

Component and ensemble density feedbacks decoupled by density-independent processes

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Abstract

Analysis of long-term trends in abundance provide insights into population dynamics. Population growth rates are the emergent interplay of fertility, survival, and dispersal, but the density feedbacks on some vital rates (component) can be decoupled from density feedback on population growth rates (ensemble). However, the mechanisms responsible for this decoupling are poorly understood. We simulated component density feedbacks on survival in age-structured populations of long-living vertebrates and quantified how imposed nonstationarity (density-independent mortality and variation in carrying-capacity) modified the ensemble feedback signal estimated from logistic-growth models to the simulated abundance time series. The statistical detection of ensemble density feedback was largely unaffected by density-independent processes, but catastrophic and proportional mortality eroded the effect of density-dependent survival on ensemble-feedback strength more strongly than variation in carrying capacity. Thus, phenomenological models offer a robust approach to capture density feedbacks from nonstationary census data when density-independent mortality is low.

1 **ECOLOGY LETTERS**

2 **Letter**

3

4 **Component and ensemble density feedbacks decoupled by density-**
5 **independent processes**

6

7 **Running title:** Phenomena hiding signals of density feedback

8

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33

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36 Population growth rates are the emergent interplay of fertility, survival, and dispersal, but the
37 density feedbacks on some vital rates (component) can be decoupled from density feedback
38 on population growth rates (ensemble). However, the mechanisms responsible for this
39 decoupling are poorly understood. We simulated component density feedbacks on survival in
40 age-structured populations of long-living vertebrates and quantified how imposed
41 nonstationarity (density-independent mortality and variation in carrying-capacity) modified
42 the ensemble feedback signal estimated from logistic-growth models to the simulated
43 abundance time series. The statistical detection of ensemble density feedback was largely
44 unaffected by density-independent processes, but catastrophic and proportional mortality
45 eroded the effect of density-dependent survival on ensemble-feedback strength more strongly
46 than variation in carrying capacity. Thus, phenomenological models offer a robust approach
47 to capture density feedbacks from nonstationary census data when density-independent
48 mortality is low.

50 **INTRODUCTION**

51 Compensatory density feedback describes a population's ability to return to the
52 environment's carrying capacity in response to an increase in population size (*sensu*
53 Herrando-Pérez *et al.* 2012b). This phenomenon is driven by adjustments to individual fitness
54 imposed by variation in per-capita resource availability, and associated processes of
55 predation, competition, parasitism, and dispersal (Fowler 1981; Matthysen 2005; Eberhardt *et al.*
56 2008; Herrando-Pérez *et al.* 2012a). As survival and fertility rates ebb and flow in
57 response to variation in population density, it is theoretically possible to detect the density-
58 feedback signal in time series of abundance monitored at regular intervals over a sufficient
59 period (Brook & Bradshaw 2006; Herrando-Pérez *et al.* 2012a). There is now considerable
60 evidence that survival and fertility track population trends in many vertebrate (Eberhardt
61 2002; Paradis *et al.* 2002; Owen-Smith & Mason 2005; Pardo *et al.* 2017; Saunders *et al.*
62 2018; Doyle *et al.* 2020; Margalida *et al.* 2020; Morrison *et al.* 2021; Stillman *et al.* 2021)
63 and invertebrate (Azerefegne *et al.* 2001; Bonsall & Benmayor 2005; McGeoch & Price
64 2005; Jepsen *et al.* 2009; Maud *et al.* 2015; Marini *et al.* 2016; Ma 2021) species. Therefore,
65 given the irreplaceable importance of long-term monitoring of population size in applied
66 ecology and conservation (Herrando-Pérez *et al.* 2012a), assessing the presence of

67 compensatory signals in censuses of population abundance remains an essential tool in the
68 ecologist's toolbox (Bellier *et al.* 2016).

69 The family of self-limiting population-growth models including logistic growth curves
70 ('phenomenological models' hereafter) are convenient for describing density-feedback
71 signals in abundance time series (Eberhardt *et al.* 2008). These models use census data to
72 quantify the net effect of population size N on the rate of instantaneous growth r (Berryman
73 & Turchin 2001). Expressed as a proportional change in N between two time (t) steps (e.g.,
74 years or generations), the assumption is that $r = \log_e(N_{t+1}/N_t)$ summarises the combination or
75 'ensemble' (Herrando-Pérez *et al.* 2012a) of all feedback mechanisms operating on
76 individual 'component' demographic rates (Münster-Swendsen & Berryman 2005). The
77 problem is that population growth rates can be insensitive to variation in particular
78 demographic rates (Kolb *et al.* 2010; Battaile & Trites 2013; Bürgi *et al.* 2015). Thus, across
79 109 observed censuses of bird and mammal populations, the strength of 'component density
80 feedback' (on demographic rates) explained only $< 10\%$ of the strength of 'ensemble density
81 feedback' (on population growth rate) using phenomenological models and after controlling for
82 time-series length and body size (Herrando-Pérez *et al.* 2012a). The reasons for such
83 decoupling are not well understood.

84 Determining the partial effects of different underlying mechanisms responsible for the
85 decoupling of component and ensemble density feedbacks is most often impossible for real
86 abundance time series. This analytical limitation occurs because the multiple, density-
87 dependent and -independent mechanisms generating population fluctuations change
88 themselves through time — a condition known as 'nonstationarity' (*sensu* Turchin 2003). We
89 therefore constructed stochastic, age-structured populations with known, component density
90 feedback on survival and imposed nonstationarity to population size via multiple
91 demographic scenarios emulating density-independent mortality and variation in carrying
92 capacity through time. We then simulated multiannual time series of abundance from those
93 populations and estimated the strength of ensemble density feedbacks from these. Our
94 prediction was that ensemble density feedbacks should track component feedbacks if survival
95 has a demographic impact, mediated by population size, on the population growth rate of
96 long-lived vertebrates, while our demographic framework allowed the quantification of true
97 and false detection of ensemble density feedbacks.

98

99 **METHODS**

100 Our overarching aim was to simulate populations of long-living species and their time series
101 of abundance with various sources of nonstationarity. We describe below the set of test
102 species, the simulation of the base population model, component density feedbacks on
103 survival and time series of population abundance, the demographic scenarios considered, and
104 the phenomenological models used to quantify ensemble density feedbacks.

105

106 ***Test species***

107 As the variability in population growth rates is driven primarily by survival rates for slower
108 life-history species of mammals (Heppell *et al.* 2000; Oli & Dobson 2003) and birds (Sæther
109 & Bakke 2000), we parameterised the simulated populations to characterise the plausible
110 dynamics of 21 long-lived species of extant ($n = 8$) and extinct ($n = 13$) Australian
111 vertebrates from five taxonomic/functional groups (herbivore vombatiformes and
112 macropodiformes, large omnivore birds, carnivores, and invertivore monotremes), spanning
113 mean adult body masses of 1.7–2786 kg and generation lengths of 2.3–21 years (Bradshaw *et al.*
114 *al.* 2021; Table 1). These species differ in their resilience to environmental change, and
115 represent the slow end of the slow-fast continuum of life histories (Herrando-Pérez *et al.*
116 2012c) where high survival rates make it possible that reproductive efforts are dispersed over
117 the lifetime of individuals (Gaillard *et al.* 1989). A full justification of the selection of our
118 test species can be found in Bradshaw *et al.* (2021).

119

120 ***Base (age-structured) population model***

121 The population model for each test species was a stochastic (parameters resampled within
122 their uncertainty bounds) Leslie transition matrix (\mathbf{M}) following a pre-breeding design, with
123 $\omega+1$ (i) \times $\omega+1$ (j) elements (representing ages from 0 to ω years) for females only, where ω
124 represents maximum longevity. Fertility (m_x) occupied the first row of the matrix, survival
125 probabilities (S_x) occupied the sub-diagonal, and the final diagonal transition probability
126 ($\mathbf{M}_{i,j}$) was S_ω for all species — except *Vombatus ursinus* (VU; common wombat), *Thylacinus*
127 *cynocephalus* (TC; thylacine) and *Sarcophilus harrisii* (SH; devil) for which we set $S_\omega = 0$ to
128 limit unrealistically high proportions of old individuals in the population given the evidence
129 for catastrophic mortality at ω for the latter two species (Holz & Little 1995; Cockburn 1997;
130 Oakwood *et al.* 2001). Multiplying \mathbf{M} by a population vector \mathbf{n} estimates total population size
131 ($\Sigma \mathbf{n}$) at each forecasted time step (Caswell 2001). The base model was parameterised with \mathbf{n}_0
132 $= A\mathbf{D}\mathbf{M}\mathbf{w}$, where \mathbf{w} is the right eigenvector of \mathbf{M} (stable stage distribution), and A is the
133 surface area of the study zone ($A = 250,000 \text{ km}^2$) so that the species with the lowest \mathbf{n}_0 would

134 have an initial population of at least several thousand individuals at the start of the
 135 simulations. Based on theoretical equilibrium densities (D , km⁻²) calculated for each taxon
 136 (Bradshaw *et al.* 2021), the species-specific carrying capacity $K = DA$.

137

138 ***Density feedback on survival***

139 We simulated a compensatory density-feedback function by forcing a reduction modifier
 140 (S_{red}) of the S_x vector in each model according to $\Sigma \mathbf{n}$:

$$141 \quad S_{\text{red}} = \frac{a}{1 + \left(\frac{\Sigma \mathbf{n}}{b}\right)^c} \quad \text{[eq 1]}$$

142 where the a , b , and c constants for each species are adjusted to produce a stable population on
 143 average over 40 generations (40[G]; see below) (Brook *et al.* 2006; Traill *et al.* 2010). This
 144 formulation avoided exponentially increasing populations, optimised transition matrices to
 145 produce parameter values as close as possible to the maximum potential rates of
 146 instantaneous increase (r_m) (Bradshaw *et al.* 2021), and so ensured that long-term population
 147 dynamics were approximately stable at the species-specific K (see previous section). Here,

$$148 \quad G = \frac{\log\left((\mathbf{v}^T \mathbf{M})_1\right)}{\lambda_1} \quad \text{[eq 2]}$$

149 $(\mathbf{v}^T \mathbf{M})_1$ is the dominant eigenvalue of the reproductive matrix \mathbf{R} derived from \mathbf{M} , and \mathbf{v} is the
 150 left eigenvector (Caswell 2001) of \mathbf{M} . Thus, the total projection length in years (q) varied
 151 across the 21 test species, from 92 (*Dasyurus maculatus*; DM; spot-tailed quoll) to 800
 152 (*Genyornis newtoni*; GN; mihirung) years (median = 324 years with 95 % interquartiles of
 153 [108, 762] years; Table 1), with one value of abundance being simulated per year. In each
 154 iteration and annual time step, the S_x vector was β -resampled assuming a 5% standard
 155 deviation of each S_x and a Gaussian-resampled m_x vector. We deliberately avoided applying
 156 density-feedback functions to fertility to isolate the component feedback to a single
 157 demographic rate.

158

159 ***Nonstationarity***

160 We added nonstationarity to our base population model through a catastrophic (density-
 161 independent) mortality function to account for the probability of a catastrophic event (C)
 162 scaling to generation length among vertebrates (Reed *et al.* 2003):

$$163 \quad C = \frac{p_C}{G} \quad \text{[eq 3]}$$

164 where p_C = probability of catastrophe set at 0.14 given this is the mean probability per
 165 generation observed across vertebrates (Reed *et al.* 2003). Once invoked at probability C , a β -

166 resampled proportion centred on 0.5 to the β -resampled S_x vector induced a $\sim 50\%$ mortality
 167 event for that year (Bradshaw *et al.* 2013). A catastrophic event is defined as “... any 1-yr
 168 peak-to-trough decline in estimated numbers of 50% or greater” (Reed *et al.* 2003). The
 169 catastrophic function recreates the demographic effects of a density-independent process such
 170 as extreme weather events, fires, disease outbreaks, or human harvest. However, we
 171 considered the process here as a standard perturbation in all models, and then added specific
 172 types of additional perturbations per scenario (see demographic scenarios below).

173

174 *Abundance time series*

175 From the base models (parameterised to incorporate age structure, density feedbacks on
 176 survival, and catastrophic events in the Leslie matrices as described above), we generated
 177 multiannual abundance time series up to $40[G]$ for each species. We standardised projection
 178 length to $40[G]$ because there is strong evidence that the length of a time series (q) dictates
 179 the statistical power to detect an ensemble density-feedback signal in logistic growth curves
 180 (Brook & Bradshaw 2006). Here, we summed the \mathbf{n} vector over all age classes to produce a
 181 total population size $N_{t,i}$ for each year t of each iteration i . We rejected the first $[G]$ -
 182 equivalent years of each projection as a burn-in to allow the initial (deterministic) age
 183 distribution to calibrate to the stochastic expression of stability under compensatory density
 184 feedback.

185 To ascertain the degree of nonstationary in the simulated abundance time series, we used
 186 Turchin’s (2003) definition of nonstationarity as temporally variant mechanisms generating
 187 population fluctuations. In that conceptual context, we calculated the mean and variance of
 188 return time (T_R) — defined as the time required to return to equilibrium following a
 189 disturbance (Berryman 1999) — for each abundance time series as:

$$190 \quad \bar{T}_R = \frac{\sum_{m=1}^M T_{Rm}}{M} \quad [\text{eq 4}]$$

191 where \bar{T}_R is the mean T_R across M steps of the time series. For each m^{th} time step,

$$192 \quad T_{Rm} = S_{Cm} + S_{Fm} \quad [\text{eq 5}]$$

193 where: S_{Cm} is the number of complete time steps taken before reaching T_{Rm} , and S_{Fm} is the
 194 fraction of time required to reach T_{Rm} in the M^{th} (final) step:

$$195 \quad S_{Fm} = \frac{N_p - \bar{N}}{N_p - N_a} \quad [\text{eq 6}]$$

196 where \bar{N} is the mean of the abundance time series (a proxy for K), N_p is the population size
 197 prior to crossing \bar{N} , and N_a is the population size after crossing \bar{N} .

198 The variance of T_R is:

$$199 \quad \text{Var}(T_R) = \frac{\sum_{m=1}^M (T_{Rm} - \bar{T}_R)^2}{M-1} \quad [\text{eq 7}]$$

200 Thus, when $\bar{T}_R \ll \text{Var}(T_R)$ (i.e., $\bar{T}_R/\text{Var}(T_R) \ll 1$), the time series is considered to be highly
201 nonstationary (Berryman 1999).

202

203 *Demographic scenarios*

204 We generated 10,000 abundance time series over $40[G]$ for each test species in each of nine
205 demographic scenarios that combined different types and magnitudes of nonstationarity in the
206 form of density-independent (catastrophic and proportional) mortality and variation in
207 carrying capacity (K) through time. Each times series represented the idiosyncratic
208 demography of a unique population occupying an area of 250,000 km² with zero dispersal
209 (see above).

210 We split the nine scenarios into two main groups: **(1)** eight testing the probability of a
211 false negative (reduced detection of ensemble density feedback when a component feedback
212 on survival existed), and **(2)** one testing the probability of a false positive (evidence of
213 ensemble density feedback when a component feedback on survival was absent) (see details
214 in Table 2). The false-negative scenarios included three subcategories: **(1.1)** *i.* fixed K with
215 no perturbations other than the stochasticity imposed by resampling demographic rates in the
216 Leslie matrices; **(1.2)** fixed K with generationally scaled catastrophes centred on 50%
217 mortality *ii.* leading to $\bar{r} \cong 0$, *iii.* as in *ii.*, but with an additional, single ‘pulse’ perturbation of
218 90% mortality applied across the entire age structure at 20 generations, *iv.* a ‘harvest’-like
219 process where a consistent proportion of individuals is removed from the \mathbf{n} vector at each
220 time step to produce $\bar{r} \cong -0.001$ (i.e., weak, monotonic decline in average population size), or
221 *v.* as in *iv.*, but where the resultant $\bar{r} \cong -0.01$ (i.e., strong, monotonic decline in average
222 population size); and **(1.3)** K fluctuations with *vi.* stochastically resampled K with a constant
223 \bar{K} and a constant variance (via resampling the b parameter in equation [1]), *vii.* as in *vi.*, but
224 where the resampling variance doubles over the projection interval (via a linear increase in
225 the standard error used to resample the b parameter in equation [1]), and *viii.* as in *vi.*, but
226 where K declines at a rate of 0.001 over the projection interval (via decreasing the b
227 parameter in equation [1]). **2.** The false-positive scenario **2ix.** tested for false positives in the
228 ensemble signal by imposing a density-independent mortality via an increase in the
229 probability of catastrophe p_C in equation [3] to produce $\bar{r} \cong 0$ over $40[G]$. In that scenario,

230 we removed the component density-feedback on survival (i.e., setting $S_{\text{red}} = 1$) —
231 theoretically, populations lack a carrying capacity in the absence of density feedbacks.

232

233 ***Ensemble density feedbacks***

234 After generating 10,000 time series for each of the 21 species following the nine
235 demographic scenarios (totalling 189,000 individual time series), we applied
236 phenomenological models to each time series to test the statistical *evidence* for an ensemble
237 compensatory density feedback, as well as quantify the *strength* of that feedback. Our
238 phenomenological models included four variants of the general logistic growth curve
239 (Verhulst 1838) following Brook and Bradshaw (2006):

$$240 \quad r = \log_e \left(\frac{N_{t+1}}{N_t} \right) = \alpha + \beta N_t + \varepsilon_t \quad [\text{eq 8}]$$

241 where N_t = population size at time t , α = intercept, β = strength of ensemble density feedback,
242 and ε_t = Gaussian random variable with a mean of zero and a variance σ^2 reflecting
243 uncorrelated stochastic variability in the instantaneous rate of population change r . Our first
244 two models are simple density-independent models (DI): (1) random walk, where $\alpha = \beta = 0$,
245 and (2) exponential growth, where $\beta = 0$. The second two variants are density-dependent or
246 density-feedback models (DF): (3) Ricker-logistic (Ricker 1954), and (4) Gompertz-logistic
247 (Nelder 1961), where N_t on the right side of equation [8] is replaced with $\log_e(N_t)$. The latter
248 two models represent alternative situations where population growth rate varies in response
249 to unit (Ricker) or order-of-magnitude (Gompertz) changes in population size (Herrando-
250 Pérez *et al.* 2012b).

251 After fitting each of the four phenomenological models to each time series, we calculated
252 their relative likelihood by means of the Akaike's information criterion (AIC) corrected for
253 finite number of samples (AIC_c). We then expressed the *evidence* for an ensemble density-
254 feedback signal $\text{Pr}(\text{DF})$ as the sum of AIC_c weights ($w\text{AIC}_c$ = model probability) (Burnham
255 & Anderson 2002) for the Ricker- and Gompertz-logistic models (i.e., $\sum w\text{AIC}_c\text{-DF}$), and the
256 *evidence* for a lack of such signal as the sum of AIC_c weights for random walk and
257 exponential growth (i.e., $\sum w\text{AIC}_c\text{-DI}$). This follows the logic that if $\beta \neq 0$ between r and N_t
258 (Ricker) *or* $\log_e(N_t)$ (Gompertz) is more likely than $\beta = 0$ (random walk and exponential
259 growth), then there is stronger statistical support for an ensemble density feedback in the time
260 series than not (i.e., $\sum w\text{AIC}_c\text{-DF} > \sum w\text{AIC}_c\text{-DI}$ implies $\text{Pr}(\text{DF}) > 0.5$).

261 We estimated the *strength* of the ensemble density-feedback signal as the negative value
262 of $\hat{\beta}$ estimated from the Gompertz-logistic model. We used the Gompertz-logistic $\hat{\beta}$, instead

263 of the Ricker-logistic $\hat{\beta}$, to estimate this strength because only the former characterises the
264 multiplicative nature of demographic rates (Doncaster 2008; Herrando-Pérez *et al.* 2012a).
265 To compare the component density feedback applied to survival in the stochastic age-
266 structured models to the ensemble density feedback estimated from the abundance time series
267 under the nine demographic scenarios, we plotted the negative value of Gompertz $\hat{\beta}$ relative
268 to $1 - S_{\text{red}}$ across all 21 species modelled.

269 We tested the correlation between ensemble and component density-feedback strength,
270 and between ensemble strength and the degree of stationarity, across species by calculating a
271 bootstrapped estimate of Spearman's correlation ρ (treating relative differences in the metrics
272 as ranks). We uniformly resampled 10,000 times from the 95% confidence interval of each
273 metric for each species and demographic scenario, calculating ρ in turn, and then calculating
274 the median and 95% confidence interval of ρ . The relationships between ensemble and
275 component density-feedback strength (as well as between ensemble strength and stationarity)
276 showed some non-linearity, so we also fit simple exponential plateau models of the form $y =$
277 $y_{\text{max}} - (y_{\text{max}} - y_0)e^{-kx}$ to these relationships. Here, y_0 is the starting value of component strength,
278 y_{max} is the maximum component strength (- Gompertz $\hat{\beta}$), k = rate constant (in units of x^{-1}),
279 and x is the component strength ($1 - S_{\text{red}}$).

280

281 **RESULTS**

282 *Statistical evidence for density feedback*

283 For each test species, when the simulated populations were subjected to a compensatory
284 density feedback on survival (age-structured Leslie matrices), the median probability for a
285 statistical signal of ensemble compensatory density-feedback ($\text{Pr}(\text{DF}) = \Sigma w\text{AIC}_c\text{-DF}$; see
286 Materials and methods) across 10,000 times series of abundance was near unity (> 0.99) for
287 the stable ($\bar{r} \cong 0$) trajectories and most demographic scenarios (Fig. S1–S2 and S3 for
288 probability density plots of $\text{Pr}(\text{DF})$ across scenarios and the bootstrapped mean $\text{Pr}(\text{DF})$ per
289 species and scenario, respectively). Only the declining stochastic K scenario (1.3viii) had a
290 slightly smaller median $\text{Pr}(\text{DF})$ at 0.95. For the false-positive scenario (2ix), the median
291 $\text{Pr}(\text{DF})$ was 0.322. Generally, the extant dasyurid *S. harrissii* (SH; devil) and the flightless
292 bird *Dromaius novaehollandiae* (DN; emu) had the weakest evidence for density feedback
293 across the different scenarios (Fig. S3).

294 In summary, if a component density feedback was present, the phenomenological models
295 mostly detected the ensuing ensemble feedback (true positive) — regardless of whether a

296 simulated population was perturbed via density-independent removal of individuals, or
297 altered K dynamics — in > 9 of every 10 time series; while false positives (component
298 feedback absent, ensemble feedback detected) occurred in < 4 of every 10 times series.

299

300 *Degree of simulated stationarity*

301 The addition of the generationally scaled 50% catastrophic (density-independent) mortality
302 reduced stationarity from a median of $\bar{T}_R/\text{Var}(T_R) \sim 0.28$ (scenario 1.1*i*) to ~ 0.08 (scenario
303 1.2*ii*) (Fig. 1A). The scenarios imposing a catastrophic 90% mortality as a pulse at 20
304 generations (1.2*iii*), or additional proportional mortality driving a moderately (1.2*iv*; $\bar{r} = -$
305 0.01) or rapidly (1.2*v*; $\bar{r} = -0.001$) declining population over 40 generations, all reduced
306 stationarity by approximately the same amount relative to the scenario without catastrophic
307 mortality (1.1*i*) (Fig. 1C). For the scenarios emulating fluctuations in K (1.3*vi–viii*), adding
308 stochasticity to K slightly increased stationarity relative to a fixed- K scenario (Fig. 1E). Only
309 when the stochastic K was forced to decline (scenario 1.3*viii*), the abundance time series
310 became highly nonstationary (Fig. 1E). The false-positive scenario (2.*ix*) resulted in
311 negligible change to stationarity when comparing populations experiencing (Fig. 2A), or not
312 experiencing (Fig. 2B), a component density feedback on survival.

313

314 *Strength of density feedback*

315 While the magnitude of statistical evidence for density feedback was largely invariant across
316 all demographic scenarios including a component density feedback on survival (Fig. S1 and
317 S2; see above), the estimated strength of the ensemble density feedback (-Gompertz β , see
318 Materials and methods) was highly sensitive to the type of perturbation the population
319 experienced. The addition of the generationally scaled 50% catastrophic (density-
320 independent) mortality under a fixed K (scenarios 1.1*i* vs. 1.2*ii*) reduced the correlation
321 (median $\rho = 0.893$ and 0.881 , respectively) and slope between ensemble feedback strength
322 and component feedback strength ($1 - S_{\text{red}}$) across the 21 test species (Fig. 1B). The
323 catastrophic pulse scenario (1.2*iii*) returned the closest correlation (median $\rho = 0.929$)
324 between ensemble and component feedback strengths, although it also depressed the slope of
325 the relationship relative to the K_{fixed} scenario (Fig. 1D). These correlations were weakest for
326 the $\bar{r} = -0.001$ and $\bar{r} = -0.01$ scenarios (1.2*v–vi*; median $\rho = 0.009$ and -0.051 , respectively),
327 which also captured a signal of depensation (population growth rate increases with
328 population size) in some abundance time series (Fig. 1D). For the demographic scenarios

329 emulating fluctuations in K (1.3), the correlation between unit change in ensemble and
330 component density feedback strength was generally higher than those where $\bar{r} < 0$ (Fig. 1F;
331 median ρ ranging from 0.803 to 0.881), with the strongest mismatch occurring when K
332 declined by a rate of 0.001 (scenario 1.3viii) (Fig. 1F; see also Fig. S4). For the false-positive
333 scenario (2ix), all estimated ensemble feedback strengths enveloped 0 (Fig. 2B), meaning that
334 the estimated slopes of the $r \sim \log_e(N_t)$ relationships could not be differentiated from zero.

335 Overall, when an ensemble density feedback was detected from time series of abundance,
336 density-independent mortality eroded the extent by which true compensatory density
337 feedbacks on survival translated into an ensemble compensatory density feedback in
338 population trends more than fluctuations in K , with the most faulty outcome in fact inferring
339 compensatory population growth rates from some populations only experiencing density
340 compensation on survival.

341 On the other hand, the stationarity metric $\bar{T}_R/\text{Var}(T_R)$ was a weak (median $\rho =$
342 0.547, -0.086, and -0.113 for the pulse, $\bar{r} = -0.001$, and $\bar{r} = -0.01$ scenarios, respectively)
343 predictor of the estimated strength of ensemble feedback when density-independent mortality
344 was imposed (Fig. 3). However, stationarity was a reasonable (median $\rho = 0.756, 0.786,$ and
345 0.844 for the $K_{\text{stochastic}}, K_{\text{stochastic}}$ with increasing variance, and declining $K_{\text{stochastic}}$ scenarios,
346 respectively) predictor of the ensemble signal for the fluctuating K scenarios (Fig. 4; see also
347 Fig. S4).

348

349 **DISCUSSION**

350 Our simulations reveal several new insights into how ensemble (population growth rates) and
351 component (vital rates) density feedbacks can be decoupled. First, the statistical detection of
352 true ensemble feedback strength through phenomenological models is little affected by
353 nonstationarity *per se*. Second, the estimation of ensemble feedback strength through
354 phenomenological models (logistic growth curves; see Introduction) are particularly sensitive
355 to density-independent mortality leading to population decline, but they are less sensitive to
356 moderate fluctuations in carrying capacity. Third, the concern that density-independent
357 processes can invoke false evidence of ensemble signals of compensation are not borne out
358 by our simulations, at least with respect to density-independent mortality.

359 The mechanisms underlying those trends are nuanced by species' life histories. For
360 instance, in long-living terrestrial vertebrates (our focus), density feedbacks might operate on
361 fertility to compensate for pathogen-induced adult mortality (McDonald *et al.* 2016), those

362 feedbacks might be stronger on survival *versus* fertility when populations are near or far from
363 carrying capacity, respectively (Sæther *et al.* 2016), and survival can be entirely driven by
364 climatic conditions and density-independent predation (Hebblewhite *et al.* 2018). In one of
365 the best-studied systems in this regard, Soay sheep from St. Kilda Archipelago (United
366 Kingdom) demonstrate that the demographic role of density and weather varies across sexes
367 and age classes in mild winters, but survival is reduced consistently in all individuals in years
368 of bad weather and high population abundance (Coulson *et al.* 2001). Much less-studied than
369 herbivores, inter-pack aggression in carnivores with strong social hierarchies like wolves
370 might shape survival at high densities, but be demographically irrelevant at low densities
371 resulting from prey shortages and/or hunting or culling (Cubaynes *et al.* 2014). Our study
372 lends credence to the application of phenomenological models to the former types of studies
373 addressing the long-term effect of vital rates on population abundance, provided there is
374 enough information available for describing population trends.

375 Our approach and results do not, of course, explain all possible scenarios leading to the
376 decoupling of ensemble and component feedback signals. For example, many other density-
377 independent factors that we did not consider can dampen the demographic role of social and
378 trophic interactions mediated by population size (Herrando-Pérez *et al.* 2012a). Along with
379 the confounding effects of sampling error (Staples *et al.* 2004; Knape & de Valpine 2012),
380 some of those factors include immigration (Lieury *et al.* 2015), spatial heterogeneity in
381 population growth rates (Thorson *et al.* 2015; Johnson *et al.* 2016), fluctuating age structure
382 (Hoy *et al.* 2020), and environmental state shifts (Lande *et al.* 2002; Turchin 2003; Wu *et al.*
383 2007). Furthermore, our choice to limit the component mechanisms to feedback on a single
384 demographic rate (albeit, applied to all age classes) for the sake of simpler interpretation
385 could limit the application of our conclusions. For example, additional density-feedback
386 mechanisms operating independently on other demographic rates, such as fertility and
387 dispersal, could potentially complicate the interpretation derived from phenomenological
388 models.

389 Simulating closed populations potentially inflated the phenomenological model's capacity
390 to detect the component signal, because permanent dispersal could alleviate per capita
391 reductions in fitness as a population approaches carrying capacity. We also limited our
392 projections to a standardised 40 generations, but even expanding these to 120 generations
393 resulted in little change in the stationarity metric (Fig. S5). Complementary studies focussing
394 on the faster end of the life-history continuum could provide further insights, even though our
395 range of test species still precipitated a life-history signal in terms of component (Fig. S6)

396 and ensemble density-feedback strengths and stationarity (Fig. S7, S8) declining with
397 increasing generation length. However, this relationship faded when the trajectories
398 simulated declines through proportional removal. Indeed, both evidence for (Holyoak &
399 Baillie 1996) and strength of (Herrando-Pérez *et al.* 2012c) ensemble density feedback
400 generally increase along the continuum of slow to fast life histories, because species with
401 slow life histories are assumed to be more demographically stable when density
402 compensation is operating (Sæther *et al.* 2002).

403 While quantifying the true extent of all component density feedback mechanisms
404 operating in real populations will remain challenging in most circumstances,
405 phenomenological models can normally capture the evidence for and strength of the
406 component density feedback mechanism at play. Appreciating the degree of nonstationarity
407 and other types of perturbations affecting abundance time series can contextualise
408 interpretations of ensemble density-feedback signals, especially where substantial density-
409 independent mortality leads to long-term population declines. Importantly, failing to capture
410 density feedback in applied ecological models can lead to suboptimal conservation and
411 management recommendations and outcomes (Herrando-Pérez *et al.* 2012a; Horswill *et al.*
412 2017).

413

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419

420 **AUTHOR CONTRIBUTIONS**

421 CJAB conceived the idea, ran the simulations, and wrote the first draft. SHP reviewed the
422 literature. Both authors contributed to revisions.

423

424 **DATA AVAILABILITY STATEMENT**

425 All data files and R code are openly available at
426 <https://github.com/cjabradshaw/DensityFeedbackSims>.

427

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432 REFERENCES

- 433 Azerefegne, F., Solbreck, C. & Ives, A.R. (2001). Environmental forcing and high amplitude fluctuations in the
434 population dynamics of the tropical butterfly *Acraea acerata* (Lepidoptera: Nymphalidae). *Journal of*
435 *Animal Ecology*, 70, 1032-1045.
- 436 Battaile, B.C. & Trites, A.W. (2013). Linking reproduction and survival can improve model estimates of vital
437 rates derived from limited time-series counts of pinnipeds and other species. *PLoS One*, 8, e77389.
- 438 Bellier, E., Kéry, M. & Schaub, M. (2016). Simulation-based assessment of dynamic N-mixture models in the
439 presence of density dependence and environmental stochasticity. *Methods in Ecology and Evolution*, 7,
440 1029-1040.
- 441 Berryman, A. & Turchin, P. (2001). Identifying the density-dependent structure underlying ecological time
442 series. *Oikos*, 92, 265-270.
- 443 Berryman, A.A. (1999). *Principles of Population Dynamics and Their Application*. Stanley Thorners Ltd.,
444 Cheltenham, UK.
- 445 Bonsall, M.B. & Benmayor, R. (2005). Multiple infections alter density dependence in host-pathogen
446 interactions. *Journal of Animal Ecology*, 74, 937-945.
- 447 Bradshaw, C.J.A., Field, I.C., McMahon, C.R., Johnson, G.J., Meekan, M.G. & Buckworth, R.C. (2013). More
448 analytical bite in estimating targets for shark harvest. *Marine Ecology Progress Series*, 488, 221-232.
- 449 Bradshaw, C.J.A., Johnson, C.N., Llewelyn, J., Weisbecker, V., Strona, G. & Saltré, F. (2021). Relative
450 demographic susceptibility does not explain the extinction chronology of Sahul's megafauna. *eLife*, 10,
451 e63870.
- 452 Brook, B.W. & Bradshaw, C.J.A. (2006). Strength of evidence for density dependence in abundance time series
453 of 1198 species. *Ecology*, 87, 1445-1451.
- 454 Brook, B.W., Traill, L.W. & Bradshaw, C.J.A. (2006). Minimum viable population size and global extinction
455 risk are unrelated. *Ecology Letters*, 9, 375-382.
- 456 Bürgi, L.P., Roltsch, W.J. & Mills, N.J. (2015). Allee effects and population regulation: a test for biotic
457 resistance against an invasive leafroller by resident parasitoids. *Population Ecology*, 57, 215-225.
- 458 Burnham, K.P. & Anderson, D.R. (2002). *Model Selection and Multimodel Inference: A Practical Information-*
459 *Theoretic Approach*. 2nd edn. Springer-Verlag, New York, USA.
- 460 Caswell, H. (2001). *Matrix Population Models: Construction, Analysis, and Interpretation*, 2nd edn. Sinauer
461 Associates, Inc., Sunderland, USA.
- 462 Cockburn, A. (1997). Living slow and dying young: senescence in marsupials. In: *Marsupial Biology: Recent*
463 *Research, New Perspectives* (eds. Saunders, N & Hinds, L). University of New South Wales Press
464 Sydney, pp. 163-171.
- 465 Coulson, T., Catchpole, E.A., Albon, S.D., Morgan, B.J.T., Pemberton, J.M., Clutton-Brock, T.H. *et al.* (2001).
466 Age, sex, density, winter weather and population crashes in soay sheep. *Science*, 292, 1528-1531.
- 467 Cubaynes, S., Macnulty, D.R., Stahler, D.R., Quimby, K.A., Smith, D.W. & Coulson, T. (2014). Density-
468 dependent intraspecific aggression regulates survival in northern Yellowstone wolves (*Canis lupus*).
469 *Journal of Animal Ecology*, 83, 1344-1356.
- 470 Doncaster, C.P. (2008). Non-linear density dependence in time series is not evidence of non-logistic growth.
471 *Theoretical Population Biology*, 73, 483-489.
- 472 Doyle, S., Cabot, D., Walsh, A., Inger, R., Bearhop, S. & McMahon, B.J. (2020). Temperature and precipitation
473 at migratory grounds influence demographic trends of an Arctic-breeding bird. *Global Change Biology*,
474 26, 5447-5458.
- 475 Eberhardt, L.L. (2002). A paradigm for population analysis of long-lived vertebrates. *Ecology*, 83, 281-2854.
- 476 Eberhardt, L.L., Breiwick, J.M. & Demaster, D.P. (2008). Analyzing population growth curves. *Oikos*, 117,
477 1240-1246.
- 478 Fowler, C.W. (1981). Density dependence as related to life history strategy. *Ecology*, 62, 602-610.
- 479 Gaillard, J.M., Pontier, D., Allainé, D., Lebreton, J.D., Trouvilliez, J. & Clobert, J. (1989). An analysis of
480 demographic tactics in birds and mammals. *Oikos*, 56, 59-76.
- 481 Hebblewhite, M., Eacker, D.R., Eggeman, S., Bohm, H. & Merrill, E.H. (2018). Density-independent predation
482 affects migrants and residents equally in a declining partially migratory elk population. *Oikos*, 127, 1304-
483 1318.
- 484 Heppell, S.S., Caswell, H. & Crowder, L.B. (2000). Life histories and elasticity patterns: perturbation analysis
485 for species with minimal demographic data. *Ecology*, 81, 654-665.

486 Herrando-Pérez, S., Delean, S., Brook, B.W. & Bradshaw, C.J.A. (2012a). Decoupling of component and
487 ensemble density feedbacks in birds and mammals. *Ecology*, 93, 1728-1740.

488 Herrando-Pérez, S., Delean, S., Brook, B.W. & Bradshaw, C.J.A. (2012b). Density dependence: an ecological
489 Tower of Babel. *Oecologia*, 170, 585-603.

490 Herrando-Pérez, S., Delean, S., Brook, B.W. & Bradshaw, C.J.A. (2012c). Strength of density feedback in
491 census data increases from slow to fast life histories. *Ecology and Evolution*, 2, 1922-1934.

492 Holyoak, M. & Baillie, S.R. (1996). Factors influencing detection of density dependence in British birds. II.
493 Longevity and population variability. *Oecologia*, 108, 54-63.

494 Holz, P.H. & Little, P.B. (1995). Degenerative leukoencephalopathy and myelopathy in dasyurids. *Journal of*
495 *Wildlife Diseases*, 31, 509-513.

496 Horswill, C., O'Brien, S.H. & Robinson, R.A. (2017). Density dependence and marine bird populations: are
497 wind farm assessments precautionary? *Journal of Applied Ecology*, 54, 1406-1414.

498 Hoy, S.R., MacNulty, D.R., Smith, D.W., Stahler, D.R., Lambin, X., Peterson, R.O. *et al.* (2020). Fluctuations
499 in age structure and their variable influence on population growth. *Functional Ecology*, 34, 203-216.

500 Jepsen, J.U., Hagen, S.B., Karlsen, S.R. & Ims, R.A. (2009). Phase-dependent outbreak dynamics of geometrid
501 moth linked to host plant phenology. *Proceedings of the Royal Society B: Biological Sciences*, 276,
502 4119-4128.

503 Johnson, D.W., Freiwald, J. & Bernardi, G. (2016). Genetic diversity affects the strength of population
504 regulation in a marine fish. *Ecology*, 97, 627-639.

505 Knappe, J. & de Valpine, P. (2012). Are patterns of density dependence in the Global Population Dynamics
506 Database driven by uncertainty about population abundance? *Ecology Letters*, 15, 17-23.

507 Kolb, A., Dahlgren, J.P. & Ehrlén, J. (2010). Population size affects vital rates but not population growth rate of
508 a perennial plant. *Ecology*, 91, 3210-3217.

509 Lande, R., Engen, S., Sæther, B.-E., Filli, F., Matthysen, E. & Weimerskirch, H. (2002). Estimating density
510 dependence from population time series using demographic theory and life-history data. *American*
511 *Naturalist*, 159, 321-337.

512 Lieury, N., Ruetten, S., Devillard, S., Albaret, M., Drouyer, F., Baudoux, B. *et al.* (2015). Compensatory
513 immigration challenges predator control: an experimental evidence-based approach improves
514 management. *Journal of Wildlife Management*, 79, 425-434.

515 Ma, Z. (2021). A unified survival-analysis approach to insect population development and survival times.
516 *Scientific Reports*, 11.

517 Margalida, A., Jiménez, J., Martínez, J.M., Sesé, J.A., García-Ferré, D., Llamas, A. *et al.* (2020). An assessment
518 of population size and demographic drivers of the Bearded Vulture using integrated population models.
519 *Ecological Monographs*, 90.

520 Marini, G., Poletti, P., Giacobini, M., Pugliese, A., Merler, S. & Rosà, R. (2016). The role of climatic and
521 density dependent factors in shaping mosquito population dynamics: the case of *Culex pipiens* in
522 northwestern Italy. *PLoS One*, 11, e0154018.

523 Matthysen, E. (2005). Density-dependent dispersal in birds and mammals. *Ecography*, 28, 403-416.

524 Maud, J.L., Atkinson, A., Hirst, A.G., Lindeque, P.K., Widdicombe, C.E., Harmer, R.A. *et al.* (2015). How does
525 *Calanus helgolandicus* maintain its population in a variable environment? Analysis of a 25-year time
526 series from the English Channel. *Progress In Oceanography*, 137, 513-523.

527 McDonald, J.L., Bailey, T., Delahay, R.J., McDonald, R.A., Smith, G.C. & Hodgson, D.J. (2016). Demographic
528 buffering and compensatory recruitment promotes the persistence of disease in a wildlife population.
529 *Ecology Letters*, 19, 443-449.

530 McGeoch, M.A. & Price, P.W. (2005). Scale-dependent mechanisms in the population dynamics of an insect
531 herbivore. *Oecologia*, 144, 278-288.

532 Morrison, C.A., Butler, S.J., Robinson, R.A., Clark, J.A., Arizaga, J., Aunins, A. *et al.* (2021). Covariation in
533 population trends and demography reveals targets for conservation action. *Proceedings of the Royal*
534 *Society B: Biological Sciences*, 288.

535 Münster-Swendsen, M. & Berryman, A. (2005). Detecting the causes of population cycles by analysis of R-
536 functions: the spruce needle-miner, *Epinotia tedella*, and its parasitoids in Danish spruce plantations.
537 *Oikos*, 108, 495-502.

538 Nelder, J.A. (1961). The fitting of a generalization of the logistic curve. *Biometrics*, 17, 89-110.

539 Oakwood, M., Bradley, A.J. & Cockburn, A. (2001). Semelparity in a large marsupial. *Proceedings of the Royal*
540 *Society of London B: Biological Sciences*, 268, 407-411.

541 Oli, M.K. & Dobson, F.S. (2003). The relative importance of life-history variables to population growth rate in
542 mammals: Cole's predictions revisited. *American Naturalist*, 161, 422-440.

543 Owen-Smith, N. & Mason, D.R. (2005). Comparative changes in adult vs. juvenile survival affecting population
544 trends of African ungulates. *Journal of Animal Ecology*, 74, 762-773.

545 Paradis, E., Baillie, S.R., Sutherland, W.J. & Gregory, R.D. (2002). Exploring density-dependent relationships
546 in demographic parameters in populations of birds at a large spatial scale. *Oikos*, 97, 293-307.

547 Pardo, D., Forcada, J., Wood, A.G., Tuck, G.N., Ireland, L., Pradel, R. *et al.* (2017). Additive effects of climate
548 and fisheries drive ongoing declines in multiple albatross species. *Proceedings of the National Academy
549 of Sciences of the USA*, 114, E10829-E10837.

550 Reed, D.H., O'Grady, J.J., Ballou, J.D. & Frankham, R. (2003). The frequency and severity of catastrophic die-
551 offs in vertebrates. *Animal Conservation*, 6, 109-114.

552 Ricker, W.E. (1954). Stock and recruitment. *Journal of the Fisheries Research Board of Canada*, 11, 559-623.

553 Sæther, B.-E. & Bakke, Ø. (2000). Avian life history variation and contribution of demographic traits to the
554 population growth rate. *Ecology*, 81, 642-653.

555 Sæther, B.-E., Engen, S. & Matthysen, E. (2002). Demographic characteristics and population dynamical
556 patterns of solitary birds. *Science*, 295, 2070-2073.

557 Sæther, B.E., Grøtan, V., Engen, S., Coulson, T., Grant, P.R., Visser, M.E. *et al.* (2016). Demographic routes to
558 variability and regulation in bird populations. *Nature Communications*, 7, 12001.

559 Saunders, S.P., Cuthbert, F.J. & Zipkin, E.F. (2018). Evaluating population viability and efficacy of
560 conservation management using integrated population models. *Journal of Applied Ecology*, 55, 1380-
561 1392.

562 Staples, D.F., Taper, M.L. & Dennis, B. (2004). Estimating population trend and process variation for PVA in
563 the presence of sampling error. *Ecology*, 85, 923-929.

564 Stillman, R.A., Rivers, E.M., Gilkerson, W., Wood, K.A., Nolet, B.A., Clausen, P. *et al.* (2021). Predicting
565 impacts of food competition, climate, and disturbance on a long-distance migratory herbivore.
566 *Ecosphere*, 12.

567 Thorson, J.T., Skaug, H.J., Kristensen, K., Shelton, A.O., Ward, E.J., Harms, J.H. *et al.* (2015). The importance
568 of spatial models for estimating the strength of density dependence. *Ecology*, 96, 1202-1212.

569 Traill, L.W., Brook, B.W., Frankham, R. & Bradshaw, C.J.A. (2010). Pragmatic population viability targets in a
570 rapidly changing world. *Biological Conservation*, 143, 28-34.

571 Turchin, P. (2003). *Complex Population Dynamics: A Theoretical/Empirical Synthesis*. Princeton University
572 Press, Princeton, USA.

573 Verhulst, P.F. (1838). Notice sur la loi que la population poursuit dans son accroissement. *Correspondance
574 mathématique et physique*, 10, 113-121.

575 Wu, Z., Huang, N.E., Long, S.R. & Peng, C.-K. (2007). On the trend, detrending, and variability of nonlinear
576 and nonstationary time series. *Proceedings of the National Academy of Sciences of the USA*, 104, 14889.

577

TABLE 1 Taxonomy and life-history characteristics of the 21 test species (all native to Australia) used to simulate age-structured populations and time series of population abundance. *abb* = abbreviation of scientific name, *M* = body mass (kg), *GL* = generation length (years), *q* = projection length (years) (Bradshaw *et al.* 2021).

taxonomic/functional group	species	abb	<i>M</i>	<i>GL</i>	<i>q</i>	status
herbivore vombatiformes	<i>Diprotodon optatum</i>	DP	2786	18.1	724	extinct
	<i>Palorchestes azael</i>	PA	1000	15.1	604	extinct
	<i>Zygomaturus trilobus</i>	ZT	500	13.2	528	extinct
	<i>Phascolonus gigas</i>	PH	200	10.7	428	extinct
	<i>Vombatus ursinus</i>	VU	25	10.0	400	extant
herbivore macropodiformes	<i>Procoptodon goliath</i>	PG	250	8.3	332	extinct
	<i>Sthenurus stirlingi</i>	SS	150	8.1	324	extinct
	<i>Protemnodon anak</i>	PT	130	7.8	312	extinct
	<i>Simosthenurus occidentalis</i>	SO	120	7.8	312	extinct
	<i>Metasthenurus newtonae</i>	MN	55	6.0	240	extinct
	<i>Osphranter rufus</i>	OR	25	5.5	220	extant
	<i>Notamacropus rufogriseus</i>	NR	14	6.3	252	extant
large omnivore birds	<i>Genyornis newtoni</i>	GN	200	20.0	800	extinct
	<i>Dromaius novaehollandiae</i>	DN	55	5.9	236	extant
	<i>Alectura lathami</i>	AL	2.2	6.8	272	extant
carnivores	<i>Thylacoleo carnifex</i>	TC	110	9.1	364	extinct
	<i>Thylacinus cynocephalus</i>	TH	20	5.2	208	extinct
	<i>Sarcophilus harrisii</i>	SH	6.1	3.1	124	extant*
	<i>Dasyurus maculatus</i>	DM	2	2.3	92	extant
invertivore monotremes	<i>Megalibgwilia ramsayi</i>	MR	11	16.4	656	extant
	<i>Tachyglossus aculeatus</i>	TA	4	14.1	564	extant

* extant in Tasmania, currently extinct in mainland Australia

TABLE 2 Demographic scenarios to quantify the detection of ensemble density-feedback signals in time series of abundance using phenomenological models (logistic growth curves) if a component density feedback on survival is present (1. H_0 : false negatives), or absent (2. H_0 : false positives). All scenarios were simulated over 40 generations across 21 vertebrate species. Time series obtained from simulated age-structured populations (Leslie matrices) occupying 250,000 km² with no dispersal. G = generation, N = population abundance, K = carrying capacity; \bar{r} = long-term mean instantaneous rate of population change, SD = standard deviation. See test species in Table 1.

scenario	catastrophe type	description
1. H_0: false negatives (component feedback)		
<i>1.1 no catastrophic mortality or fluctuation in K</i>		
<i>i.</i> $K_{\text{fixed}}, \bar{r} \cong 0$	none	stochastically resampled survival rates in age-structured population
<i>1.2 catastrophic mortality (50%) and stable K</i>		
<i>ii.</i> $K_{\text{fixed}}, \bar{r} \cong 0$; sustained catastrophic mortality	generationally scaled	as in <i>i</i> , but with catastrophes
<i>iii.</i> $K_{\text{fixed}}, \bar{r} \cong 0$; pulsed catastrophic mortality	generationally scaled	as in <i>ii</i> , but with a single 90% mortality pulse implemented at $20G$
<i>iv.</i> $K_{\text{fixed}}, \bar{r} \cong -0.001$; sustained proportional mortality	generationally scaled	as in <i>ii</i> , but with proportional removal of individuals from the \mathbf{n} vector such that $\bar{r} = -0.001$ (slowly declining population)
<i>v.</i> $K_{\text{fixed}}, \bar{r} \cong -0.01$; sustained proportional mortality	generationally scaled	as in <i>iv</i> , but where $\bar{r} = -0.01$ (rapidly declining population)
<i>1.3 catastrophic mortality (50%) and fluctuation in K</i>		
<i>vi.</i> $K_{\text{stochastic}}, \bar{r} \cong 0$	generationally scaled	as in <i>ii</i> , but normally distributed K varying randomly at each time step (SD = 5%)
<i>vii.</i> $K_{\text{stochastic}}$ with increasing variance; $\bar{r} \cong 0$	generationally scaled	as in <i>vi</i> , but variance in K increased linearly from 5% to 10%
<i>viii.</i> $K_{\text{stochastic}}$ declining, forcing $\bar{r} < 0$	generationally scaled	as in <i>vi</i> , but K also decreases on average at a rate of -0.001
2. H_0: false positives (no component feedback)		
<i>ix.</i> no K ; $\bar{r} \cong 0$	temporally scaled	probability of catastrophe increased over time such that $\bar{r} \cong 0$ (~ average stability)

FIGURE 1 (A, C, E) Truncated violin plots showing the distribution of the stationarity index $\bar{T}_R / \text{Var}(T_R)$ across 10,000 times series of population abundance per species and all 21 test species (see list in Table 1) obtained from age-structured populations subjected to a compensatory component density feedback on survival over 40 generations, according to nine demographic scenarios (detailed in Table 2). (B, D, F) Relationship between strength of ensemble (slope coefficient β of the Gompertz-logistic model $\times [-1]$) and component (1 – the modifier S_{red} on survival) density feedback. (A–B) Scenarios without (blue: scenario 1.1i) and with (grey: scenario 1.2ii) generationally scaled 50% catastrophic (density-independent) mortality. (C–D) Stable projections with carrying capacity (K) fixed (darker grey; scenario 1.2ii), a pulse disturbance of 90% mortality at the first 20 generations (20G; lighter grey; scenario 1.2iii), weakly declining ($r \cong -0.001$; red; scenario 1.2iv), and strongly declining ($r \cong 0.01$; blue; scenario 1.2v). (E–F) Stable projections with K fixed (darker grey; scenario 1.2ii), varying stochastically (K_{stoch}) around a constant mean with a constant variance (lighter grey; scenario 1.3vi), varying stochastically with a constant mean and an increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; red; scenario 1.3vii), and varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; blue; scenario 1.3viii). The fitted curves across species are exponential plateau models of the form $y = y_{\text{max}} - (y_{\text{max}} - y_0)e^{-kx}$. Shaded regions represent the 95% prediction intervals for each type. Also shown are the mean probabilities of median density feedback (Pr(DF): sum of the Akaike’s information criterion weights for the Ricker- and Gompertz-logistic models across time series ($\Sigma w\text{AIC}_c\text{-DF}$). Compensation implies that survival and population growth wane as population abundance rises, and $\bar{T}_R \gg \text{Var}(T_R)$ implies high stationarity.

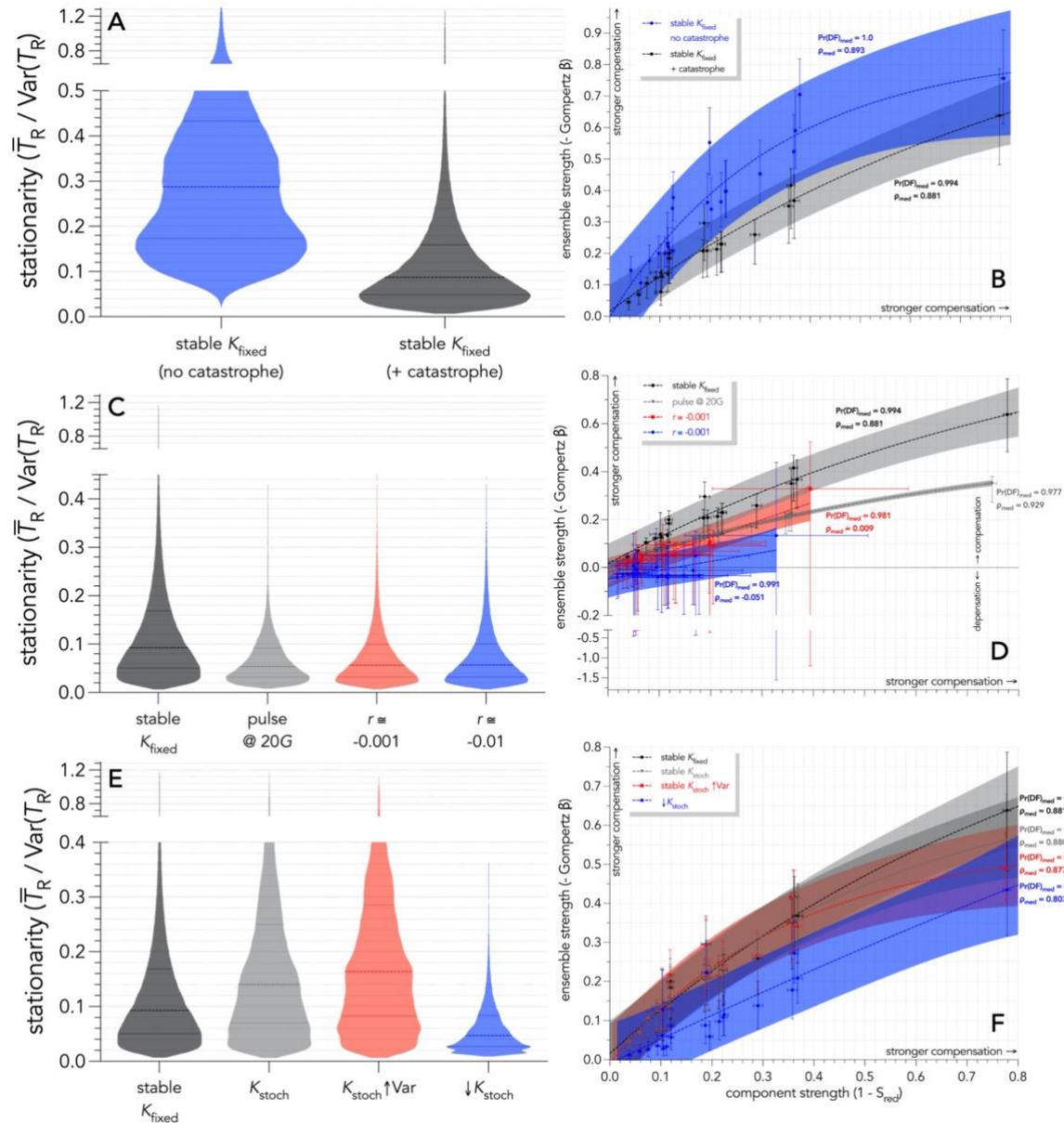


FIGURE 2 (A) Truncated violin plots showing the distribution of the stationarity index $\bar{T}_R/\text{Var}(T_R)$ across 10,000 times series of population abundance per species and all 21 species (see species list in Table 1) obtained from age-structured populations subjected to a compensatory component density feedback on survival over 40 generations, according to two demographic scenarios (detailed in Table 2). Demographic scenarios include carrying capacity (K) fixed with (darker grey, scenario 1.2*ii*) and without (lighter grey, scenario 2*ix*) component compensatory density-feedback on survival, the latter including an increase in the probability of 50% catastrophic (density-independent) mortality to produce stable population growth rates around 0 (see scenarios in Table 2). (B) Relationship between strength of ensemble (slope coefficient $\beta \times [-1]$) of the Gompertz-logistic model) and generation length (years) across the 21 species. Probabilities of density feedback (Pr(DF) = sum of the Akaike's information criterion weights for the Ricker and Gompertz models) calculated across simulations gave median Pr(DF) = 0.994 and 0.322 for the two stable scenarios without and with component feedback on survival, respectively.

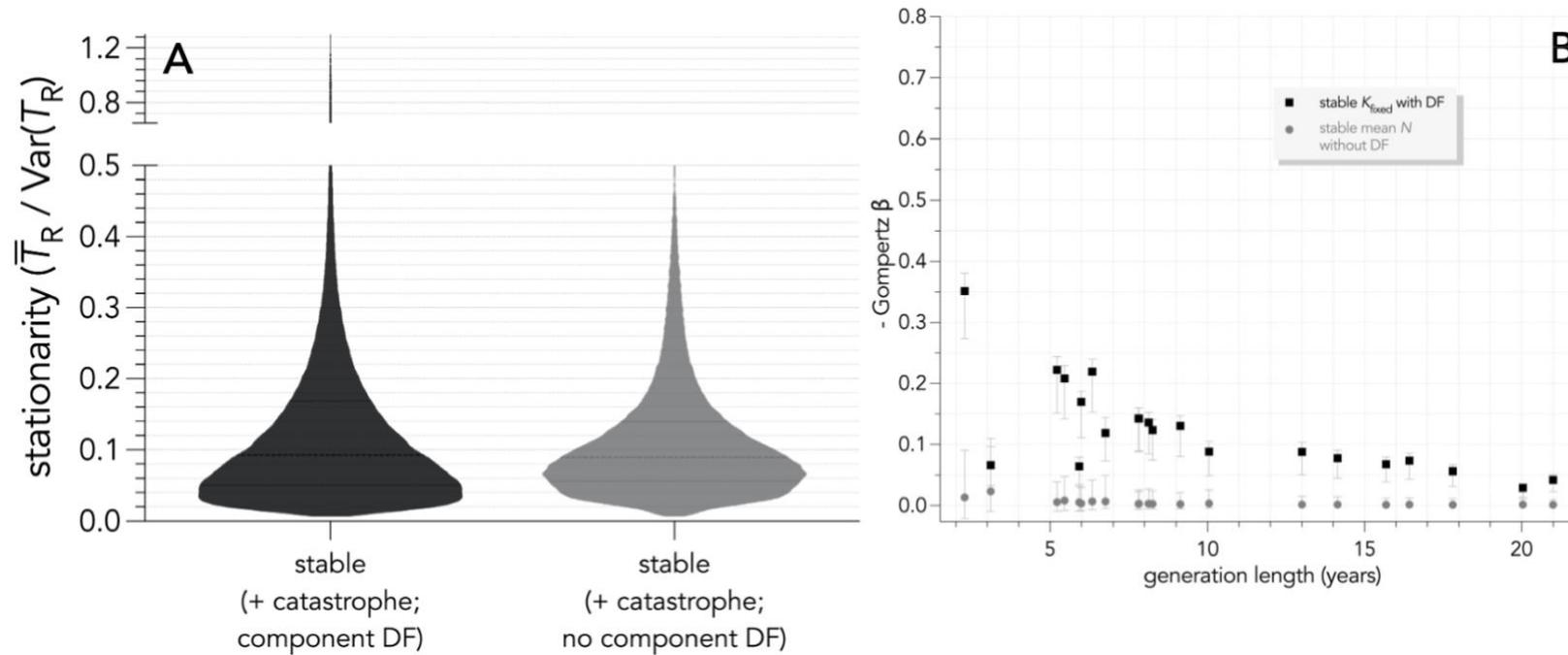


FIGURE 3 Relationships between the stationarity index $\bar{T}_R/\text{Var}(T_R)$ and the strength of ensemble density feedback (slope coefficient $\beta \times [-1]$ of the Gompertz-logistic model) for four scenarios with 50% catastrophic (density-independent) mortality across 21 test species (see Table 1) over 40 generations, including (A) carrying capacity (K) fixed (scenario 1.2ii), (B) a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2iii), (C) weakly declining ($r \cong -0.001$, scenario 1.2iv), and (D) strongly declining ($r \cong 0.01$, scenario 1.2v) populations (scenarios detailed in Table 2). The fitted curves across species exponential plateau models of the form $y = y_{\max} - (y_{\max} - y_0)e^{-\lambda x}$. Shaded regions represent the 95% prediction intervals for each type. ρ_{med} are the median Spearman's ρ correlation coefficients for the relationship between the ensemble strength and stationarity index across species (resampled 10,000 times; see Fig. S4 for full uncertainty range of ρ in each scenario).

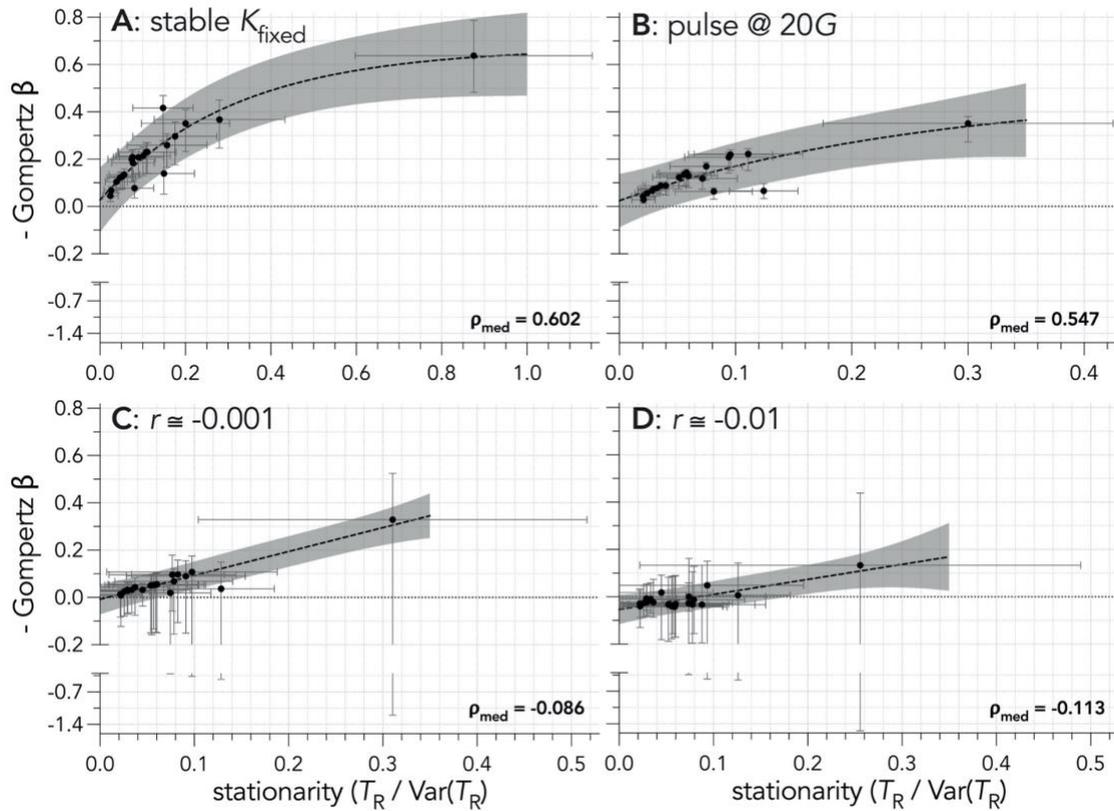
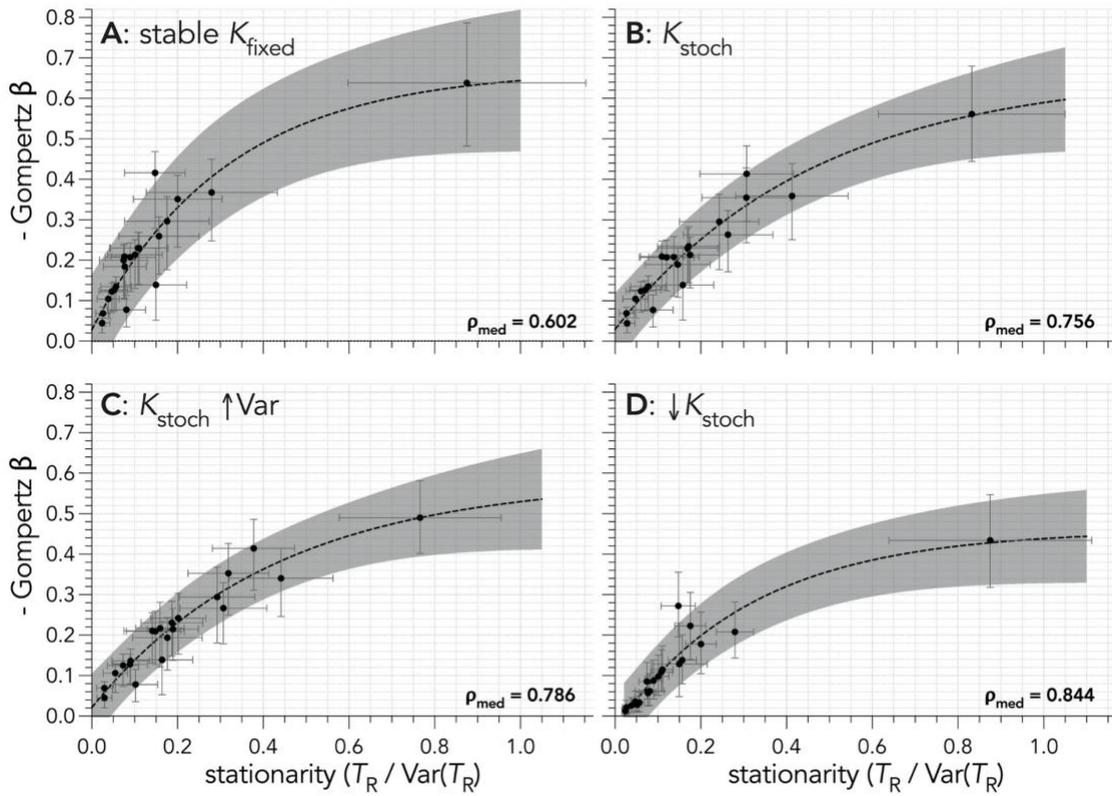


FIGURE 4 Relationships between the stationarity index $\bar{T}_R/\text{Var}(T_R)$ and the strength of ensemble density feedback (slope coefficient $\beta \times [-1]$ of the Gompertz-logistic model) across 21 test species (see list in Table 1) over 40 generations for four scenarios (scenarios detailed in Table 2) with 50% catastrophic (density-independent) mortality, including (A) carrying capacity (K) fixed (scenario 1.2ii), (B) K varying stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3vi), (C) K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$, scenario 1.3vii), and (D) K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$, scenario 1.3viii). The fitted curves across species exponential plateau models of the form $y = y_{\text{max}} - (y_{\text{max}} - y_0)e^{-kx}$. Shaded regions represent the 95% prediction intervals for each type. ρ_{med} are the median Spearman's ρ correlation coefficients for the relationship between the ensemble strength and stationarity index across species (resampled 10,000 times; see Fig. S4 for full uncertainty range under each scenario).



SUPPORTING INFORMATION

FIGURE S1 Probability of an ensemble compensatory density-feedback signal ($\text{Pr}(\text{DF}) = \sum w_{\text{AIC}_c\text{-DF}}$ = sum of Akaike's information criterion weights across the Ricker- and Gompertz-logistic models — see Materials and methods) in abundance time series for simulated populations of 21 long-lived species of Australian mammals and birds (see list in Table 1) subjected to compensatory density feedback on survival and experiencing 50 % catastrophic (density-independent) mortality over 40 generations. Each probability surface represents one of the 21 test species (see list in Table 1), so plots show the overlapping median probability density over 10,000 times series of abundance per species and for each of four demographic scenarios (detailed in Table 2), including (A) a carrying capacity is fixed (K_{fixed}) with 50 % catastrophic (density-independent) mortality (scenario 1.2ii), (B) a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2iii), and (C) weakly declining ($\bar{r} \cong -0.001$; scenario 1.2iv) and (D) strongly declining ($\bar{r} \cong -0.01$; scenario 1.2v) populations. See Fig. S3 for bootstrapped mean Spearman correlation coefficients for each scenario.

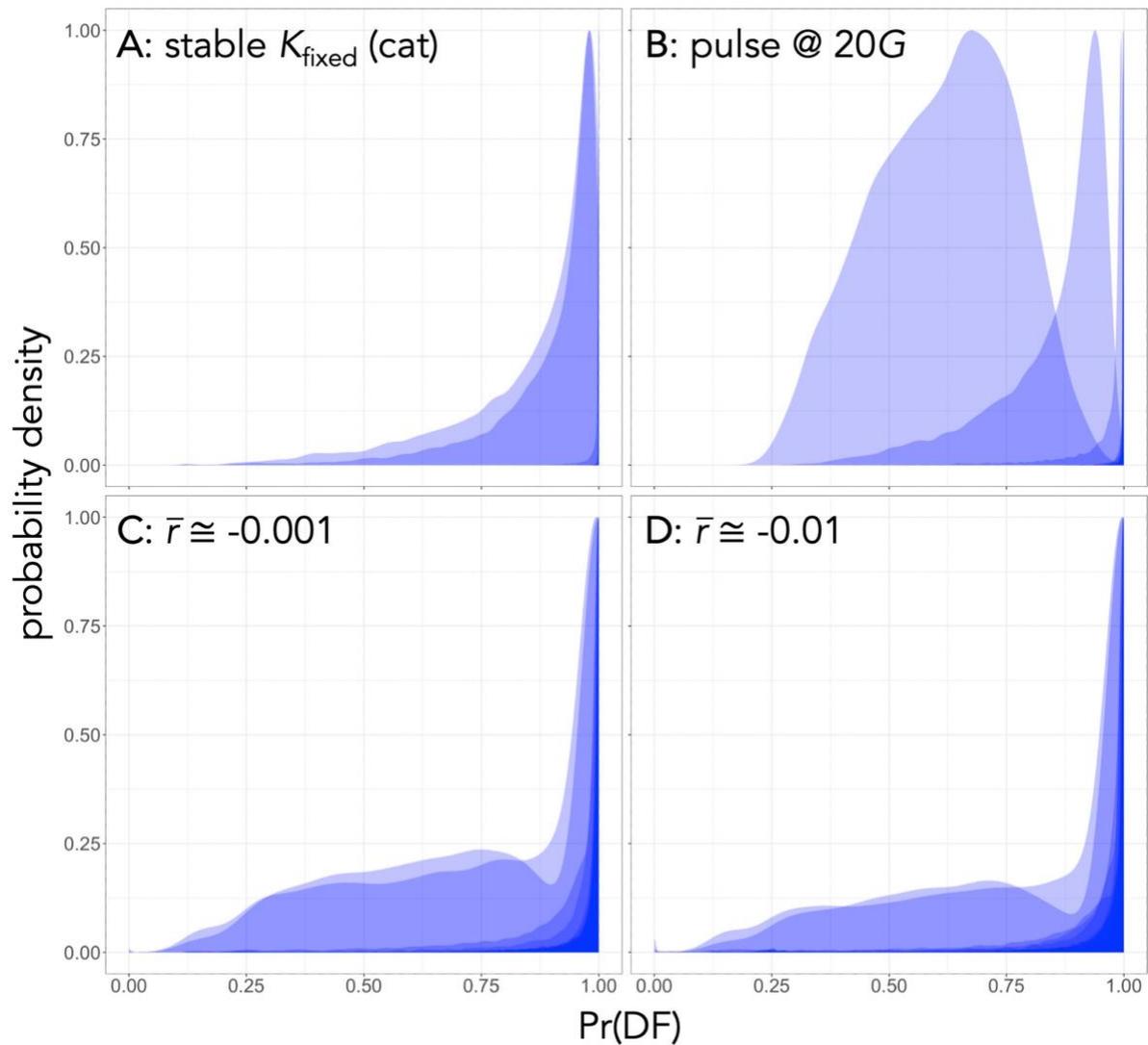


FIGURE S2 Probability of an ensemble compensatory density-feedback signal ($\text{Pr}(\text{DF}) = \sum w \text{AIC}_c\text{-DF}$ = sum of Akaike's information criterion weights across the Ricker- and Gompertz-logistic models — see Materials and methods) in abundance time series for simulated populations of 21 long-lived species of Australian mammals and birds (see list in Table 1) subjected to compensatory density feedback on survival and experiencing fluctuations in carrying capacity (K) along with 50 % catastrophic (density-independent) mortality over 40 generations. Each probability surface represents one of the 21 test species (see list in Table 1), so plots show the overlapping median probability density over 10,000 times series of abundance per species and for each of four demographic scenarios (detailed in Table 2), including (A) a stable demographic projection where K is fixed (K_{fixed}) (scenario 1.2ii), (B) K varies stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3vi), (C) K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; scenario 1.3vii), and (D) K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; scenario 1.3viii). See Fig. S3 for bootstrapped mean Spearman correlation coefficients for each scenario.

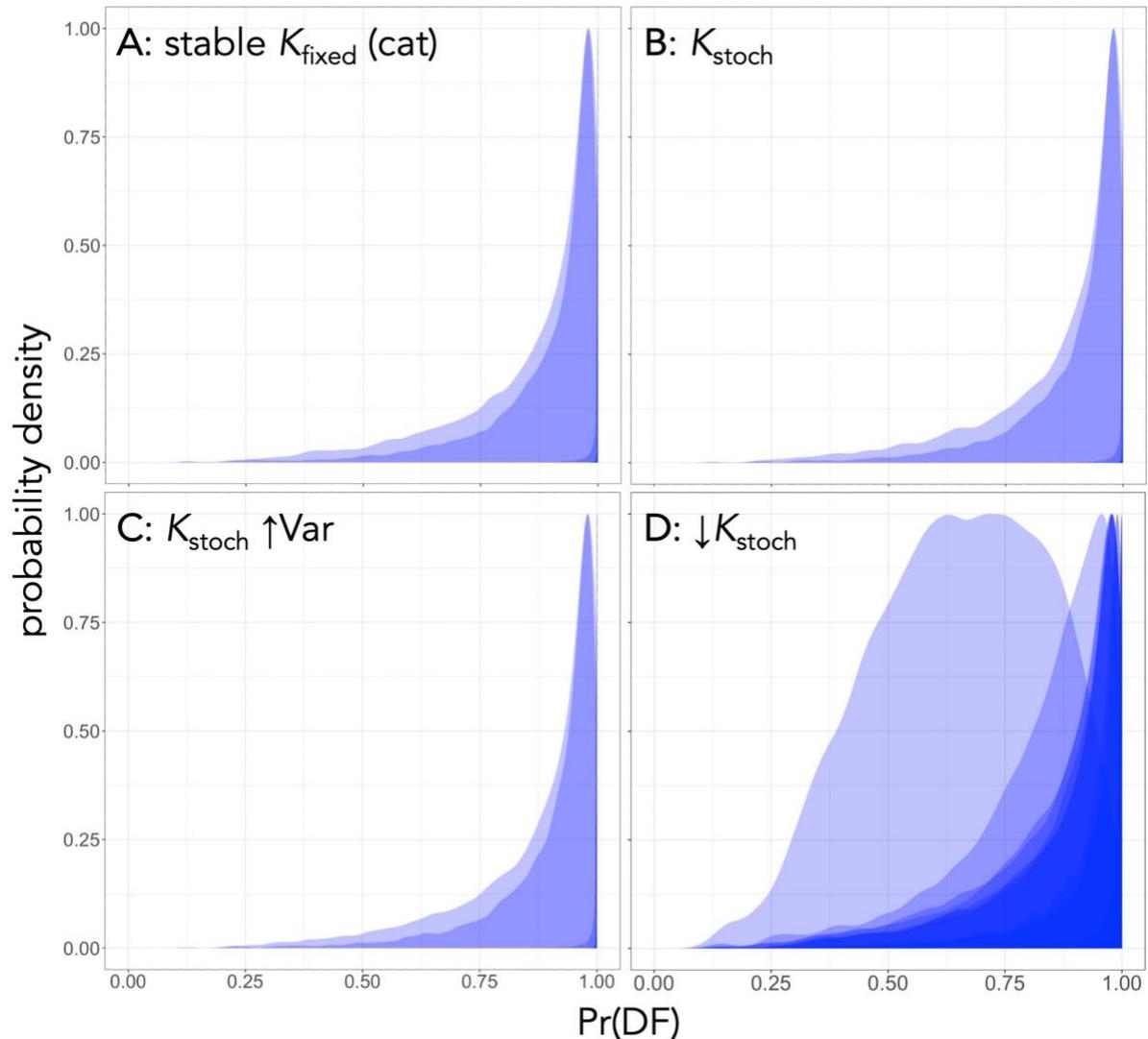


FIGURE S3 Bootstrapped mean (with 80 % confidence intervals; 100,000 resamples) probability of an ensemble compensatory density-feedback signal ($\text{Pr}(\text{DF}) = \Sigma w\text{AIC}_c\text{-DF}$ = sum of Akaike’s information criterion weights across the Ricker- and Gompertz-logistic models — see Materials and methods) in abundance time series for simulated populations of 21 long-lived species of Australian mammals and birds for populations (see list in Table 1) subjected to compensatory density feedback on survival and experiencing fluctuations in carrying capacity (K) and/or 50 % catastrophic (density-independent) mortality (scenarios detailed in Table 2). Demographic scenarios (see details in Table 2) include (A) K fixed (K_{fixed}) with no catastrophic mortality (no cat; scenario 1.1*i*), and with catastrophic mortality in combination with (B) K_{fixed} (cat; scenario 1.2*ii*), (C) a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2*iii*), (D) weakly declining ($\bar{r} \cong -0.001$; scenario 1.2*iv*) and (E) strongly declining ($\bar{r} \cong 0.01$; scenario 1.2*v*) populations, (F) K varying stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3*vi*), (G) K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; scenario 1.3*vii*), and (H) K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; scenario 1.3*viii*). The vertical dashed line at $\text{Pr}(\text{DF}) = 0.5$ in each panel is the point below which the evidence for a density-independent model [$\text{Pr}(\text{DI}) = \Sigma w\text{AIC}_c\text{-DI}$ = sum of Akaike’s information criterion weights across the random walk and exponential models] is greater than $\text{Pr}(\text{DF})$. See Table 2 for species abbreviations.

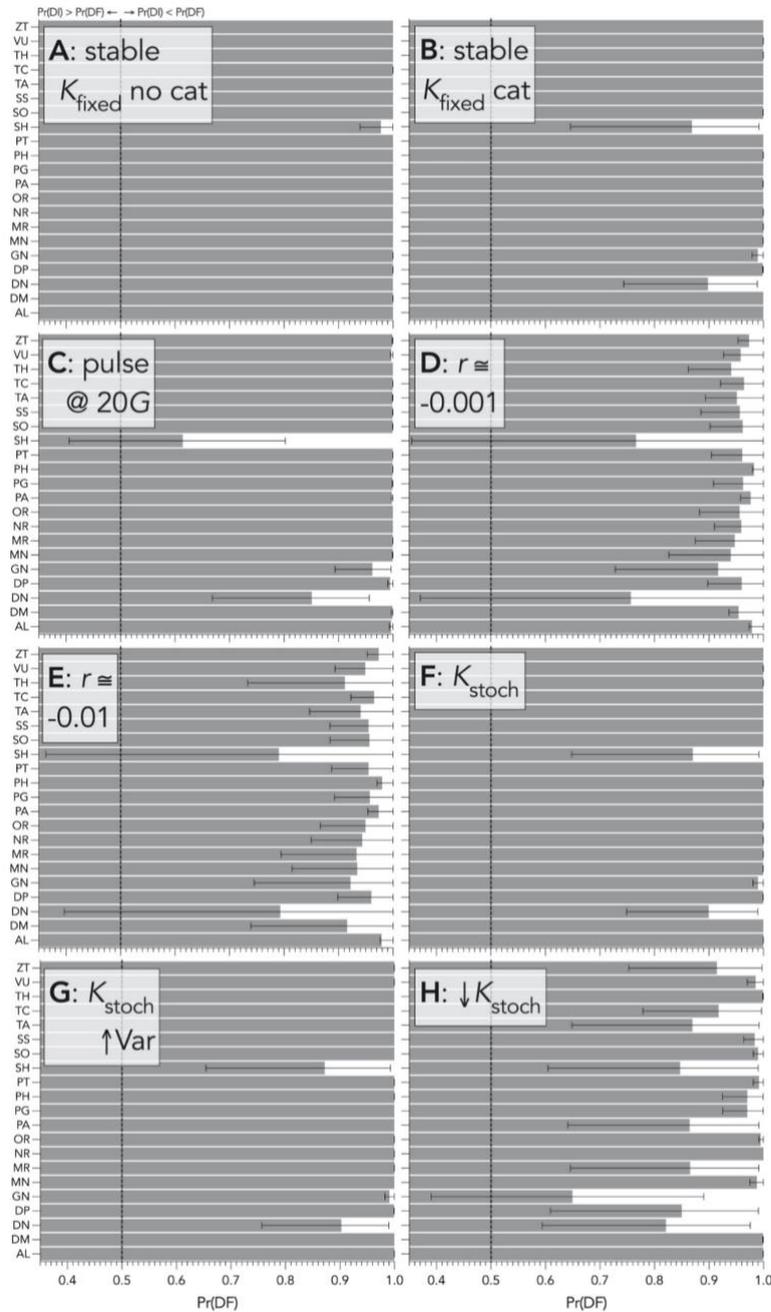


FIGURE S4 Bootstrapped (10,000 iterations) Spearman’s correlation ρ between (A) ensemble density feedback strength (- Gompertz slope β , the reduction of survival as population density increases) and component feedback strength on survival ($1 - S_{\text{red}}$, the reduction in survival as population density increases), and (B) ensemble feedback strength and the stationarity metric $\bar{T}_R/\text{Var}(T_R)$ for 10,000 simulated populations across each of 21 long-lived species of Australian mammals and birds for populations (see list in Table 1) subjected to compensatory density feedback on survival and experiencing fluctuations in carrying capacity (K) and/or 50 % catastrophic (density-independent) mortality (scenarios detailed in Table 2) . Demographic scenarios include K fixed (K_{fixed}) with no catastrophic mortality (no cat; scenario 1.1*i*), and catastrophic mortality in combination with K_{fixed} (cat; scenario 1.2*ii*), a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2*iii*), weakly declining ($\bar{r} \cong -0.001$; scenario 1.2*iv*) and (E) strongly declining ($\bar{r} \cong 0.01$; scenario 1.2*v*) populations, K varying stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3*vi*), K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; scenario 1.3*vii*), and K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; scenario 1.3*viii*).

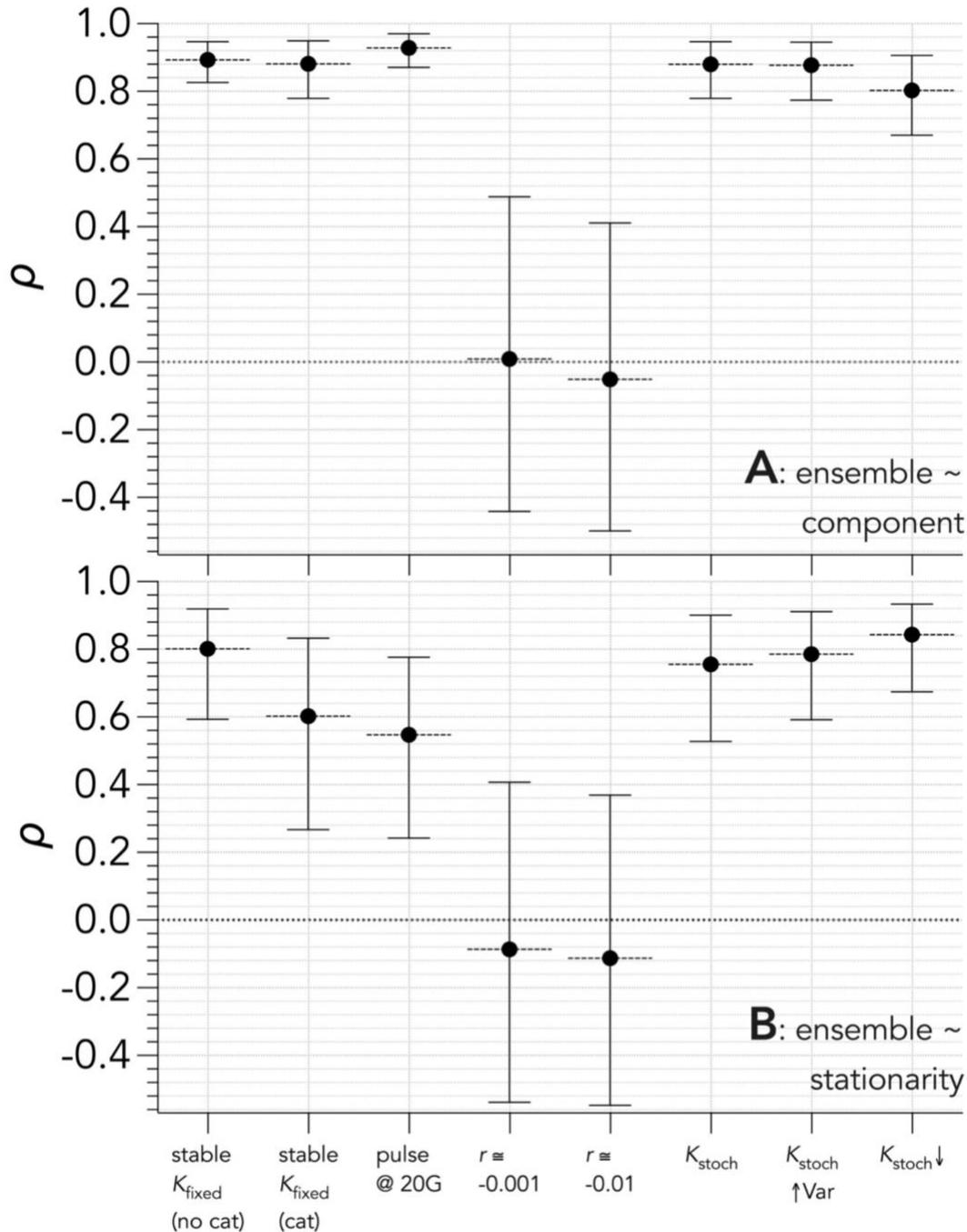


FIGURE S5 Truncated violin plots showing the distribution of the stationarity index $\bar{T}_R/\text{Var}(T_R)$ across 10,000 time series of population abundance per species and all 21 species (see species list in Table 1) obtained from age-structured populations for scenarios showing carrying capacity fixed with component compensatory density-feedback on survival and 50% catastrophic (density-independent) mortality to produce stable population growth rates around 0 over 40 (scenario 1.2ii; detailed in Table 2) and 120 generations (G).

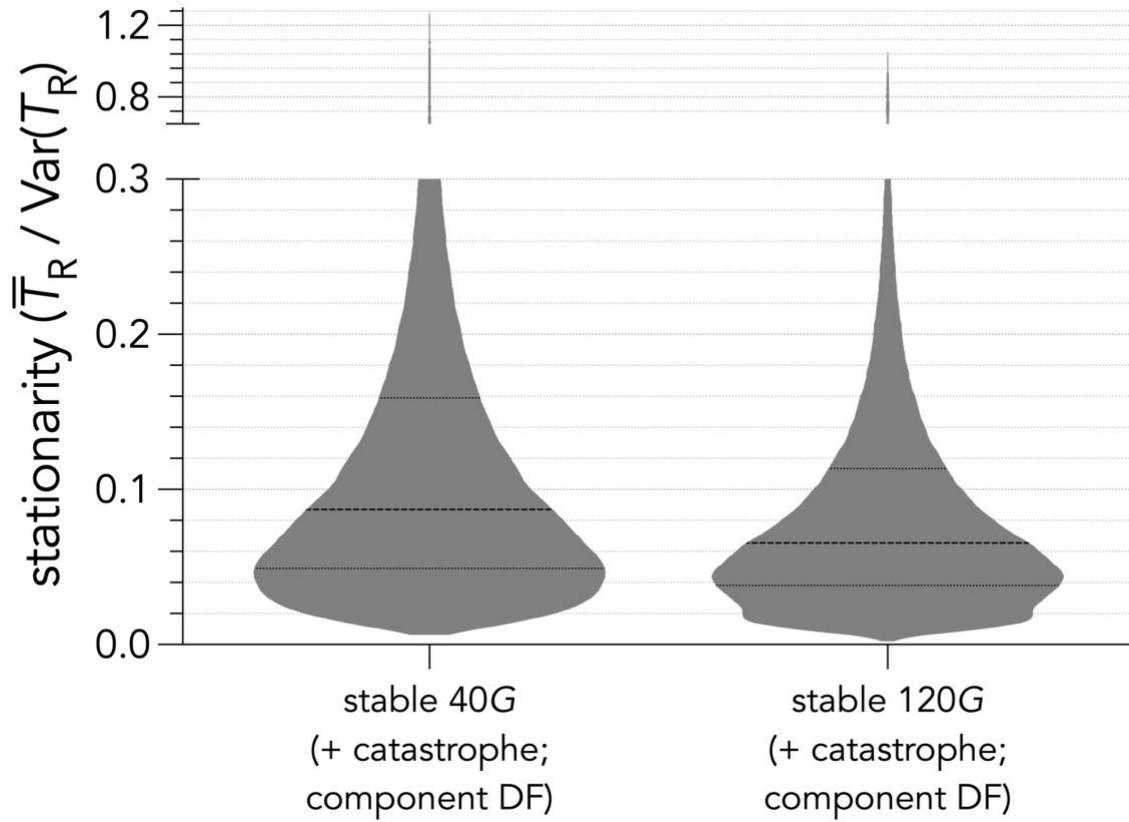


Fig. S6. Relationship between strength of component density feedback and generation length (years) across 10,000 time series of population abundance for each of 21 test species (see list in Table 1) obtained from age-structured populations subjected to a compensatory component density feedback on survival over 40 generations for a demographic scenario with constant carrying capacity and no catastrophic (density-independent) mortality (scenario 1.1i; detailed in Table 2). The dashed grey line indicates a least-squares-fitted (adjusted coefficient of regression $R^2 = 0.58$) exponential plateau model of the form: $y = y_{\max} - (y_{\max} - y_0)e^{-kG}$, where y_0 = starting value of component strength, y_{\max} = maximum component strength, k = rate constant (years^{-1}) and G = generation time (years). Species notation: DP = *Diprotodon optatum*, PA = *Palorchestes azael*, ZT = *Zygomaturus trilobus*, PH = *Phascolonus gigas*, VU *Vombatus ursinus* (herbivore vombatiform); PG = *Procoptodon goliath*, SS = *Sthenurus stirlingi*, PT = *Protemnodon anak*, SO = *Simosthenurus occidentalis*, MN = *Metasthenurus newtonae*, OR = *Osphranter rufus* (herbivore macropodiformes); GN = *Genyornis newtoni*, DN = *Dromaius novaehollandiae* (large omnivore birds), AL = *Alectura lathami*; TC = *Thylacoleo carnifex*, TH = *Thylacinus cynocephalus*, SH = *Sarcophilus harrisii* (carnivores), DM = *Dasyurus maculatus*; MR = *Megalibgwilia ramsayi*; TA = *Tachyglossus aculeatus* (invertebrate monotremes).

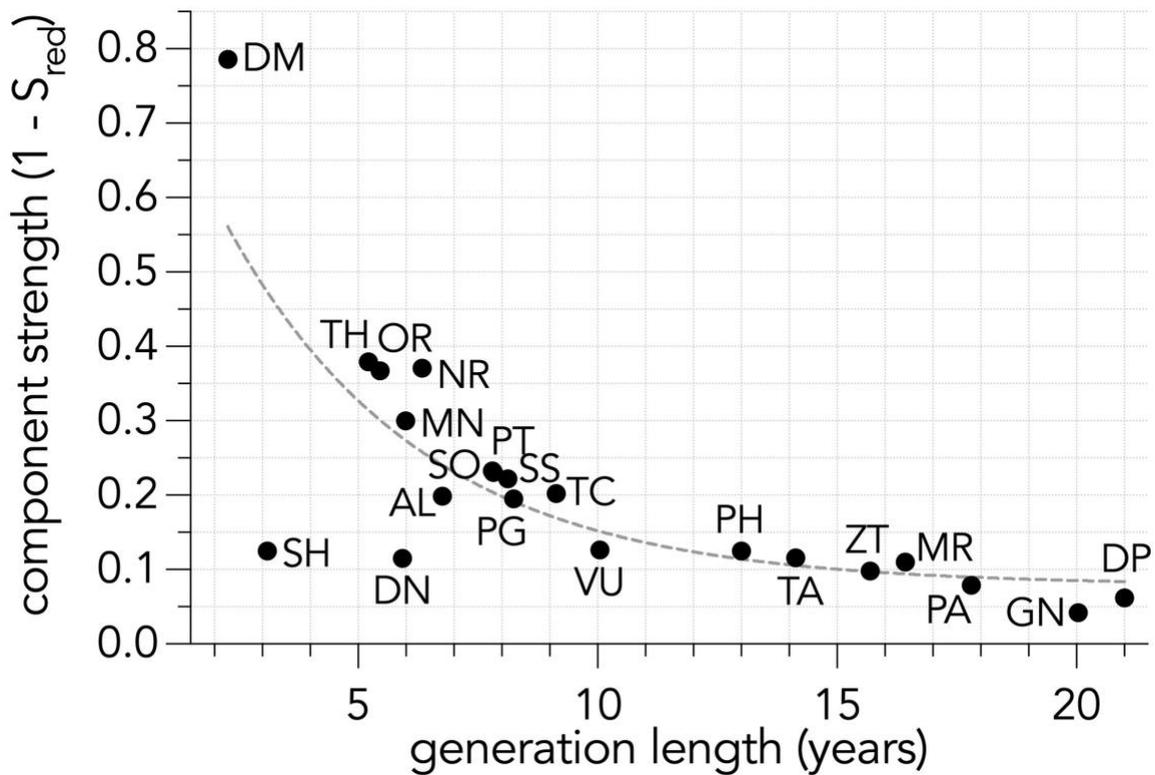


FIGURE S7 Relationships between the stationarity index $\bar{T}_R/\text{Var}(T_R)$ and generation length across 10,000 times series of population abundance per species and all 21 test species (see list in Table 1) obtained from age-structured populations subjected to a compensatory component density feedback on survival over 40 generations, according to seven demographic scenarios (detailed in Table 2). Demographic scenarios include (A) carrying capacity K fixed (K_{fixed} ; scenario 1.2ii), (B) a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2iii), (C) weakly declining ($\bar{r} \cong -0.001$; scenario 1.2iv) and (D) strongly declining ($\bar{r} \cong 0.01$; scenario 1.2v) populations, (E) K varying stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3vi), (F) K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; scenario 1.3vii), and (G) K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; scenario 1.3viii).

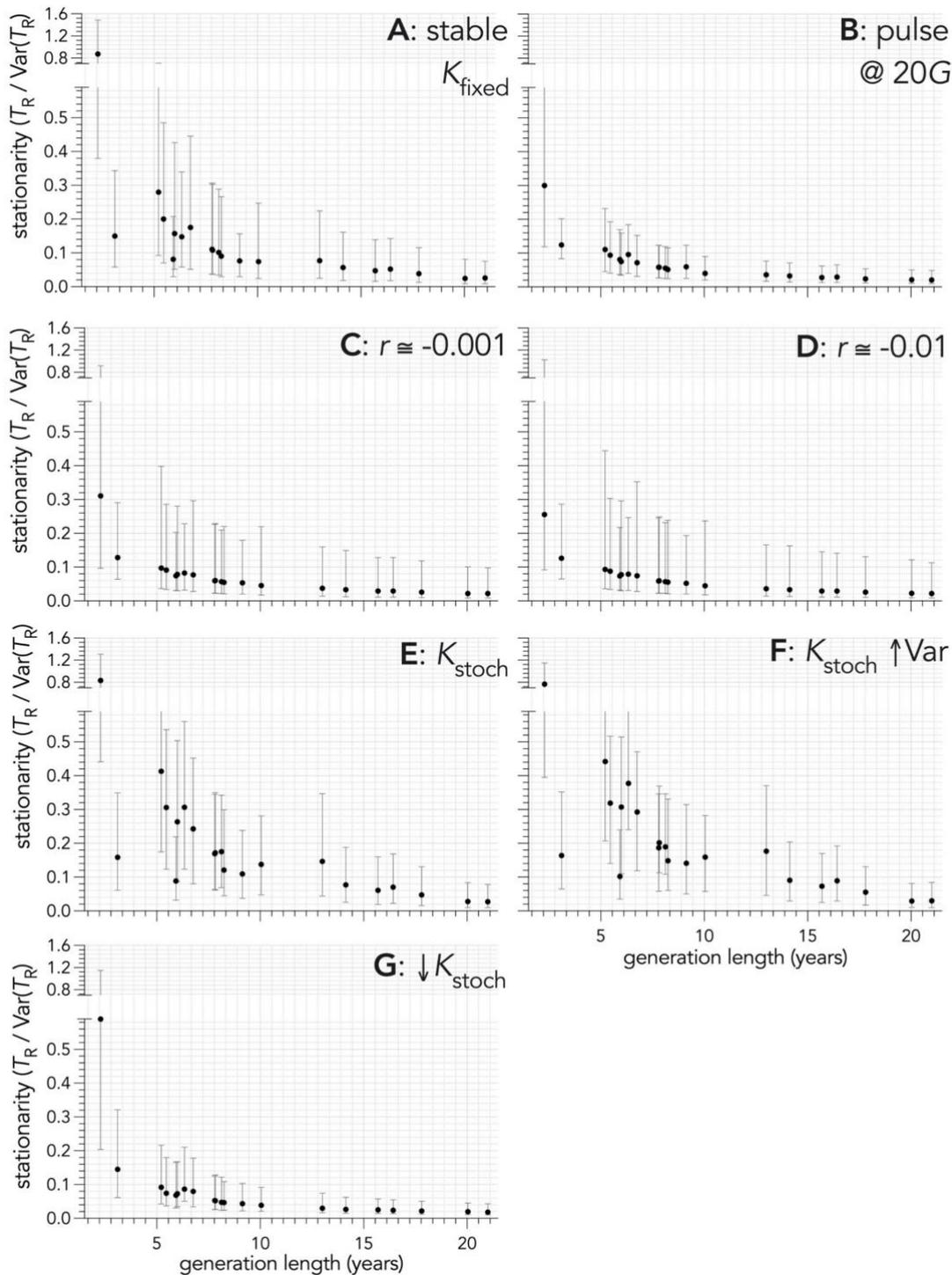


FIGURE S8 Relationships between the strength of ensemble (- Gompertz slope β , the reduction of survival as population density increases) and generation length across 10,000 times series of population abundance per species and all 21 test species (see list in Table 1) obtained from age-structured populations subjected to a compensatory component density feedback on survival over 40 generations, according to seven demographic scenarios (detailed in Table 2). Demographic scenarios include (A) carrying capacity K fixed (K_{fixed} ; scenario 1.2ii), (B) a pulse disturbance of 90% mortality at 20 generations (20G; scenario 1.2iii), (C) weakly declining ($\bar{r} \cong -0.001$; scenario 1.2iv) and (D) strongly declining ($\bar{r} \cong 0.01$; scenario 1.2v) populations, (E) K varying stochastically (K_{stoch}) around a constant mean with a constant variance (scenario 1.3vi), (F) K varying stochastically with a constant mean and increasing variance ($K_{\text{stoch}} \uparrow \text{Var}$; scenario 1.3vii), and (G) K varying stochastically with a declining mean and a constant variance ($\downarrow K_{\text{stoch}}$; scenario 1.3viii).

