

Hunting alters viral transmission and evolution

Nicholas Fountain-Jones¹, Simona Kraberger², Roderick Gagne², Marie Gilbertson³, Daryl Trumbo², Michael Charelston¹, Patricia Salerno², Chris Funk², Kevin Crooks², Ken Logan⁴, Mathew Alldredge⁴, Simon Dellicour⁵, Guy Baele⁵, Xavier Didelot⁶, Sue VandeWoude², Scott Carver¹, and Meggan Craft⁷

¹University of Tasmania

²Colorado State University

³University of Minnesota Twin Cities

⁴Colorado Parks and Wildlife

⁵KU Leuven

⁶University of Warwick

⁷University of Minnesota

March 07, 2024

Abstract

Hunting can fundamentally alter wildlife population dynamics, but the consequences of hunting on pathogen transmission and evolution remain poorly understood. Here we present a study that leverages a unique landscape-scale experiment coupled with pathogen transmission tracing, network simulation and phylodynamics to provide insights into how hunting shapes viral dynamics in puma (*Puma concolor*). We show that removing hunting pressure enhances the role of males in transmission, increases the viral population growth rate and the role of evolutionary forces on the pathogen (higher purifying and diversifying selection) compared with when hunting was reinstated. Changes in transmission could be linked to short term social changes as male population increases. These findings are supported through comparison with a region with stable hunting management over the same time period. This study shows that routine wildlife management can have profound impacts on pathogen transmission and evolution not previously considered.

Hunting alters viral transmission and evolution

Authors: Nicholas M. Fountain-Jones ^{1, 2*}, Simona Kraberger (simona.kraberger@gmail.com) ³, Roderick Gagne (rgagne@vet.upenn.edu) ³, Marie L.J. Gilbertson (mjones029@gmail.com)¹, Daryl R. Trumbo (daryl.trumbo@gmail.com) ⁴, Michael Charleston (Michael.Charleston@utas.edu.au) ², Patricia Salerno (patriciasalerno@gmail.com) ^{4,5}, W. Chris Funk (Chris.Funk@colostate.edu) ⁴, Kevin Crooks (kevin.crooks@colostate.edu) ⁶, Kenneth Logan (ken.logan@state.co.us)⁷, Mathew Alldredge (Mat.allredge@state.co.us) ⁸, Simon Dellicour (simon.dellicour@kuleuven.be) ^{9,10}, Guy Baele (guy.baele@kuleuven.be) ¹⁰, Xavier Didelot (xavier.didelot@warwick.ac.uk) ¹¹, Sue VandeWoude (Sue.Vandewoude@colostate.edu) ³, Scott Carver (Scott.Carver@utas.edu.au) ², and Meggan E. Craft (craft004@umn.edu) ^{1,12}

Affiliations

¹ Department of Veterinary Population Medicine, University of Minnesota, St Paul, Minnesota 55108.

² School of Natural Sciences, University of Tasmania, Hobart Australia 7001.

³ Department of Microbiology, Immunology, and Pathology, Colorado State University, Fort Collins, CO 80523.

⁴ Department of Biology, Graduate Degree Program in Ecology, Colorado State University, Fort Collins, CO 80523.

⁵ Universidad Regional Amazónica IKIAM, Km 7 Vía Muyuna, Napo, Ecuador

⁶ Department of Fish, Wildlife, and Conservation Biology, Colorado State University, Fort Collins, CO 80523.

⁷ Colorado Parks and Wildlife, Montrose, CO 81401 USA.

⁸ Colorado Parks and Wildlife, Fort Collins, CO 80526 USA.

⁹ Spatial Epidemiology Lab (SpELL), Université Libre de Bruxelles, CP160/12 50, av. FD Roosevelt, 1050 Bruxelles, Belgium.

¹⁰ Department of Microbiology, Immunology and Transplantation, Rega Institute, KU Leuven, Herestraat 49, 3000 Leuven, Belgium.

¹¹ School of Life Sciences and Department of Statistics, University of Warwick, Coventry, CV47AL, United Kingdom.

¹² Department of Ecology, Evolution and Behavior, University of Minnesota, St Paul, Minnesota, 55108

*Nick.FountainJones@utas.edu.au

Key words: Feline immunodeficiency virus, Mountain lion, Selection, Perturbation effect, Phylogenetic diversity, Puma, *Puma concolor*

Authorship statement: NFJ conducted the analysis and wrote the initial draft of the paper to which all authors contributed. KL and MA studied the puma populations in the field and provided the blood samples. SK, DT, PS, RG and SV collected virus and host genetic data. SD, GB, MC and XD contributed to the phylogenetic and transmission tree analyses. MG contributed to the spatial analysis. MEC, SV, KC, CF and SC conceived of the project.

Data accessibility:

DNA sequences—GenBank accession: MN563193 - MN563239. All other data and code to perform the analysis will be available on github.

Running title: Hunting alters viral evolution

Article type: Letter

Number of words (abstract): 144

Number of words (body): 4380

Number of references: 50

Number of figures: 3

Number of tables: 0

Number of text boxes: 0

Abstract:

Hunting can fundamentally alter wildlife population dynamics, but the consequences of hunting on pathogen transmission and evolution remain poorly understood. Here we present a study that leverages a unique landscape-scale experiment coupled with pathogen transmission tracing, network simulation and phylodynamics to provide insights into how hunting shapes viral dynamics in puma (*Puma concolor*). We show

that removing hunting pressure enhances the role of males in transmission, increases the viral population growth rate and the role of evolutionary forces on the pathogen (higher purifying and diversifying selection) compared with when hunting was reinstated. Changes in transmission could be linked to short term social changes as male population increases. These findings are supported through comparison with a region with stable hunting management over the same time period. This study shows that routine wildlife management can have profound impacts on pathogen transmission and evolution not previously considered.

Human actions commonly alter wildlife populations. A classic example of an alteration is hunting, which often has density and demographic effects on a population (Milner-Gulland *et al.* 2003; Whitman *et al.* 2004; Packer *et al.* 2009; Treves 2009). However, the consequences of these actions on pathogen transmission and evolution are largely unknown, and the few available studies report contradictory findings. Theory predicts that for density-dependent pathogens hunting-induced changes to density should decrease transmission rates yet make little difference to transmission dynamics for frequency-dependent pathogens. However, in practice models suggest that reducing host density can decrease (Lloyd-Smith *et al.* 2005; Potapov *et al.* 2012) or even increase pathogen transmission and prevalence (Choisy & Rohani 2006; Beeton & McCallum 2011). The complex interplay between host density, demography and behavior also makes predicting the impacts of hunting on pathogen dynamics complex. Limited empirical work shows that population reduction can increase pathogen prevalence via a ‘perturbation effect’ (Woodroffe *et al.* 2004, 2006; Carter *et al.* 2007; Carr *et al.* 2019). For example, culling induced changes or ‘perturbations’ to badger (*Meles meles*) territorial behavior was considered a driver of increased bovine tuberculosis transmission between badgers and cattle (e.g., Woodroffe *et al.* 2006). However, there is also evidence that population reduction has little impact on canine rabies (Morters *et al.* 2013) or Tasmanian devil facial tumor disease (Lachish *et al.* 2010) dynamics. Recent advances in high-resolution pathogen sequencing and analytic approaches can now elucidate patterns of pathogen transmission and evolution (Smith *et al.* 2015; Didelot *et al.* 2017; Grubaugh *et al.* 2019) that were previously out of reach. Here we address the effects of hunting on pathogen dynamics by capitalizing on pathogen sequences collected from a detailed study on the demographic effects of hunting (Logan & Runge 2020b) as well as from sequences obtained over the same time period in a region where little hunting occurred. Our approach enables us to provide insights into the cascading consequences of hunting on pathogen-host dynamics.

RNA viruses are ideal agents for examining the effect of hunting on pathogen transmission and evolution. Genomic variation rapidly accrues in RNA viruses, enabling estimation of fine-scale epidemiological processes (such as transmission between hosts) and the basic reproduction number R_0 (the average number of secondary cases per infection, Biek *et al.* 2015; Didelot *et al.* 2017). Altered transmission dynamics and the arrival of new lineages can imprint distinctive evolutionary signatures on RNA viruses as they adapt quickly to changes in host populations they encounter (Woolhouse *et al.* 2014; Pybus *et al.* 2015). For example, if the cessation of hunting led to a higher frequency of transmission events, we expect that the transmission bottleneck would lead to high purifying selection since within-host mutations are lost with transmission (e.g., Pybus & Rambaut 2009). Conversely, if new mutations entering the host population allow the pathogen to escape immune detection, we may expect an increase in diversifying selection. Altered transmission dynamics and new lineages will also shape the phylogenetic diversity of the pathogen (Fountain-Jones *et al.* 2018). For example, if novel pathogen lineages are frequently arriving into a host population with limited transmission, we would expect to see a pattern of phylogenetic dispersion (i.e., higher phylogenetic diversity than expected by chance (Webb 2000)). In contrast, phylogenetic clustering (i.e., lower phylogenetic diversity than expected by chance (Webb 2000)), may be a marker of increased transmission events within a population.

Here we leverage viral data collected from closely monitored puma (*Puma concolor*) in two areas in Colorado during the same time period: a ‘treatment region’ in which hunting pressure changed over time and a ‘stable management region’ acting as a control (hereafter ‘stable region’). We sequenced viral genes sampled from captured puma for an endemic RNA retrovirus, puma feline immunodeficiency virus (FIV_{pco}), which is a host-specific pathogen considered relatively benign and not associated with overt disease outcomes (Biek *et al.* 2003). Even though FIV_{pco} is endemic in puma populations, novel infections can spread in susceptible and previously infected individuals (Malmberg *et al.* 2019). Evidence suggests FIV_{pco} is transmitted via

aggressive interactions in most instances, although vertical transmission is also possible (Biek *et al.* 2003; Fountain-Jones *et al.* 2017). We analyzed these viral data in both regions using a transmission network approach (Didelot *et al.* 2017; Fountain-Jones *et al.* 2018) that incorporates a stochastic epidemiological model with pathogen genomic data to trace transmission between individual puma. The treatment region consisted of puma in a $\sim 12000\text{km}^2$ area in western Colorado in which hunting prior to our study was common practice (see Logan & Runge 2020a). Hunting was excluded for a five-year period (2004-2009, “no-hunting period”) and reinstated for a further five years afterwards (2009-2014, “hunting period”). During the no-hunting period in the treatment region, the population of independent pumas (i.e., adults and sub-adults) increased from an estimated 23 (2005) to 57 (2009) individuals with much of this growth occurring 2007-2010 (Logan & Runge 2020a, i.e., after a two year lag 2004-6 hereafter “Lag 1”). Adult and sub-adult male survival was significantly higher in the no-hunting period (Logan & Runge 2020a). When hunting resumed, the overall population declined after a lag of two years (hereafter ‘Lag 2’). However, the decline in abundance and demographic effects on males were severe and rapid with males > 6 years old apparently eliminated from the population after two hunting seasons (Logan & Runge 2020b). In contrast, over the same period, the stable region in the Front Range of Colorado experienced minimal hunting pressures and no change in management practice. Nearly all the individuals sampled in both regions were adults and both sexes were evenly represented. While density was higher in the stable region (see Table S1), individual survival probabilities in the stable region were unaltered across years (Moss *et al.* 2016). By comparing the treatment and stable regions, we were able to test how demographic changes caused by hunting cessation and reinstatement perturb viral transmission networks and epidemiological parameters (e.g., R_0), and also alter pathogen diversity and evolution. In doing so we begin to untangle the complex interplay between wildlife management and pathogen transmission, which is crucial for pathogen-orientated conservation and disease management strategies.

Cessation of hunting shifts transmission networks and increases R_0

We found that reducing hunting mortality had profound effects on FIV_{pco} transmission dynamics. Even though the populations in the treatment and stable region were of comparable size (Table S1), our estimates of R_0 for the same virus were two-fold higher in the treatment region compared to the stable region (with non-overlapping 95% high probability density intervals indicating that the difference is significant, Fig. 1). Other model parameters, such as generation time (time between initial FIV_{pco} infection and onward transmission, Fig. S2) and the proportion of missing cases (Fig. S3) yielded similar estimates in both regions. This burst of transmission in the treatment population was likely a product of transmission between males as they were dominant in the network (Fig. 1a). In the treatment population, males had an overall mean weighted degree (i.e., the number of likely transmission events per individual, weighted by probability of transmission occurring) double that of females (0.14 compared to 0.37), only one putative transmission event occurring between sexes, and no detected female-female transmission events. When we assessed weighted degree homophily of male-male transmission events (i.e., the number of edges only between males), and simulations revealed that the dominance of male-male transmission events in the network was not random (1000 simulated annealing network iterations, $p < 0.001$, Fig. S4a). Putative transmission events largely occurred when hunting mortality was eliminated (Fig. 1a), during which time the survival of adults and subadult males was high, age structure increased, and the abundance of independent pumas increased (Logan & Runge 2020b). Male survival rates in the hunting period were also lower than for either sex in the treatment region (Logan & Runge 2020a). Females were, however, much less connected in the transmission network in the treatment region compared to the stable region, where they were more central (Fig. 1b). In contrast to the treatment region, the stable region showed evidence of transmission from females to both females and males. Average weighted degree was higher overall for males than females in the stable region (0.46 vs 0.29). Even though weighted female-female degree homophily was higher between regions (0 vs 0.05), our simulations show that we could not reject the null hypothesis that this difference was by chance ($p = 0.692$, Fig. S4b). Female-to-female transmission events occurred between highly related females supporting our previous findings of a significant role of host relatedness in FIV spread for puma living in this region (Fountain-Jones *et al.* 2019). Taken together, our results indicate that lower hunting mortality was associated

with an increase in the number of transmission events and a transmission shift towards males.

After hunting was prohibited, the greater survival and increasing abundance of males probably resulted in greater competition between males for mates. As the dominant transmission mode for FIV_{pco} is considered to be via aggressive contacts (VandeWoude & Apetrei 2006), increased male competition for mates appears a probable explanation for the change in transmission dynamics. Further interrogation of our transmission network supports this theory, as in all but two instances, male-to-male transmission occurred between individuals with overlapping territories in the treatment region (Fig. 2/S5/S6). One transmission pair was unusual in having less spatial proximity, yet one puma of this pair was a likely immigrant to the region (M133) and could have passed through M73’s territory at some point (Fig. 2). With the exception of M73 (~6 y.o. at time of infection), all individuals involved in these transmission events were between 1-3 y.o., which is a period when males are establishing new territories and are starting to compete for access to females (Logan & Sweanor 2001; Hornocker & Negri 2010). Our results suggest it is unlikely that these males transmitted to each other prior to dispersal or via maternal or paternal contacts—since these individuals were not related based on genomic data (Trumbo *et al.* 2019). While our estimates suggest that we were able to sample approximately 40% of the FIV_{pco} infections in both regions (Fig. S3)—arguably good coverage for secretive, free-ranging wildlife—our models account for this type of missing data (Didelot *et al.* 2017). For example, nearly all putative transmission events we identified from our transmission networks were between individuals on the landscape at the same time and in most cases were captured in close spatial proximity to each other. The biological plausibility of these transmission events demonstrates the power of adapting transmission network models to trace transmission and gain epidemiological insights in systems that are difficult to observe.

Hunting alters diversity and selective pressure on the virus

Altered transmission dynamics at a population level were associated with changes in viral evolution and diversity in the treatment region. The increased number of transmission events in the no-hunting period was supported by the strong phylogenetic clustering (isolates with less phylogenetic diversity than expected by chance) detected relative to the hunting period (Fig. 3a). The link between reduced hunting pressure and increased transmission events was further supported as we did not find similar phylogenetic clustering in the stable region or hunting period (Fig. 3a). Moreover, our transmission tree results and previous phylogenetic reconstruction (Fountain-Jones *et al.* 2019) found little evidence for new lineages arriving during the no-hunting period in the treatment region (Fig. 1a). We further interrogated viral diversity patterns across time using *skygrowth* demographic analyses (Volz & Didelot 2018). Viral genetic diversity rapidly accrued at the end of the no-hunting period (~2009) before markedly declining after ~2011 when hunting was reinstated (Fig. 3b), closely mirroring male population size estimates ($R^2 = 0.8$, $p = 0.010$, Fig 3c). Female population size was not significantly correlated to viral population growth rate ($R^2 = 0.190$, $p = 0.630$, Fig. 3d) adding further evidence for the enhanced role of male interactions in transmission dynamics when hunting mortality was reduced. While we have no data on puma behavior across time, it is possible that the hunting-induced increase in male density increased competition for mates and thus aggressive interactions (Logan & Sweanor 2001). No such increase in FIV_{pco} diversity and growth rate was detected in the stable population (Fig. S7b/c).

Within the treatment region, the increase in viral diversity was underpinned by greater effects of both purifying and diversifying selection acting on individuals infected during the no-hunting period compared to the hunting period ($p = 0.01$, likelihood ratio = 6.31). Purifying selection, potentially as a signature of rapid transmission events (e.g., 22), was dominant in both periods (97.25% sites $\omega < 1$), as is often the case in error-prone RNA viruses, but stronger in the non-hunting period ($\omega_{2\text{nh}} = 0$, $\omega_{2\text{h}} = 0.1$). In contrast, there was no shift in evolutionary pressure in the same periods in the stable population ($p = 0.5$, likelihood ratio = 0.43). While impacting a smaller proportion of the loci overall (2.79% loci $\omega > 1$), there was strong diversifying selection in the no-hunting period as well ($\omega_{3\text{nh}} = 21.46$, $\omega_{3\text{h}} = 2.8$). We identified five FIV_{pco} loci under diversifying selection using the MEME routine in both regions (cutoff: p [?] 0.1). Two of these loci were only found in isolates in males and, based on our transmission models, likely infected by FIV_{pco} in

the no-hunting period. There was no signature of diversifying or purifying selection in the envelope gene (*env*), which was surprising given that *env* is generally under greater evolutionary pressure as it is responsible for the virus binding to the host cells (Kenyon & Lever 2011). All loci under diversifying selection were detected in the FIV_{pol} integrase region. Putting these lines of evidence together, we not only detected population-level impacts of demographic changes due to cessation of hunting on viral mutation, but also at the individual scale with stronger evolutionary pressure on viruses infecting males. Increased evolutionary pressure on the virus may increase the probability of a new FIV_{pco} phenotype occurring in this population. Systematic shifts in evolutionary pressure are known to occur when viruses switch hosts (e.g., Tamuri *et al.* 2009; Forniet *al.* 2017); however here we show that selective constraints on a virus can be altered in response to host demographic changes caused by wildlife hunting. We stress that FIV_{pco} is largely apathogenic in puma and therefore our findings demonstrate the types of changes in pathogen transmission dynamics that can be caused by hunting induced changes in wildlife populations.

Perturbation, management and disease

Our work provides a valuable case study on how hunting can have unexpected consequences for pathogen transmission and evolution across scales. On the surface our results seem to contradict the perturbation effect hypothesis (Carter *et al.* 2007), as hunting in our treatment region reduced the number of putative transmission events as some theoretical models suggest (Lloyd-Smith *et al.* 2005; Potapovet *al.* 2012). However, in our case the cessation of hunting in a population (which was previously hunted) facilitated demographic change via increased male survivorship and abundance that perturbed the system to a different demographic state (Logan & Runge 2020a). Even though the ‘perturbation’ was different here, as reduced hunting pressure may have resulted in more interactions, likely via enhanced male-male competition, the underlying mechanism could be similar. An expansion of the perturbation effect to include any management action that leads to demographic or behavioral change may be warranted.

Our results also reveal potential shortcomings of relying on population estimates of prevalence to understand the impact of wildlife management actions on pathogen transmission. In our case, population estimates of FIV_{pco} prevalence across time alone could not detect shifts in transmission associated with hunting and were not sensitive to changes in population size (Figs. S8/S9). The lack of signal from prevalence data may be a contributing factor behind the variability of the effects of culling on disease dynamics in empirical systems (Prentice *et al.* 2019). Prevalence data may be better able to detect shifts in population demography where the pathogen causes acute infections with shorter periods of immunity. The collection of pathogen molecular data from well-sampled wildlife populations across time is a logistical challenge, yet with ever cheaper and more mobile sequencing platforms, the potential to use approaches similar to ours is increasing, even for slowly evolving pathogens such as bacteria (Biek *et al.* 2015). This molecular and analytical approach can not only provide novel insights into the broader consequences of wildlife management on disease dynamics but can also help understand evolutionary relationships between hosts and pathogens in free-ranging species more broadly.

Materials and Methods

Study area and puma capture

Our study was conducted in two regions in the Rocky Mountains in Colorado separated by ~500 km but at similar elevations and with similar puma densities (Lewis *et al.* 2015), vegetative and landscape attributes, yet with differing degrees of urbanization (see Fig. S10 and Lewis *et al.* 2015). In the treatment region in the Uncompahgre Plateau on the Western Slope of Colorado, blood samples were taken from 114 individuals (2004-2011: 50 females & 33 males, 2012-2014: 21 females & 19 males) and monitored intensively (e.g., GPS collars) until their death or the end of the study in 2014. In the stable management region in the Front Range of Colorado, blood samples were taken from 56 individuals from 2005-2014 (2005-2011: 11 females & 5 males, 2012-2014: 21 females & 19 males). Captured pumas were anesthetized with established sedative and tranquilizer protocols (Logan 2012) and released after blood, serum, and oral swabs were collected. Animal sex, age, and capture location were recorded. See Fountain-Jones *et al.* (2019) for sample storage,

FIV_{pco} DNA extraction and sequencing details. In brief, for samples that were qPCR positive for FIV_{pco}, the complete *ORFA* and *pol* gene regions were isolated using a nested PCR protocol (Fountain-Jones *et al.* 2019). Recombination was removed and the genes were concatenated together. See Table S1 for a summary of the sequence data and a comparison of study area size, host mortality, and host genetic diversity between regions.

Transmission and phylogenetic trees

We constructed transmission trees between pumas in each region using the R package TransPhylo (Didelot *et al.* 2017). TransPhylo uses a time-stamped phylogeny to estimate a transmission tree to gain inference into “who infected whom” and when. Briefly, this approach computes the probability of an observed transmission tree given a phylogeny using a stochastic branching process epidemiological model; the space of possible transmission trees is sampled using reversible jump Markov chain Monte Carlo (MCMC) (Didelot *et al.* 2017). This approach is particularly useful for pathogens where the outbreak is ongoing, and not all cases are sampled (Didelot *et al.* 2017), as is the case here. We leveraged our FIV_{pco} Bayesian phylogenetic reconstructions from previous work and focused on the two clades of FIV_{pco} that predominantly occurred in each region (see Fountain-Jones *et al.* 2019). Whilst the TransPhylo approach makes few assumptions, a generation time distribution (the time from primary infection to onward transmission) is required to calibrate the epidemiological model (Didelot *et al.* 2017). We assumed that generation time could be drawn from a Gamma distribution ($k = 2$, $\vartheta = 1.5$) estimating onward transmission on average 3 years post-infection (95% interval: 0.3 - 8 years, based on average puma age estimates (Logan & Sweanor 2001)). Based on previous work (Lewis *et al.* 2015), we were confident that the proportion of cases (π) sampled was high, therefore we set the starting estimate of π to be 0.6 (60% of cases tested in each region), and allowed it to be estimated by the model. We ran multiple MCMC analyses of 400,000 iterations and assessed convergence by checking that the parameter effective sample size (ESS) was > 200 . We computed the posterior distributions of R_0 , π , and the realized generation time from the MCMC output. We also estimated likely infection time distributions for each individual and compared these estimates to approximate puma birth dates to ensure that these infection time distributions were biologically plausible. We then computed a consensus transmission tree for each region to visualize the transmission probabilities between individuals through time. Lastly, we reformatted the tree into a network object (nodes as individual puma and edges representing transmission probabilities) and plotted it using the *igraph* package (Csárdi & Nepusz 2006) and overlaid puma sex as a trait. Overall weighted degree and weighted degree for each sex, including edges representing homophily (e.g., male-male) and heterophily (e.g., male-female), were also calculated using *igraph*.

Simulation modelling

To test for non-random patterns of weighted degree between each sex, we applied a simulated network annealing approach from the Ergm R package (Handcock *et al.* 2018). To generate each simulated network, we fitted a variety of probability distributions to edge weight and degree of both treatment and stable regions, then used AIC to select the best fitting target distribution. Edge density, network size and the number of isolated nodes were fixed based on each observed network. We added sex to each simulated node attribute drawing from a Bernoulli distribution (probability= 0.5). Using these network characteristics, we generated 1000 ‘null’ networks and compared the homophily weighted degree distribution of each sex (i.e., the average weighted degree for each individual based on putative male-male or female-to-female transmission events) of the null networks to the observed and calculated a bootstrap p-value.

Selection analyses

To test if the demographic changes driven by hunting resulted in a reduction in the intensity of natural selection on FIV_{pco}, we examined selective pressure in both time periods in each region using the RELAX hypothesis testing framework (Wertheim *et al.* 2015). The method builds upon random effects branch-site models (BS-REL) (Kosakovsky Pond *et al.* 2011) that estimates the ω ratio (the ratio of non-synonymous to synonymous mutations or dN/dS) along each branch from a discrete distribution of three ω ratio classes allowing selection pressure to vary across the phylogeny (Wertheim *et al.* 2015). A ω ratio of one corresponds

to neutral selection with values > 1 being evidence for diversifying (positive) selection along a branch, and < 1 evidence for purifying (negative) selection along a branch. Briefly, RELAX tests for relaxation of selection pressure by dividing branches into three subsets; test branches (T), reference branches (R) and unclassified branches (U) (Wertheim *et al.* 2015) with ω_T (resp. ω_R) being the estimated dN/dS ratio on test (resp. reference) branches. The discrete distribution of ω is calculated using BS-REL for each branch class, and then branches belonging to each subset are compared. The reference estimates of ω are raised to the power of k (an intensity parameter) so that $\omega_T = \omega_R^k$ in order to simplify model comparison. The null RELAX model is when the ω distribution and thus selective pressure is the same in R and T (when $k = 1$). The null model is compared to an alternate model (using a likelihood ratio test) that allows k to vary so that when $k > 1$ selection pressure on the test branches was intensified or $k < 1$ indicating that selection pressure has been relaxed (Wertheim *et al.* 2015). In the relaxed scenario, $k < 1$ branches in R are under stronger purifying and diversifying selection compared to T branches (e.g., ω shifts from 0.1 to 0.001 or from 10 to 2). See Wertheim *et al.* (2015) for model details. T and R were selected from leaf branches (all other branches were Unassigned, U); individuals sampled from 2005-2011 (to the end of the lag period) were assigned to the R set and those sampled from 2012-2014 were assigned to T set. All branches not directly connecting to the tips were classified as ‘U’ as the majority had low phylogenetic support (posterior probability < 0.6). To further interrogate the sequence data to identify individual sites under selection, we performed the MEME (mixed-effects model of evolution) pipeline (Murrell *et al.* 2012). We performed both MEME and RELAX models using the Datamonkey web application (Weaver *et al.* 2018).

Population growth rate

We applied the non-parametric *skygrowth* method (Volz & Didelot 2018) to examine if the $FIV_{P_{CO}}$ population growth rate fluctuated across time and if this was related to changes in male or female population size in the treatment region. We did not do the same for the stable region as similar estimates were not available. We fitted these models using MCMC (100,000 iterations) assuming that $FIV_{P_{CO}}$ population size fluctuated every 6 months over a 14-year period (the estimated time to most recent common ancestor of this clade, Fig. S7). Otherwise, the default settings were used. We then performed a Pearson correlation test to assess if the trend in $FIV_{P_{CO}}$ population growth was related to male and female population size estimates (Logan & Runge 2020a). Measuring the correlation between population size estimates and patterns of population growth using generalized linear models (Gill *et al.* 2016; Volz & Didelot 2018) was not feasible due to the relatively small size of this dataset.

Phylogenetic diversity

To quantify phylogenetic diversity in each time period in each region, we calculated the standardized effect size (SES) for Faith’s phylogenetic richness that accounts for differing sample sizes (SES for Faith’s PD, (Faith 1992)). Faith’s PD (hereafter PD) is the sum of the branch lengths of the phylogenetic tree linking all isolates for each subset (in this case the two time periods). As the number of isolates in each contrast differed (stable region 2005-2011: 11 isolates, stable region 2012-13: 5 isolates, treatment region 2005-2011: 10 isolates, treatment region 2012-14: 5 isolates) we calculated the standardized effect size (SES) by comparing the PD we observed to a null model that accounts for number of tips (i.e., how much phylogenetic diversity would we see for a given number of isolates by chance). We denote the standardized PD as SES.PD from here on; this was calculated across a subset of posterior phylogenetic trees from our previous Bayesian phylogenetic analyses (Fountain-Jones *et al.* 2019). To capture phylogenetic uncertainty in these estimates, we utilized the computational efficiency of the *PhyloMeasures* R package algorithm (Tsirogiannis & Sandel 2016) to calculate SES.PD and apply this across a 1000 tree subsample of posterior trees (Fountain-Jones *et al.* 2019). Specifically, for each calculation of SES.PD we compared our observed PD to a uniform null model (i.e., isolate samples are taken with equal [uniform] probability). The code and data to perform these operations as well as the transmission tree analysis above can be found here: <https://github.com/nfj1380/TransmissionTreeCode>

Acknowledgments: This project was funded by the National Science Foundation Ecology of Infectious Diseases research program grants (DEB 1413925) and an Australian Research Council Discovery Project Grant (DP190102020). M.L.J.G. was supported by the Office of the Director, National Institutes of Health

under award number NIH T32OD010993. The content is solely the responsibility of the authors and does not necessarily represent the official views of the National Institutes of Health. SD is supported by the *Fonds National de la Recherche Scientifique* (FNRS, Belgium). GB acknowledges support from the Interne Fondsen KU Leuven / Internal Funds KU Leuven under grant agreement C14/18/094, and the Research Foundation – Flanders (‘Fonds voor Wetenschappelijk Onderzoek – Vlaanderen’, G0E1420N). MEC was funded by the National Science Foundation (DEB-1654609 and 2030509) and the CVM Research Office UMN Ag Experiment Station General Ag Research Funds. Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

Acknowledgments: This project was funded by the National Science Foundation Ecology of Infectious Diseases research program grants (DEB 1413925) and an Australian Research Council Discovery Project Grant (DP190102020). M.L.J.G. was supported by the Office of the Director, National Institutes of Health under award number NIH T32OD010993. The content is solely the responsibility of the authors and does not necessarily represent the official views of the National Institutes of Health. SD is supported by the *Fonds National de la Recherche Scientifique* (FNRS, Belgium). GB acknowledges support from the Interne Fondsen KU Leuven / Internal Funds KU Leuven under grant agreement C14/18/094, and the Research Foundation – Flanders (‘Fonds voor Wetenschappelijk Onderzoek – Vlaanderen’, G0E1420N). MEC was funded by the National Science Foundation (DEB-1654609 and 2030509) and the CVM Research Office UMN Ag Experiment Station General Ag Research Funds. Any use of trade, firm, or product names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

References:

- Beeton, N. & McCallum, H. (2011). Models predict that culling is not a feasible strategy to prevent extinction of Tasmanian devils from facial tumour disease. *J. Appl. Ecol.* , 48, 1315–1323.
- Biek, R., Pybus, O.G., Lloyd-Smith, J.O. & Didelot, X. (2015). Measurably evolving pathogens in the genomic era. *Trends Ecol. Evol.*
- Biek, R., Rodrigo, A.G., Holley, D., Drummond, A., Anderson, C.R., Ross, H.A., *et al.* (2003). Epidemiology, genetic diversity, and evolution of endemic feline immunodeficiency virus in a population of wild cougars. *J. Virol.* , 77, 9578–9589.
- Carr, A.N., Milleson, M.P., Hernández, F.A., Merrill, H.R., Avery, M.L. & Wisely, S.M. (2019). Wildlife management practices associated with pathogen exposure in non-native wild pigs in Florida, U.S. *Viruses* , 11.
- Carter, S.P., Delahay, R.J., Smith, G.C., Macdonald, D.W., Riordan, P., Etherington, T.R., *et al.* (2007). Culling-induced social perturbation in Eurasian badgers *Meles meles* and the management of TB in cattle: An analysis of a critical problem in applied ecology. *Proc. R. Soc. B Biol. Sci.* , 274, 2769–2777.
- Choisy, M. & Rohani, P. (2006). Harvesting can increase severity of wildlife disease epidemics. *Proc. R. Soc. B Biol. Sci.* , 273, 2025–2034.
- Csárdi, G. & Nepusz, T. (2006). The igraph software package for complex network research. *InterJournal Complex Syst.* , 1695.
- Didelot, X., Fraser, C., Gardy, J. & Colijn, C. (2017). Genomic infectious disease epidemiology in partially sampled and ongoing outbreaks. *Mol. Biol. Evol.* , 34, msw075.
- Faith, D.P. (1992). Conservation evaluation and phylogenetic diversity. *Biol. Conserv.* , 61, 1–10.
- Forni, D., Cagliani, R., Clerici, M. & Sironi, M. (2017). Molecular Evolution of Human Coronavirus Genomes. *Trends Microbiol.*
- Fountain-Jones, N.M., Kraberger, S., Gagne, R., Trumbo, D.R., Salerno, P., Funk, W.C., *et al.* (2021). Host relatedness and landscape connectivity shape pathogen spread in a large secretive carnivore. *Communications Biology* 4, 12 <https://doi.org/10.1038/s42003-020-01548-2>

- Fountain-Jones, N.M., Packer, C., Troyer, J.L., VanderWaal, K., Robinson, S., Jacquot, M., *et al.* (2017). Linking social and spatial networks to viral community phylogenetics reveals subtype-specific transmission dynamics in African lions. *J. Anim. Ecol.* , 86, 1469–1482.
- Fountain-Jones, N.M., Pearse, W.D., Escobar, L.E., Alba-Casals, A., Carver, S., Davies, T.J., *et al.* (2018). Towards an eco-phylogenetic framework for infectious disease ecology. *Biol. Rev.* , 93, 950–970.
- Gill, M.S., Lemey, P., Bennett, S.N., Biek, R. & Suchard, M.A. (2016). Understanding past population dynamics: Bayesian coalescent-based modeling with covariates. *Syst. Biol.* , 65, 1041–1056.
- Grubaugh, N.D., Ladner, J.T., Lemey, P., Pybus, O.G., Rambaut, A., Holmes, E.C., *et al.* (2019). Tracking virus outbreaks in the twenty-first century. *Nat. Microbiol.*
- Hornocker, M.G. & Negri, S. (2010). *Cougar : ecology and conservation* . University of Chicago Press.
- Handcock M, Hunter D, Butts C, Goodreau S, Krivitsky P, Morris M (2018). *ergm: Fit, Simulate and Diagnose Exponential-Family Models for Networks*. The Statnet Project (<http://www.statnet.org>). R package version 3.9.4, <https://CRAN.R-project.org/package=ergm>.
- Karcher, M.D., Palacios, J.A., Bedford, T., Suchard, M.A. & Minin, V.N. (2016). Quantifying and mitigating the effect of preferential sampling on phylodynamic inference. *PLOS Comput. Biol.* , 12, e1004789.
- Kenyon, J.C. & Lever, A.M.L. (2011). The molecular biology of feline immunodeficiency virus (FIV). *Viruses* .
- Kosakovsky Pond, S.L., Murrell, B., Fourment, M., Frost, S.D.W., Delpont, W. & Scheffler, K. (2011). A random effects branch-site model for detecting episodic diversifying selection. *Mol. Biol. Evol.* , 28, 3033–3043.
- Lachish, S., McCallum, H., Mann, D., Pukk, C.E. & Jones, M.E. (2010). Evaluation of selective culling of infected individuals to control tasmanian devil facial tumor disease. *Conserv. Biol.* , 24, 841–851.
- Lewis, J.S., Logan, K.A., Alldredge, M.W., Bailey, L.L., VandeWoude, S. & Crooks, K.R. (2015). The effects of urbanization on population density, occupancy, and detection probability of wild felids. *Ecol. Appl.* , 25, 1880–1895.
- Lloyd-Smith, J.O., Cross, P.C., Briggs, C.J., Daugherty, M., Getz, W.M., Latta, J., *et al.* (2005). Should we expect population thresholds for wildlife disease? *Trends Ecol. Evol.*
- Logan, K. & Runge, J. (2020a). Effects of hunting on a puma population in Colorado. *Wildl. Monogr.*
- Logan, K.A. (2012). *Wildlife Research Report, Colorado Division of Parks and Wildlife:62* .
- Logan, K.A. & Runge, J.P. (2020b). *Effects of hunting on a puma population in Colorado* . Denver.
- Logan, K.A. & Sweanor, L.L. (2001). *Desert puma : evolutionary ecology and conservation of an enduring carnivore* . Island Press.
- Malmberg, J.L., Lee, J.S., Gagne, R.B., Kraberger, S., Kechejian, S., Roelke, M., *et al.* (2019). Altered lentiviral infection dynamics follow genetic rescue of the Florida panther. *Proc. R. Soc. B Biol. Sci.* , 286, 20191689.
- Milner-Gulland, E.J., Bukreeva, O.M., Coulson, T., Lushchekina, A.A., Kholodova, M. V., Bekenovil, A.B., *et al.* (2003). Reproductive collapse in saiga antelope harems. *Nature* , 422, 135.
- Morters, M.K., Restif, O., Hampson, K., Cleaveland, S., Wood, J.L.N. & Conlan, A.J.K. (2013). Evidence-based control of canine rabies: a critical review of population density reduction. *J. Anim. Ecol.* , 82, 6–14.
- Moss, W.E., Alldredge, M.W. & Pauli, J.N. (2016). Quantifying risk and resource use for a large carnivore in an expanding urban-wildland interface. *J. Appl. Ecol.* , 53, 371–378.

- Murrell, B., Wertheim, J.O., Moola, S., Weighill, T., Scheffler, K. & Kosakovsky Pond, S.L. (2012). Detecting individual sites subject to episodic diversifying selection. *PLoS Genet.* , 8.
- Packer, C., Kosmala, M., Cooley, H.S., Brink, H., Pintea, L., Garshelis, D., *et al.* (2009). Sport hunting, predator control and conservation of large carnivores. *PLoS One* , 4.
- Potapov, A., Merrill, E. & Lewis, M.A. (2012). Wildlife disease elimination and density dependence. *Proc. R. Soc. B Biol. Sci.* , 279, 3139–3145.
- Prentice, J.C., Fox, N.J., Hutchings, M.R., White, P.C.L., Davidson, R.S. & Marion, G. (2019). When to kill a cull: Factors affecting the success of culling wildlife for disease control. *J. R. Soc. Interface* , 16.
- Pybus, O.G. & Rambaut, A. (2009). Evolutionary analysis of the dynamics of viral infectious disease. *Nat. Rev. Genet.*
- Pybus, O.G., Tatem, A.J. & Lemey, P. (2015). Virus evolution and transmission in an ever more connected world. *Proc. R. Soc. B Biol. Sci.*
- Smith, M.D., Wertheim, J.O., Weaver, S., Murrell, B., Scheffler, K. & Kosakovsky Pond, S.L. (2015). Less is more: An adaptive branch-site random effects model for efficient detection of episodic diversifying selection. *Mol. Biol. Evol.* , 32, 1342–1353.
- Tamuri, A.U., dos Reis, M., Hay, A.J. & Goldstein, R.A. (2009). Identifying Changes in Selective Constraints: Host Shifts in Influenza. *PLoS Comput. Biol.* , 5, e1000564.
- Treves, A. (2009). Hunting for large carnivore conservation. *J. Appl. Ecol.*
- Trumbo, D., Logan, K., Gagne, E., Kozakiewicz, C., Krabberger, S., Fountain-Jones, N., *et al.* (2019). Urbanization impacts apex predator gene flow but not genetic diversity across an urban-rural divide. *Mol. Ecol.* , 28, 4926–4940.
- Tsirogiannis, C. & Sandel, B. (2016). PhyloMeasures: a package for computing phylogenetic biodiversity measures and their statistical moments. *Ecography (Cop.)* , 39, 709–714.
- VandeWoude, S. & Apetrei, C. (2006). Going wild: lessons from naturally occurring T-lymphotropic lentiviruses. *Clin. Microbiol. Rev.* , 19, 728–762.
- Volz, E.M. & Didelot, X. (2018). Modeling the growth and decline of pathogen effective population size provides insight into epidemic dynamics and drivers of antimicrobial resistance. *Syst. Biol.* , 67, 719–728.
- Weaver, S., Shank, S.D., Spielman, S.J., Li, M., Muse, S. V. & Kosakovsky Pond, S.L. (2018). Datamonkey 2.0: A modern web application for characterizing selective and other evolutionary processes. *Mol. Biol. Evol.* , 35, 773–777.
- Webb, C.O. (2000). Exploring the phylogenetic structure of ecological communities: An example for rain forest trees. *Am. Nat.* , 156, 145–155.
- Wertheim, J.O., Murrell, B., Smith, M.D., Pond, S.L.K. & Scheffler, K. (2015). RELAX: Detecting relaxed selection in a phylogenetic framework. *Mol. Biol. Evol.* , 32, 820–832.
- Whitman, K., Starfield, A.M., Quadling, H.S. & Packer, C. (2004). Sustainable trophy hunting of African lions. *Nature* , 428, 175–178.
- Woodroffe, R., Cleaveland, S., Courtenay, O., Laurenson, M.K. & Artois, M. (2004). Infectious disease in the management and conservation of wild canids. In: *The Biology and Conservation of Wild Canids* . Oxford University Press, pp. 123–142.
- Woodroffe, R., Donnelly, C.A., Jenkins, H.E., Johnston, W.T., Cox, D.R., Bourne, F.J., *et al.* (2006). Culling and cattle controls influence tuberculosis risk for badgers. *Proc. Natl. Acad. Sci. U. S. A.* , 103, 14713–14717.

Woolhouse, M.E.J., Adair, K. & Brierley, L. (2014). RNA viruses: A case study of the biology of emerging infectious diseases. In: *One Health* . American Society of Microbiology, pp. 83–97.

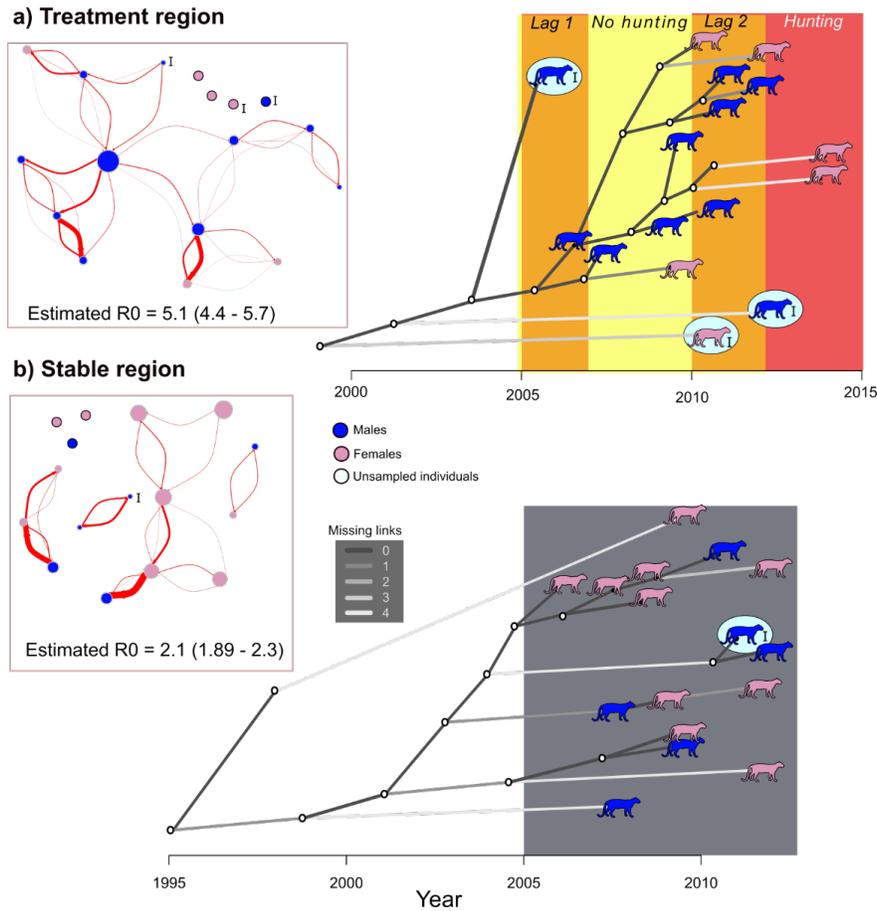


Fig. 1. Males (blue nodes/puma silhouettes) were dominant in the FIV_{pco} transmission network in the treatment region (a) whereas females (pink nodes/puma silhouettes) were more central in the transmission network in the stable region (b). Nodes connected to each other via edges indicate the probability of transmission in either direction. Node size in the networks (left) is scaled based on the number of edges estimated for each individual. Edge width is scaled according to the probability of the transmission events, where wider edges indicate a more likely transmission event (see Fig. S1). R_0 estimates (with 95% highest posterior density (HPD)) are based on the stochastic branching epidemiological model underlying each transmission network (see *Materials & Methods* , (Didelot *et al.* 2017)). Transmission trees (right) show these putative transmission events through time with branch color indicating how many missing edges are likely between individuals. Orange: lag 1 period from the start of the no-hunting period when males are recruited into the study area and lag 2 period at the start of the hunting period until the population declines; yellow: hunting pressure relieved; red: hunting pressure resumed; grey: stable region. White nodes: unsampled individuals estimate by the model. I: individuals that were likely immigrants in this region based on (Trumbo *et al.* 2019). See Figure S2 for the FIV_{pco} generation time distributions for each region and Figure S3 for the estimate of missing cases across years

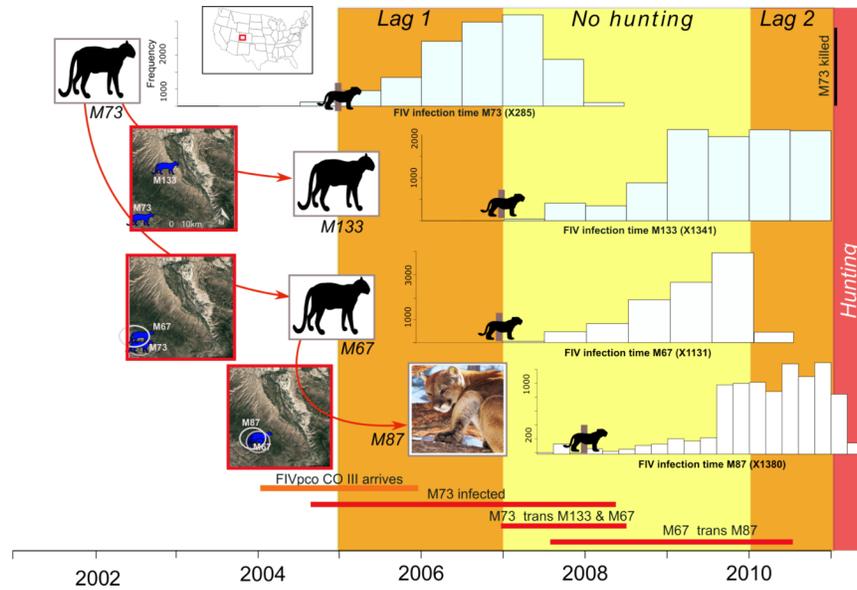


Fig. 2: Infection time distributions from our transmission network model for individuals involved in a putative transmission chain, along with the likely direction of transmission (red arrows) and the spatial context (see Fig. S5/S6 for information on other transmission events in the treatment and stable region). Grey circles encompassing puma silhouettes in the map insets represent known territorial overlap between individuals (based on observations from K. Logan) and is not representative of home range size. Orange: lag 1 period from the start of the no-hunting period when males are recruited to the study area and lag 2 period at the start of the hunting period until the population declines, light yellow: hunting pressure relieved, red: hunting pressure. Birth year is indicated by the cub silhouette, and death year of M73 is indicated by the black horizontal line. The orange horizontal line indicates when the FIV_{pcc} CO III lineage was introduced into this population based on node estimates from (Fountain-Jones *et al.* 2019). Red horizontal lines indicate transmission time distributions (overlap between infection time distributions) and ‘trans’ means ‘likely transmitted to’.

Fig. 3. Eliminating hunting mortality led to: (a) overall greater phylogenetic clustering (i.e., lower phylogenetic diversity) of FIV_{pcc} isolates standardized for sample number and (b) an increase in FIV_{pcc} population growth rate that was (c) strongly correlated with male population size rather than (d) female population size. (a) Standardized phylogenetic diversity (*: SES.PD, standardized effect size phylogenetic diversity calculated from 1000 posterior trees) estimates revealed strong patterns of phylogenetic clustering (smaller distances between isolates than expected by chance) when hunting pressure was relieved (negative values of SES.PD). Otherwise, FIV_{pcc} isolates were more dispersed across the tree (SES.PD ~ 0, indicated by the dashed line). Estimates of FIV_{pcc} prevalence (number of qPCR positives/total number sampled) are provided next to each box and whisker plot with number of individuals tested shown in parentheses (see Fig. S8 for estimates of prevalence across years). (b) Viral population growth rate was estimated using Bayesian phylodynamic reconstruction (Volz & Didelot 2018). See Fig. S7a for the corresponding skyline plot (effective population size through time estimated via the *phylodyn* model (Karcher *et al.* 2016)) for the treatment region and Fig. S7b/c for complementary plots for the FIV_{pcc} clade dominant in the stable region. Lg refers to the period between hunting being reinstated and the start of population decline.