hyacinth patch size is likely the result of mechanical 450 break-down due to boat traffic and possibly higher flow velocities (Petrell & Bagnall, 1991). 451 This comparison across studies bears many uncertainties, mainly because it assumes that 452 the temporal variation in hyacinth and plastic densities is negligible between May 2020 and 453 May 2021. Given the high temporal variability in plastic densities observed in this study, 454 and the intrannual variability in hyacinth coverage found in Janssens et al. (2022), such an 455 assumption is probably incorrect. For instance, between 2018-2020, hyacinth coverage was 456 found to vary by as much as a factor of eight for the month of May (Janssens et al., 2022). 457 This factor however, remains much lower than the difference found in hyacinth coverage 458 between L1 and L2 (of a factor of 15). We can therefore reasonably infer that hyacinth 459 coverage decrease and plastic transport and densities increase along the river course still 460 holds. Upstream of Ho Chi Minh City, hyacinth can cover a large extent of the river surface, 461 up to 24% of the river surface (Janssens et al., 2022). As the hyacinth drift downstream 462 of the city, patches get destabilized and break-down into smaller patches. Overall, the 463 hyacinth coverage decreases, covering on average less than 0.1% in its most downstream 464 section. Conversely, the plastic densities at the river surface and within hyacinth are higher 465 downstream than upstream of Ho Chi Minh City. The higher quantities in plastic result in 466 higher plastic transport downstream than upstream of the city. 467

In addition to spatial variation between upstream and downstream locations, the horizontal spatial variability (i.e.: across the river width) is also an important factor to understand the nexus between hyacinth abundance and plastic accumulation and transport

processes. Overall, we did not find that plastic densities, plastic transport and hyacinth 471 abundance all followed a similar horizontal spatial distribution (Fig. 6 A-B). Our findings 472 show that high transport of plastics can coincide with both high hyacinth coverage, which 473 occurs in the lateral reaches of the river; or with low hyacinth coverage in the middle of 474 channel. Our observations suggest that the drivers for these two high transport modes are 475 of different nature. The first is mainly driven by the mobilization of hyacinth patches, the 476 second is more closely tightened to variations in flow velocities and plastic quantities found 477 in the river. 478

479 Hyacinths tend to accumulate on the sides of the river channel, where the flow velocity is lower. Both the coverage and number of patches gradually decrease towards the middle 480 of the river channel (Fig. 6A). River surface plastic density follows a similar distribution 481 (Fig. 6B) and was found to be positively correlated with hyacinth abundance (hyacinth 482 coverage: $\rho=0.84$, p-value<0.01, hyacinth patch: $\rho=0.47$, p-value<0.05). A peak in river 483 surface plastic density was however observed at 80 m from the West bank, in a section 484 of the river with low hyacinth coverage (<4% on average). Hyacinth plastic density and 485 plastic transport, on the other hand, have a more complex and chaotic spatial distribution, 486 with a succession of peaks and drops in values (Fig. 6B). An overall trend is difficult to 487 establish. No strong significant correlation was found between these variables and hyacinth 488 abundance, or among themselves (all $\rho < 0.2$). For plastic transport, two main areas where 489 high plastic transport typically occur can be distinguished. One is at around 25 m from the 490 West riverbank, in an area with generally high hyacinth coverage and high plastic densities. 491 Plastic transport is also relatively high at approximately 120 m from the West riverbank, 492 in an section with low hyacinth coverage. The discrepancies in the spatial distribution of 493 plastic densities is explained by the fact that one considers the river area as its reference, and the other the hyacinth coverage. High hyacinth plastic densities can be observed in 495 areas with low surface plastic densities and hyacinth abundance, notably in the case of high 496 quantities of plastic present in small hyacinth patches. Overall, we can distinguish four 497 modes of transport and accumulation across the river (Fig. 6C). On both lateral sides of 498 the river channel high coverage of hyacinth dominates. This high accumulation is combined 499 with both low and high transport rates. Both hyacinth and plastic tend to accumulate in 500 this area, due to low current velocities. When the current increases, hyacinths get mobilized 501 in batches and important quantities of plastic and hyacinth are then washed out, resulting 502 in high plastic transport. On the lateral reaches of the river, plastic transport is therefore 503 intermittent, alternating periods of low plastic transport and high accumulation (deposition 504 dominated) with 'washed-out' periods (transport dominated). One main unknown is the 505 thresholds in flow conditions (stream velocity and water level) necessary to destabilize these 506 temporary deposition zones. In the middle reach of the river channel, both high and low 507 plastic transport rates can be found as well, hyacinth coverage is generally low. Hyacinth 508 patches do not cover large portion of the river surface there, are highly mobile and generally 509 510 present in small amount.

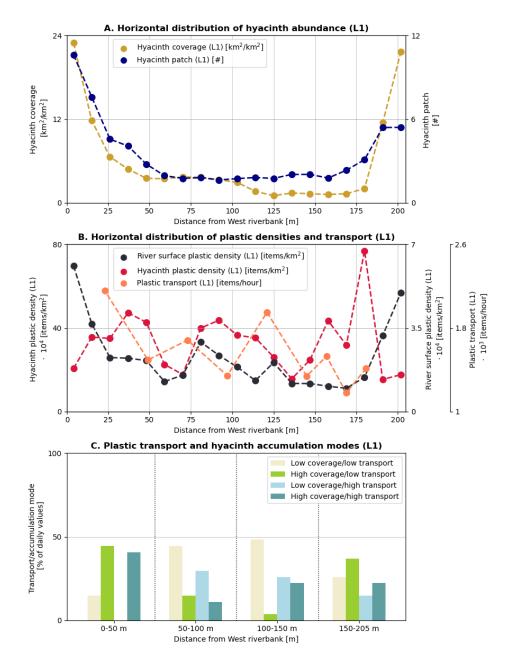


Figure 6: Horizontal distribution of hyacinth abundance (A) and plastic densities and transport (B) Plastic transport / hyacinth accumulation modes (C). Daily values were averaged across the river section.

511 4 Synthesis and Outlook

512 **4.1 Summary**

In this study, we demonstrated the role that hyacinths have in accumulating and transporting floating plastic. We found that $\sim 60\%$ of transported items are trapped within hyacinth patches, and that hyacinth plastic densities are on average one order of magnitude higher than otherwise found at the river surface. In comparison, the highest plastic densities found in the Great Pacific Garbage Patch are 190 times lower. Hyacinths function as

major temporary sinks for floating plastics; however this trapping effect varies greatly both 518 in time and space. Our analysis showed that on a temporal scale, high plastic transport and 519 hyacinth coverage tend to co-occur, especially when considering a monthly to seasonal scale. 520 This is likely the result of a time-lag between plastic transport and hyacinth coverage peak 521 events at a sub-monthly scale. Plastic densities, hyacinth coverage and plastic transport are 522 all higher during the dry season (Dec-May) when compared with the wet season (Jun-Nov) 523 At a spatial scale, we identified different transport modes in relation to hyacinth coverage. 524 Depending on the sections of the river, different mechanisms can explain high plastic trans-525 port rates. In the lateral sections of the river, low surface flow velocities and the abundance 526 of high hyacinth coverage promote the temporary deposition of large quantities of items, 527 with limited transport rates (Fig. 7A). Increased surface flow velocities mobilize in batches 528 of these temporary accumulation zones, leading to high plastic transport rates (Fig. 7B). 529 In the middle of the channel, plastic items are less affected in their trajectories by hyacinth-530 water interactions, and move therefore more freely at the water surface. We hypothesize 531 that the intermittent transport on the lateral reaches of the river is mainly governed by 532 semi-diurnal variations in river flow, caused by tidal dynamics. 533

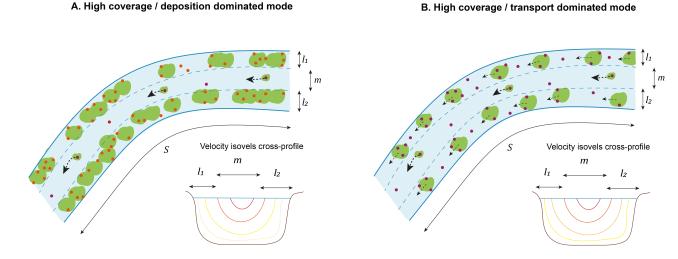


Figure 7: Variation in plastic transport modes on the lateral reaches of the river channel depending on hyacinth coverage. A. Deposition dominated mode, hyacinths and plastic have limited mobility during low flow conditions. B. Transport dominated mode, the hyacinths and plastic are mobilized in batches. S is the total longitudinal section of the river considered, l_1 and l_2 correspond to the lateral sections of the river, and m signifies the middle section. The cross-sectional views schematize the velocity isovels, with lower flow velocity on the lateral section of the river during a deposition dominated mode.

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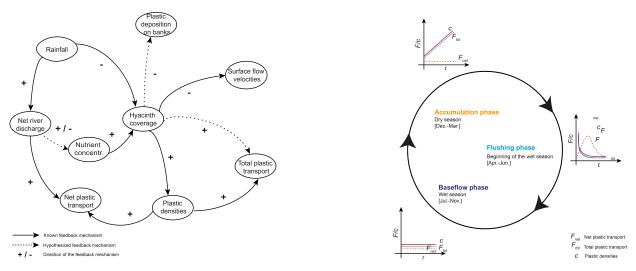
4.2 Conceptual model for plastic-hyacinth interactions

Fluvial plastic transport is affected by several hydro-meteorological and ecological fac-535 tors (Schreyers, van Emmerik, Luan Nguyen, Castrop, et al., 2021; Hurley et al., 2018; 536 Roebroek et al., 2021; van Emmerik, de Lange, et al., 2022) (Fig. 8A), of which hyacinth 537 coverage is a key component for tropical rivers. Low rainfall rates during the dry sea-538 son both limit freshwater discharge (Fig. S2) and net plastic transport, but generate an 539 increase in hyacinth coverage (Camenen et al., 2021; Harun et al., 2021; Janssens et al., 540 2022). This is likely the result of the higher nutrient concentrations found in the water 541 during periods of low net river discharge. In turn, increased hyacinth coverage also alters 542

plastic transport, with high rates of (temporary) deposition within hyacinths. In-stream 543 vegetation (floating and submerged) can function as a resistance force to water flows in 544 certain systems (Wharton et al., 2020; Sand-Jensen, 1998; Cornacchia et al., 2020) and 545 ultimately regulate surface flow velocities and water levels. We hypothesize that because 546 of this influence on the riverine flow dynamics, high hyacinth abundance also affects fluvial 547 plastic transport, by causing a (temporary) trapping of items, thus ultimately decelerating 548 transport transport. During the wet season, the lower coverage of hyacinths results in lower 549 deposition/accumulation rates of plastic items within patches compared to the dry season. 550 The role of hyacinths as aggregators and temporary sinks is therefore more limited during 551 this season. The fate of plastic could be affected by this in two distinct ways. Items could 552 flow more freely at the water surface, probably leading to longer transport trajectories. An-553 other likely scenario is that limited hyacinth coverage facilitates the contact and deposition 554 of plastic in other compartments, such as riparian vegetation or riverbanks. In such a case, 555 the higher hyacinth abundance during the dry season can be considered as a barrier to other 556 accumulation processes. Plastic deposition in these compartments would probably result in 557 longer deposition periods, because they can be considered more stable (e.g. less frequently 558 affected by hydrological dynamics). 559

The above-mentioned interactions between hyacinth coverage, plastic densities and net 560 discharge affect the seasonality in plastic transport (Fig. 8B). We can distinguish three 561 phases in the annual plastic transport cycle: an accumulation phase, a flushing phase and 562 a baseflow phase. The accumulation phase corresponds to the bulk of the dry season (Dec-563 Mar). During this phase, the Saigon's net discharge is low, with even negative net discharge 564 monthly values registered for some years (Camenen et al., 2021). Net discharge estimates 565 were not available for the year 2021. However, measured freshwater discharge and rainfall 566 rates in 2021 also suggest low net discharge rates for the period spanning from December 567 to March (Fig. S2). In this accumulation phase, plastic densities are gradually increasing 568 due to the cumulative effect of additional plastic inputs and limited net downstream plastic 569 transport. Most plastics therefore remain into the river, moving upstream and downstream 570 depending on the flow direction. High total plastic transport rates are observed, mainly 571 governed by the high plastic densities found in the river. A large part of the transported 572 items are most likely not flushed out of the system, because of the relative low net discharge. 573

At the beginning of the wet season (Apr-Jun), the increased net discharge generates a 574 575 flushing effect (Fig. S2). Most items are transported downstream and plastic densities in the river channel therefore decrease. Higher flow velocities destabilize hyacinths, which tend 576 to break-down more easily. Other studies also observed that increased precipitation rates 577 can be associated with the seasonal reduction in hyacinth coverage during the wet season 578 (Janssens et al., 2022; Harun et al., 2021). During the wettest months (Jul-Nov), rainfall 579 levels, freshwater discharge (Fig. S2) and thus net discharge (Camenen et al., 2021) are at 580 their highest. However, plastic transport rates are low during this period, as a result of a 581 drop in plastic densities during the previous flushing phase (baseflow phase). 582



A. Processes and feedback mechanisms governing riverine plastic transport B. Schematic annual cycle in plastic transport

Figure 8: Overview of plastic transport processes in the Saigon river, under the assumption of constant plastic influx. A. Processes and feedback mechanisms governing riverine plastic transport B. Schematic annual cycle in plastic transport.

4.3 Discontinuity in plastic transport

This study confirmed that hyacinths alter fluvial plastic transport by generating dis-584 continuous effects at various levels. The majority of plastics are trapped by hyacinths 585 $(\sim 60\%)$, despite hyacinth only covering $\sim 6\%$ of the river surface, thus confirming the role 586 of hyacinths as major accumulators of plastics (*physical discontinuity*). The fate of plastic 587 transport in rivers is therefore impacted not only by water-particle interactions but also 588 by the interactions with floating vegetation. As a result, the presence of hyacinths gener-589 ates different transport modes, with different accumulation and release dynamics between 590 areas where hyacinth are abundant and less affected areas. For instance, an intermittent 591 transport mode was observed on the lateral sections of the river (spatial discontinuity). In 592 addition, the seasonality in net river discharge similarly affects both hyacinth coverage and 593 plastic transport. In the accumulation phase, high hyacinth coverage alters the temporary 594 deposition and release dynamics of plastics, because items are more often and more likely 595 temporarily deposited and released by hyacinths. Without such large hyacinth coverage, the 596 deposition mechanisms of plastic would likely be dominated by interactions with the banks, 597 with seasonal release timescale, whereas transport mechanisms would be entirely governed 598 by daily flow dynamics (*temporal discontinuity*). 599

4.4 Outlook

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For tropical river systems heavily affected by tidal influence and seasonal variation in 601 river net discharge, such as the Saigon river, distinguishing between net and total plastic 602 transport is essential. Estimating the net plastic transport is however challenging, as it 603 requires to take into account: a) ebb and flood phases during the semi-diurnal tidal cycles; 604 b) neap and spring tides, and c) annual net discharge cycle. The net transport of items over 605 a river longitudinal section can be expressed as the product of seaward and landward plastic 606 transport within a predetermined area. Factors affecting the variation between seaward 607 and landward transport include tidal dynamics, variations in freshwater discharge and the 608

resulting seasonality in net water discharge, as well as deposition mechanisms in other river
 compartments.

Current observation techniques and protocols are limited in time and space, and do not 611 enable accurate quantification of net plastic transport in rivers. The timescales of observa-612 tions are inappropriate to estimate plastic mobility, especially in systems with fluctuating 613 transport regimes such as tidal rivers, and systems heavily affected by temporary deposition 614 mechanisms (for instance due to high floating vegetation presence). For our observations, 615 both landward and seaward plastic transport were considered, but monitoring was not car-616 ried throughout entire tidal cycles and thus we could not accurately quantify net plastic 617 transport. Continuous measurements throughout tidal cycles are needed to further charac-618 terize plastic mobility in rivers. Techniques such as tracking of the mobility of individual 619 particles, for instance with GPS trackers (Ledieu et al., 2022; Newbound, 2021; Tramoy 620 et al., 2020) and continuous transport measurements over tidal cycles could help in better 621 understanding plastic transport mechanisms. 622

Despite characterizing hyacinths as temporary sinks of plastics, we could not quantify 623 deposition times of plastics within hyacinths, nor do we know the hydraulic conditions at 624 which accumulations of hyacinth-plastic are entrained. This aspect is particularly impor-625 tant as it likely determines the alternation between mobile and relatively stable phases of 626 hyacinths-plastic mobility and ultimately affects the timescale of fluvial plastic transport. 627 Furthermore, to better understand the overall role of hyacinths as temporary sinks of plas-628 tics and how this relates to other components of the river system, studies on transport and 629 temporary deposition mechanisms across various riverine compartments are needed. Ex-630 tending direct observations is one way forward, but presents certain challenges. First, it 631 is usually time consuming and can sometimes be costly. Second, isolating the explaining 632 variables is often challenged by the inherent complexity and heterogeneity of the observed 633 systems. Another way forward could involve testing hypothesis on deposition and transport 634 dynamics of plastic within vegetation and other sinks through controlled laboratory exper-635 iments. This could be done for instance by building physical models that test under which 636 hydraulic conditions floating plastics are mobilized and deposited in various river compart-637 ments. Nevertheless, extending field-based research to other tropical systems is a necessary 638 step to further explore the role of hyacinth in trapping and transport plastics. 639

5 Conclusions

Hyacinth function as a major temporary sink for riverine floating plastics. Plastic densities in hyacinths were found to be 10 times higher than at the river surface and \sim 60% of the total transported items were trapped by hyacinth patches. These plant-plastic dynamics are not unique to the main observation location, as similar findings were also found for another location in the Saigon river. This suggests that the results are transferable to other sites within the river, as well as to other fluvial systems invaded by hyacinths.

Temporally, peaks in plastic transport and hyacinth coverage coincide, especially on 647 a monthly to seasonal scale. A time-lag in peak events was observed at a sub-monthly 648 scale. These findings suggests that to a certain extent, hyacinth coverage could be used as 649 a proxy for plastic pollution. In addition, we showed that hyacinths are a key component 650 in explaining plastic transport mechanisms. Peaks in plastic transport are caused either by 651 high transport governed by daily flow dynamics - mainly in the middle of the channel -, or 652 by high accumulation of hyacinth-plastics in the lateral sections of the river, which can be 653 mobilized in batches. 654

We linked hyacinth coverage and plastic accumulation to hydrological factors in a conceptual model, which can be used to explain spatio-temporal variations in plastic transport. A crucial aspect is the distinction between net and total river discharge, which likely drives changes in net/total plastic transport and hyacinth coverage at the river scale. We identified three phases (accumulation, flushing, baseflow phases) throughout the year which explain the annual variation in net and total plastic transport within the river.

Overall, hyacinth abundance in tropical rivers alters floating plastic transport because it interferes with the two-way interaction between water and plastic items. Because they trap the majority of plastic items, the mechanisms driving hyacinth movement and temporary deposition at the river surface also influence plastic propagation in rivers. As major temporary (and mobile) sinks of plastics, hyacinth abundance lead to increased discontinuity in plastic transport.

667 6 Data availability

All the UAV images used in this study are publicly available at https://doi.org/10.4121/21648152.v1.
 All remaining data will be made publicly available upon publication and have been included
 in the submission documents.

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Supporting Information for "Discontinuity in fluvial plastic transport increased by floating vegetation"

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Supporting Information 1: Extended method

UAV surveys

We used the DJI Phantom 3 UAV, with comes with a FC6310 camera, equipped with a 1/2.3 inch CMOS sensor. The sensor has a maximum resolution of 12.76 megapixels and a camera resolution of 2992 x 3992 pixels. The UAV operated automatically, from take-off to landing. The programming was done with the Drone Harmony app. All images were captured at nadir, i.e. perpendicular (90°± 0.02°) to the direction of the flight, to facilitate surface calculations. Each flight lasted approximately ten minutes. The UAV imagery analysis involved coverage detection of hyacinths. The pixel area had to be converted to real-ground area, by calculating the ground sampling distance (d_g)

X - 2

[m/pixels], as follows:

$$d_g = \frac{S_w \cdot H_f}{F_l \cdot w_i} \tag{1}$$

⁷ Here, S_w is the sensor width of the camera [m], H_f is the flight height [m], F_l is the ⁸ focal length of the camera [m] and w_i is the image width [pixels]. All variables as the ⁹ camera used did not change and the flight height was set at 10 m. A d_g value of $3.8 \cdot 10^{-3}$ ¹⁰ m/pixel was found.

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Hyacinth and plastic detection with UAV imagery analysis

In this section, we detail the processing steps taken for both hyacinth and plastic de-11 tection (Fig. S1). The choice in the RGB threshold values was done by trial and errors 12 over a subset of the imagery dataset. For the hyacinth detection, the same threshold 13 values were applied for all the analyzed images. For the detection of plastic, changes in 14 brightness between images did not allow to use the same threshold values for the entire 15 dataset. A few combinations were therefore retained and tested over batches of images 16 (corresponding usually to the same measurement day). The best fitting threshold values 17 were retained for the batch of images analyzed. For hyacinth detection, images were then 18 blurred with a Gaussian filter, to reduce noise. Noise in hyacinth detection is the result 19 of the configuration of patches. In general, patches were relatively loose (with gaps and 20 holes in-between) with highly irregular edges. Various filter sizes were tested (see Sen-21 sitivity analysis in the Validation subsection). Ultimately, a filter size of 13×13 pixels 22 was retained for the hyacinth detection. No Gaussian blurring was necessary for the de-23 tection of plastic items, as the target objects are of relatively small size and the detection 24 approach sought to maximize edge detection from the background elements rather than 25 reduce noise. For hyacinth detection, a dilate operation was necessary to reduce unnec-26

essary details at the edges of patches. A final kernel size of 17×17 pixels was selected after trial and errors through visual inspection. A fill in (e.g.: binary closing) operation was performed for both detection approaches. This allows to fill in small gaps within the detected objects of interest. The closing is applied around a circle of a specified diameter [in pixels]. A diameter of 10 pixels was chosen for both hyacinth and plastic detection.

Sun glint and false positives with plastic detection

No recurring distinct shapes of sun glints that could be of use to automatically filter these areas out were recognized throughout the entire UAV imagery dataset. We do not deemed feasible therefore to implement an automatic detection of sun glint and opted for manual removal of sun glint affected area, using a simple cropping operation. The cropping was done by batch of images. In images taken during the same UAV flight and same overpass direction, the area covered by sun glint was generally located in the same region of the images.

Validation

³⁹ Sensitivity analysis for hyacinth detection

We explored the sensitivity of the output variables for hyacinth abundance [hyacinth coverage and count of patches] to variations in input parameters for the three morphological operations performed (Gaussian blur, dilate and fill-in operations). The sensitivity analysis was performed over a representative subset of the imagery dataset (n = 156 images, 4% of the total number of images analyzed). We performed a Mood's median test to compare the median of the two datasets. The alpha risk value was set at 0.05. We found a p-value > 0.05 (p-value=0.11), indicating that the null hypothesis is

47 confirmed and no significant difference can be assumed between the two sample
48 populations.

For each morphological parameter, we calculated the change in output values for the count of patches and mean and median coverage area [%], based on changes in input parameters [%]. Changes in input parameters were computed for approximately -50, -30, -10, 10, 30 and 50%. Given that kernel sizes have to be odd numbers, small deviations from the above-mentioned changes in input were sometimes necessary to fulfill this requirement. Ultimately, we expressed the sensitivity in terms of slope factor [%], calculated as the ratio between the change of output and the change of input parameters:

$$s = \frac{c_o}{c_i} \tag{2}$$

⁵⁶ Here, c_o is the change in output parameter and c_i in input parameter. The sensitivity ⁵⁷ analysis results (Table S1) show that the dilate parameter is the most sensitive, with a ⁵⁸ higher dilate kernel leading to a lower number of patches and higher hyacinth coverage.

59 Assessment of plastic detection

We assessed the accuracy of our detection approach of floating plastic items by manually 60 labelling items on a subset of our dataset (n = 273, 10% of the image dataset used for 61 plastic detection). This validation set of images was selected randomly, using the sample 62 function in Python. We again performed a Mood's median test to compare the median 63 of the two datasets and test the whether the validation set can be considered 64 representative of the entire imagery dataset. We found a p-value > 0.05 (p-value=0.22), 65 indicating that the null hypothesis is confirmed and no significant difference can be 66 assumed between the two sample populations. 67

⁶⁸ We manually identified and counted all floating items, irrespective of their size, on the ⁶⁹ validation set. An accuracy ratio [%] a_r was computed for each image, as follows:

$$a_r = 100\% - \frac{|c_d - c_m|}{c_m} \cdot 100\%$$
(3)

Here, c_d is the total number of floating items detected with the detection approach on a 70 given image and c_m the total number of floating items manually labelled. The overall 71 accuracy ratio [%] was computed as the mean of accuracy ratios per image. We found 72 an overall accuracy ratio of 75%. The number of floating items was found to be exactly 73 the same between the validation and our detection approaches for 52% of the images (n 74 = 141). For 37% of the images (n = 102), the detection approach underestimated the 75 number of floating items when compared with the manual labelling. Only for a minority 76 of the images (11%, n = 31) the detection approach overestimated the number of 77 floating plastic items. 78

Supporting Information 2: Rainfall and freshwater discharge at the Saigon river

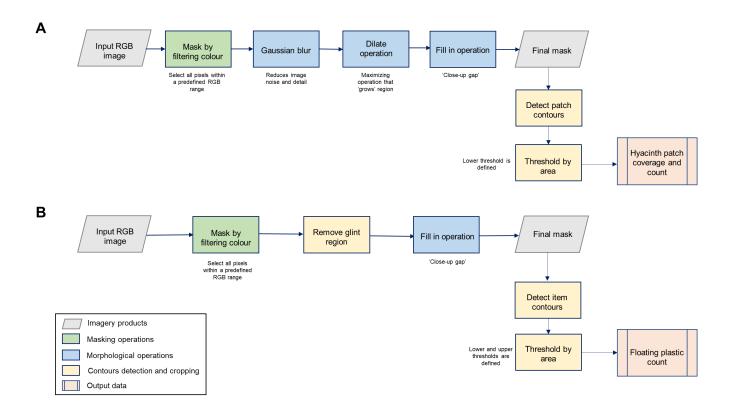
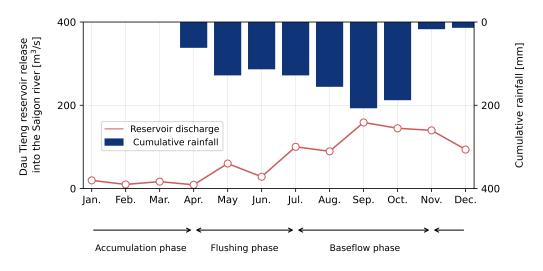


Figure S1: Processing steps to detect: A. Hyacinth patches and B. Floating plastic items.

Table S1: Sensitivity analysis for input parameters (morphological operations) in hyacinth detection on UAV images. This table reports the slope factor s, expressed in %.

	Dilate	Gaussian	Closing
Hyacinth patch	-54	-21	-5
Mean hyacinth coverage	55	25	4
Median hyacinth coverage	64	28	12



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Figure S2: Monthly rainfall and freshwater discharge at the Saigon river, for the year 2021. The rainfall data was monitored at the Mc inh Chi station in District 1, Ho Chi Minh City. The freshwater discharge (mean values) from the Dau Tieng reservoir into the Saigon river was measured at the Tây Ninh station. The three phases indicated refer to plastic transport/hyacinth coverage phases, as conceptualized in section 4.2.