Variation in microbial biomass and community composition based on long-term fertilization regimes in paddy soil profiles

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Abstract

Fertilization is a common approach to increase or sustain soil fertility, but its impact on microbial biomass and community structure remains controversial, particularly in paddy soils. In this study, we investigated the effect of different long-term fertilization strategies, beginning in 1986, namely no fertilization, mineral fertilization, mineral fertilization combined with rice straw or chicken manure, on microbial biomass and community composition at four soil depths (0-10, 10-20, 20-30, and 30-40 cm). The extracted soil phospholipid fatty acids (PLFAs) were pooled into gram-positive (G+) bacteria, gram-negative (G-) bacteria, fungi, and actinomycetes groups. Results showed that irrespective of the fertilization type, the abundance of PLFAs decreased with soil depth in the following order due to nutrient decrease along soil profiles: fungi > G- bacteria > G+ bacteria > actinomycetes. Mineral fertilizers are utilized more by G+ bacteria than by other microbial groups. Partial replacement of mineral fertilizer with manure further stimulates G+ bacteria at all depths. Redundancy analysis showed obvious microbial separation at the 0-20 and 20-40 cm soil depths due to the rhizodeposition effect and also revealed that the microbial communities were significantly correlated with nutrient content (soil organic carbon and available N) and pH. Overall, our findings highlight microbial community shifts due to different fertilizer types, which provides basic information for understanding how substrate availability controls the structure of soil microbial communities in paddy soil systems.

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Key words : mineral fertilizer –rhizodeposition effect –straw addition – soil depth – organic manure – paddy soil.

INTRODUCTION

Paddy soil is estimated to cover a total area of approximately 161 million ha worldwide, with 18.9% of this area in China (FAOSTAT, 2016). Mineral or organic fertilizers, or their combinations, are usually applied to paddy fields to maintain soil fertility and increase rice crop yields. Organic fertilizers commonly consist of crop residues or animal manure. Crop residue and manure addition can increase soil organic carbon (SOC) content (Yan et al., 2007; Liu et al., 2014; Zhang et al., 2016), but they can also be a source of CH₄emissions in paddy soil (Shen et al., 2007; Tang et al., 2016). Application of manure or rice straw combined with mineral fertilizer in paddy soils was demonstrated to be effective in increasing soil fertility (mainly due to increased microbial biomass and SOC) (Xu et al., 2018). Similarly, a mixture of 70% manure in combination with 30% chemical fertilizer improved rice yields and bacterial diversity, and also alleviated soil acidification (Chen et al., 2017).

Soil microorganisms play a key role in nutrient cycling, and maintaining or increasing microbial diversity is crucial for soil health (Keeler et al., 2009; Whiteside et al., 2012; Fierer, 2017). Microbial diversity is affected by several factors, such as soil physical and chemical properties (Lynn et al., 2017; van Leeuwen et al., 2017), fertilization (Huang et al., 2018), and irrigation (Azziz et al., 2016; Das et al., 2016). Most previous studies on microbial communities focused on topsoil (based on tillage or direct fertilization)because the composition is more variable in the surface horizons (Eilers et al., 2012), which generally refers to depths of 0-20 cm for paddy soils. However, a previous study suggested that nutrients can translocate from the topsoil to the deeper layers in maize and wheat fields (Kramer et al., 2013). Another previous study determined that up to 30% of microbial biomass could still be found in the C horizon (55-65 cm in grassland and cropland) (van Leeuwen et al., 2017). Hence, soil depth should be considered as an important environmental gradient that structures soil-microbial communities (Eilers et al., 2012), especially the subsoil close to the plough layers. Mineral fertilizer plus manure addition has been demonstrated to increase different taxa of bacteria and archaea for topsoil (0–20 cm) and subsoil (below 20 cm) compared to mineral fertilizer alone in long-term paddy fields (Gu et al., 2017).

Previous studies have shown that different fertilization strategies could alter different microbial-community structures. Mineral N fertilizer was reported to increase the gram-positive to gram-negative bacteria (G+ to G-) ratio in paddy soil (Zhang et al., 2007) and temperate grassland soil (Denef et al., 2009), while no ratio difference was observed in other studies (Zhang et al., 2012; Dong et al., 2014). It has been suggested that G-bacteria preferentially use labile plant-derived C and G+ bacteria that are able to utilize recalcitrant compounds from soil organic matter (SOM) (Kramer & Gleixner, 2008; Fanin et al., 2019). Complex litter addition was found to greatly promote fungi than bacteria and also increased the G- to G+ bacteria ratio in subtropical forest soil (Wang et al., 2014); the straw residue was demonstrated to promote more G- bacteria than G+ bacteria (Tang et al., 2018). Manure was reported to promote G-bacteria in rotation cropping (Peacock et al., 2001) and paddy soils (Zhang et al., 2012). However, there is still limited information regarding the microbial-community change along the soil profiles of paddy soil supplied with equal amounts but different types of fertilizer based on long-term observation. It has been demonstrated that, unlike inorganic fertilizer, straw and manure addition increase the relative abundance of *Gemmatimonadetes* and *Planctomycetia* (Tang et al., 2019).

In the present study, we investigated the microbial biomass and responses of specific microbial groups (the fungi to bacteria, G+ to G-, G+ to actinomycetes, and fungi to G- bacteria ratios) to different fertilizer types (mineral and organic combination with equal amounts of mineral fertilization [NPK]) at 0–10, 10–20, 20–30, and 30–40 cm soil depths in paddy soil systems. The microbial biomass and community composition were

measured using chloroform fumigation extraction and phospholipid fatty acid(PLFA) analyses, respectively. We hypothesized that 1) the microbial abundance and biomass would decrease with soil depth; 2) fertilization would increase microbial abundance and biomass; and 3) mineral fertilizer would promote G+ bacteria and thereby increase the G+ to G- bacteria ratio, while more complex organic fertilizer (straw and manure) would stimulate more G- bacteria.

MATERIALS AND METHODS

2.1 Site description and fertilization strategies

Soil samples were collected in April 2015 (before early rice transplanting) from long-term fertilization experiment stations at Ningxiang in the subtropical region of the Hunan Province, China (111°54'-112deg18'E, 28deg07'-28deg37'N). The mean annual temperature and rainfall of the site are 17degCand1681 mm, respectively. The soil was derived from river alluvium, Quaternary red clay, and plate shale, and the cropping system used was milk vetch-rice-rice. The main physical and chemical properties of the soil at the start of the long-term field fertilization trials were as follows: pH 5.8, SOC17.1 g kg⁻¹, total N 1.76 g kg⁻¹, available N 144mg kg⁻¹, and Olsen-P 12.8 mg kg⁻¹ (Yuan et al., 2012).

The experiment started in 1986 with the following four treatments: no fertilizer (CK), mineral fertilizer (NPK) (urea, superphosphate, and potassium chloride), rice straw combined with mineral fertilizer (NPK + ST), and 70% NPK +30% chicken manure (NPK +OM). The rice straw produced from the straw plots was returned to the corresponding plots in July and November after the early and late rice harvests at rates of 2775 kg ha⁻¹ a⁻¹ and 3600 kg ha⁻¹ a⁻¹, respectively. The experiment design ensured that all fertilizer treatments received the same amount of N, phosphorus pentoxide (P₂O₅), and potassium oxide (K₂O). The total amount of N, P₂O₅, and K₂O was the sum of the chemical fertilizer plus that from the rice straw residue or manure, corresponding to143 kg N ha⁻¹, 23.2 kg Pha⁻¹, and 52.3 kg K ha⁻¹; and 158.0 kg N ha⁻¹, 18.6 kg P ha⁻¹, and 67.2 kg K ha⁻¹per fertilized treatment for the early and late rice seasons every year, respectively. The total amounts of fertilizer applied in each fertilized treatment during the milk vetch season were identical to those for the late rice season.

2.2 Soil sampling and chemical analysis

Soil samples were collected from three replicate plots for each treatment. In each 33 m² plot, five soil cores were taken from the four layers at 0-10, 10–20, 20–30, and 30–40 cm depths and mixed thoroughly. The composite soil samples were immediately placed in a gas-permeable plastic bag and stored at 4degCuntil analysis. Subset soil samples were air-dried and sieved through a 2-mm mesh to remove plant debris and stores. The physical and chemical properties of the soil were then analysed, as presented in Table 2.

Microbial biomass carbon (MBC) and nitrogen (MBN) were determined using the fumigation-extraction method (Brookes et al., 1985; Wu et al., 1990) and analysed using a total organic carbon analyser (TOC-VWP; Shimadzu Corporation, Kyoto, Japan) and flow-injection auto-analyser (Tecator FIA Star 5000 Analyzer, Foss Tecator, Sweden), respectively. Soil pH was measured using a pH meter (Delta 320; Mettler Toledo, Columbus, OH, USA) with a soil/water ratio of 1:2.5. SOC content was determined using dry combustion in an elemental analyser (VarioMAX C/N; Elementar, Langenselbold, Germany). The soil alkali-hydrolysable nitrogen content (available N) was determined using alkali solution diffusion absorption. Olsen-P was extracted using 0.5molL⁻¹NaHCO₃(Olsen et al., 1954).

2.3 PLFA extraction

For the PLFA extraction, 2 g of freeze-dried (2-mm sieved) soil was extracted and derivatized to fatty acid methyl esters following the previously reported method of Bligh and Dyer (1959) as adapted by White et al.(1979). Briefly, the soil was first extracted twice using a 22.8-mL single-phase solution of chloroform: methanol: citrate buffer (1:2:0.8 v/v/v,0.15 M, pH 4.0). The supernatant was transferred to test tubes and split into two phases by adding equal amounts of CHCl₃ and citrate buffer. Phospholipids were then separated from neutral lipids and glycolipids on a silica acid column (Supelco, Bellefonte, PA, USA). Methyl nonadecanoate fatty acid (19:0) was added prior to derivatization as an internal standard to quantify the concentration of phospholipids. Following methylation of the phospholipids, the PLFA methyl esters were separated and identified using gas chromatography (GC; N6890; Agilent Technologies, Inc., Santa Clara, CA, USA) fitted with a MIDI Sherlock microbial identification system (version 4.5; MIDI, Inc., Newark, DE, USA). Individual PLFAs were quantified from the combined area of the peaks with mass-to-charge values (m/z 44, 45, and 46) relative to the internal standard added to each sample (Thornton et al., 2011). Iso- and anteiso-branched fatty acids (except for 10Me-branched PLFAs) were used as indicators for G+ bacteria, while monounsaturated and cyclopropyl fatty acids were used as indicators for G- bacteria (Wang et al., 2016; Ma et al., 2018). The 10Me-branched PLFAs were used as actinomycete biomarkers, while 18:2 ω 6c and arbuscular mycorrhizal fungi 16:1 ω 5c were used as fungal biomarkers. The sum of the G+ bacteria, G-bacteria, and actinomycetes was used for ratio calculation as bacteria PLFAs.

2.4 Soil respiration measurement

The subset of air-dried soil samples was then pre-incubated for 14 days at 25°C under flooded conditions. For each measurement of respiration rate, approximately 20 g (equivalent dry weight) of each soil sample was incubated in a 500-mL container at 25°C for 24 h. At the end of this period, CO₂ concentrations in the headspace were measured using an Agilent-7890a gas chromatograph equipped with a flame ionization detector (Agilent Technologies, Inc.). Soil respiration rates were calculated from the net accumulation of CO₂ over time. The soil metabolic quotient (qCO₂) was calculated by dividing the soil respiration rate by the MBC content (Plaza et al., 2004).

2.5Statistical analysis

Data were analysed using one-way analysis of variance for different groups of soil depth within each fertilization regime and different groups of fertilization regimes within each soil depth, and were compared using Duncan's test in SPSS 19.0 (SPSS Inc., Chicago, IL, USA), with significance defined as P < 0.05. The relationship between the microbial structures and physicochemical properties was analysed using redundancy analysis (RDA) using Canoco 5.0 for Windows (Microcomputer Power, Ithaca, NY, USA).

RESULTS

3.1 Effect of fertilization strategies on microbial communities along the paddy soil profile

The absolute abundance of total and specific microbial groups (G-, G+ bacteria, actinomycetes, and fungi) in all treatments significantly decreased with soil depth (Table 1, P < 0.05). The abundance of total and specific PLFAs were the lowest and highest in the control and NPK + OM treatments, respectively, at the 0-40 cm soil depth. Compared with that in the NPK treatments, NPK + ST showed significantly higher bacterial (G-, G+, and actinomycetes) and fungal abundance at the 10-20 cm soil depth (P < 0.05). The fungi to bacteria, G+ bacteria to actinomycetes, and fungi to G- bacteria ratios gradually decreased with soil depth in all treatments between 0 and 30 cm. The G+ to G-bacteria ratio increased with soil depth for all treatments (Figure 1).

Both mineral and organic fertilizers significantly increased the G+ to G- bacteria and actinomycetes ratios compared with those in the control group at 0–30 cm soil depths (Figures 1b and c, P < 0.05). NPK + ST showed the lowest fungi to bacteria and G- bacteria ratios at 0–10 cm compared with those in the other treatments (Figures 1a and d, P < 0.05). At 10-40 cm, the fertilized treatments showed higher fungi to bacteria ratios compared to those of the control.

There was no uniform pattern throughout the soil profile regarding the influence of organic fertilizers on microbial functional PLFA ratios. With equal amounts of fertilizer input, compared with NPK, NPK + OM further stimulated more G+ bacteria than G- bacteria and actinomycetes at all depths (Figures 1b and c). Soil respiration rates gradually decreased with soil depth and the lowest values for each soil depth were observed in the control (Figure 2a, P < 0.05). NPK + ST significantly increased the soil respiration rate at 0-20 cm (P < 0.05). NPK + OM showed a constant soil qCO₂ between the 0 and 40 cm soil depths, while both the NPK treatments and the control significantly increased the qCO₂ values with soil depth. NPK + ST treatment showed the lowest qCO₂ at 10-20 cm compared to that at other depths (Figure 2b, P < 0.05).

3.2 Relationship between microbial communities and soil properties

SOC content and available N decreased with soil depth (Table 2, P < 0.05). Compared to that of the control, NPK +ST and NPK +OM showed significantly higher SOC contents at soil each depth. NPK addition alone resulted in lower SOC content at 0–10 cm compared to that in the control. In contrast, soil pH exhibited the opposite pattern of SOC, which was significantly increased with soil depth, with the highest value observed in the control (Table 2, P < 0.05).

MBC and MBN decreased with soil depth in all treatments (Table 2, P < 0.05). Compared to those in the control, NPK and NPK+ OM significantly increased MBC and MBN at0–40 cm, except for 10–20 cm in NPK. In contrast, NPK + ST significantly increased MBC at0–30 cm compared to that in the control (P < 0.05). Compared to those under NPK addition alone, the NPK + ST treatment increased MBC and MBN at 10–30 cm and 0–30 cm, respectively; NPK+ OM treatments showed significantly higher MBC and MBN at 0–40 cm (Table 2, P < 0.05).

The RDA showed that SOC significantly contributed to the soil microbial communities and explained 87% of the variance in the first two axes (Figure 3a, P < 0.001). Soil pH was also significantly correlated with the microbial communities and explained 3.7% of the total variation. Available N and SOC contributed significantly to the microbial communities at both the 0–20 and 20–40 cm soil depths and explained 79.3% and 4.5%, and 3.1% and 86.4% of the total variation for each depth, respectively (P < 0.05). Soil pH only showed a significant correlation to the microbial communities at the 0–20 cm soil depths and explained 4.9% of total the variation (P < 0.013) (Figures3b and c).

DISCUSSION

4.1 Microbial groups along soil depth

The decline in the abundance of microbial PLFAs with soil depth (Table 1, P < 0.05) could arise from the reduction in energy (carbon) and nutrient availability along the profile. A reduction in the ratio of fungi to bacteria with soil depth (Figure 1a) is also an indication that the fungal population declined much faster than that of bacteria along soil profiles. Similarly, the decreasing trends of the fungi to G-bacteria and G+ bacteria to actinomycetes ratios, as well as the increasing G+ to G-bacteria ratio with soil depth could mean that microbial dependence on nutrients adhered to the following order: fungi > G- bacteria> G+ bacteria > actinomycetes. Fungal abundance and activity are considered low in paddy soil because of prolonged anaerobic conditions. However, in paddy soils, rice plants could release oxygen from the root through aerenchyma (Frenzel et al., 1992) and created an aerobic environment in the rhizosphere. Previous studies have demonstrated that fungi have the highest capacity for assimilating rhizodeposits in paddy soils (Ge et al., 2016). Such translocation to fungi can be rapidly detected in PLFAs due to the direct connection of fungal hyphae and mycorrhizal fungi with rice roots (Yuan et al., 2016). Therefore, oxygen limitation caused by root biomass reduction with soil depth could largely suppress fungal abundance. G-bacteria have been reported to preferentially use labile compounds such as rhizodeposits and exudates from plant biomass, and G+ bacteria are able to use recalcitrant compounds from SOM (Kramer and Gleixner, 2008; Fanin et al., 2019). Zhu et al. (2017) also found that G+ bacteria increased after rhizo- and micro-C addition. Labile substrates decrease with increasing soil depth, thereby causing G-bacteria to decline faster than G+ bacteria. As actinomycetes are well known for their key role in degrading complex compounds (Acosta-Martinez et al., 2008), they are relatively the least sensitive to labile substrate reduction with soil depth. Thus, the sensitivity trends of the PLFAs with soil depth were not altered by fertilization type.

The decrease in MBC and MBN with soil depth corresponded with the decrease in SOC and available N content along the soil profile (Table 2). This indicates that the growth of microorganisms is related to the availability of C and N nutrients (Loeppmann et al., 2016; van Leeuwen et al., 2017). SOC levels are fundamentally determined by the balance between organic matter inputs and their losses through decomposition (Six, Frey, Thiet, & Batten, 2006). In our current study, rhizodeposition and fertilization were the two main sources of microbial C nutrient availability. Zhu et al. (2017) found that up to 45% of rice rhizo-C was stabilized within SOC. In paddy soils, more than 50% of the total root biomass is allocated to the first 5

cm of surface soil; therefore, the upper soil layer can receive more C input than the soils below (Li et al., 2004; Li & Yagi, 2004). The decrease in the δ^{13} C isotopic signature of soil and PLFA from C₄ plants with soil depth after C₃-C₄ vegetation change in a previous study also demonstrated the decreasing influence of plants on deeper soil (Kramer & Gleixner, 2008). Higher MBC and MBN contents were observed at 0–20 cm in the control, indicating that soil layers with fresh C input can increase microbial biomass. RDA results also showed an obvious separation of the 0–20 and 20–40 cm soil layers, even in the control without fertilizer, confirming the rhizodeposition effect on the 0–20 cm soil layer (Figure 3a).

4.2 Effect of fertilization strategies on microbial community structure

All fertilized treatments increased microbial PLFAs (Table 1), showing a fertilizer-induced increase in microbial abundance. Large amounts of N from mineral and organic fertilizers could benefit microbes through a reduction in nutrient competition with rice plants (Jackson, Burger, & Cavagnaro, 2008; Zhu et al., 2018). Moreover, the increase of plant biomass caused by N fertilization would have a direct impact on rhizodeposition and the high availability of labile C to microbes (Christopher &Lal, 2007; Zhu, Vivanco, & Manter, 2016). G- bacteria are usually considered to benefit more from rhizodeposits (Fanin et al., 2019). However, the G+ to G-bacteria ratio was enriched $(0.55^{\circ}0.83 \text{ after fertilizers vs. } 0.45^{\circ}0.53 \text{ in control})$ at 0-30 cm soil depths (Figures 2b and c, P < 0.05). Moreover, the abundance of G+ bacteria exceeded that of actinomycetes at 0-30 cm. This indicated that both mineral and organic fertilizers preferentially promote G_{+} bacteria. Pure manure has been demonstrated to induce G-bacteria and reduce G+ bacteria (Peacock et al., 2001). Wang et al. (2014) also found more C incorporation by G-bacteria, which had originally been dominant in the soils. In parallel, they found that litter addition would reduce the G+ to G- bacteria ratio, but its combination with mineral N could buffer this decrease. This indicated that complex substrates induced more G-bacteria and mineral N promoted G+ bacteria, thereby reducing the G+ to G-bacteria ratio. Relatively increased G+ bacteria levels due to mineral N fertilization have also been reported in grassland soils (Denef et al., 2009) and corn fields (Peacock et al., 2001). Consistent with these observations, all treatments with mineral N fertilizers in the present study induced G+ bacteria, even when the soil was predominated by G-bacteria. This indicates a positive response of soil G+ bacteria to mineral N fertilizers.

After partial replacement of inorganic fertilizer with straw, PLFA biomarkers increased at 10–20 cm compared with those under NPK addition alone (Table 2, P < 0.05). The fungi to bacteria and G- bacteria ratios were also relatively increased at this soil depth (Figures 1a and d). This indicated that fungi were especially promoted at 10–20 cm, probably owing to more suitable anaerobic conditions for straw decomposition. Simultaneously, qCO₂ was observed to be the lowest at this soil depth. Some studies have suggested that higher qCO₂ means higher stress or disturbance for microorganisms (Wardle & Ghani, 1995). Under the same level of fertilizer input comparable to straw addition, MBC and PLFAs were largely induced by chicken manure throughout the soil profile (Tables 1 and 2, P < 0.05). Manure addition further increased the G+ to G- bacteria and G+ bacteria to actinomycetes ratios (Figures 1b and c), which may indicate that manure induced more G+ bacteria. In addition, the NPK + OM treatments showed a consistently low qCO₂ throughout the 0-40 cm soil profile (Figure 2). Plaza et al. (2004) observed that qCO₂ initially decreased after pig manure amendment to soil and then increased with further manure addition. Hence, qCO₂ was not negatively correlated with soil fertility or organic fertilizer input. The findings above may preferentially support the connection between low qCO₂ and favourable soil conditions for microorganisms.

In addition to its effect on microbial communities through nutrient availability, fertilization can indirectly alter microbial communities through its effect on pH at 0–20 cm soil depths (Figures 3a and b). More organic acids such as oxalic acid (Keiluweit et al., 2015) and acetate (Farrar, Hawes, Jones, & Lindow, 2003) released from rhizodeposits could partly cause a reduction in soil pH, especially in the rooted layers compared with the deeper layers (Table 2, Fig 3a-b, P < 0.05). In addition, N fertilizer has been reported to cause soil acidification during N cycling (Bolan, Headley, & White, 1991; Geissler & Scow, 2014). However, in contrast to general assumptions, higher SOC contents were observed in the control compared with those in NPK at 0–10 cm. The soil respiration rate and qCO₂ were higher in NPK in this soil layer (Figures 2a and b, P<0.05), and carbon use efficiency was also found to be lower in NPK (Zhran et al., 2020). A greater amount of respired than accumulated C in biomass can partially explain the decrease in SOC. In addition, Loeppmann et al. (2016) found that maize rhizodeposition can decrease the proportional activity of C- to N-cycling enzymes. Mineral NPK fertilizers can increase root biomass and then increase rhizodeposition, but it can also stimulate microorganisms that need both C and N. More C-cycling enzymes will be released by microbes to decompose SOC to maintain a certain microbial C:N ratio (Devi & Yadava, 2006) if microorganisms need more C than rhizodeposits can supply.

5. Conclusions

Long-term mineral and organic fertilization were demonstrated to have a positive effect on microbial biomass enhancement. Our data clearly showed that the soil microbial biomass and PLFAs decreased with soil depth owing to decreasing available nutrients along soil profiles. The microbial sensitivity of nutrients declined with soil depth in the following order: fungi > G- bacteria > G+ bacteria > actinomycetes; this order cannot be changed by fertilization type. Moreover, mineral fertilizers induced G+ bacteria more than G- bacteria and actinomycetes, which suggested a close link between G+ bacteria and mineral fertilizers. With equal fertilizer input, straw addition especially promoted fungi at 10-20cm soil depth, while chicken manure further stimulated G+ bacteria. In broader perspective, understanding the mechanisms of microbial group shift due to fertilization in paddy systems may help improving soil quality through better fertilization management. Further studies should include isotope analyses (^{13}C , ^{14}C , ^{15}N) and microbial molecular analyses (16S rRNA sequencing, PLFA, SIP, etc) to fully reveal the fertilizer nutrient turnover *in situ* in paddy soil systems.

Conflict of interest statement: The authors declare no conflict of interest.

Data Availability Statement: The data that support the findings of this study are available from the corresponding author upon reasonable request.

References

Acosta-Martínez, V., Dowd, S., Sun, Y., & Allen, V. (2008). Tag-encoded pyrosequencing analysis of bacterial diversity in a single soil type as affected by management and land use. *Soil Biology and Biochemistry*, 40 (11), 2762-2770. doi:10.1016/j.soilbio.2008.07.022

Azziz, G., Trasante, T., Monza, J., & Irisarri, P. (2016). The effect of soil type, rice cultivar and water management on ammonia-oxidizing archaea and bacteria populations. *Applied Soil Ecology*, 100, 8-17. doi:10.1016/j.apsoil.2015.11.009

Bligh, E. G., & Dyer, W. J. (1959). A rapid method of total lipid extraction and purification. *Canadian Journal of Biochemistry and Physiology*, 37 (8), 911-917. doi:10.1139/o59-099

Bolan, N., Hedley, M., & White, R. (1991). Processes of soil acidification during nitrogen cycling with emphasis on legume based pastures. *Plant and Soil*, 134 (1), 53-63. doi:10.1007/BF00010717

Brookes, P. C., Landman, A., Pruden, G., & Jenkinson, D. S. (1985). Chloroform fumigation and the release of soil nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen in soil. *Soil Biology & Biochemistry*, 17 (6), 837-842. doi:10.1016/0038-0717(85)90144-0

Chen, D., Yuan, L., Liu, Y., Ji, J., & Hou, H. (2017). Long-term application of manures plus chemical fertilizers sustained high rice yield and improved soil chemical and bacterial properties. *European Journal of Agronomy*, 90, 34-42. doi:10.1016/j.eja.2017.07.007

Christopher, S. F., & Lal, R. (2007). Nitrogen management affects carbon sequestration in North American cropland soils. *Critical Reviews in Plant Sciences*, 26 (1), 45-64. doi:10.1080/07352680601174830

Das, S., Chou, M. L., Jean, J. S., Liu, C. C., & Yang, H. J. (2016). Water management impacts on arsenic behavior and rhizosphere bacterial communities and activities in a rice agro-ecosystem. *Science of the Total Environment*, 542 (Pt A), 642-652. doi:10.1016/j.scitotenv.2015.10.122

Denef, K., Roobroeck, D., Manimel Wadu, M. C. W., Lootens, P., & Boeckx, P. (2009). Microbial community composition and rhizodeposit-carbon assimilation in differently managed temperate grassland soils. *Soil Biology and Biochemistry*, 41 (1), 144-153. doi:10.1016/j.soilbio.2008.10.008

Devi, N. B., & Yadava, P. S. (2006). Seasonal dynamics in soil microbial biomass C, N and P in a mixed-oak forest ecosystem of Manipur, North-east India. *Applied Soil Ecology*, 31 (3), 220-227. doi:10.1016/j.apsoil.2005.05.005

Dong, W.-Y., Zhang, X.-Y., Dai, X.-Q., Fu, X.-L., Yang, F.-T., Liu, X.-Y., . . . Schaeffer, S. (2014). Changes in soil microbial community composition in response to fertilization of paddy soils in subtropical China. *Applied Soil Ecology*, 84, 140-147. doi:10.1016/j.apsoil.2014.06.007

Eilers, K. G., Debenport, S., Anderson, S., & Fierer, N. (2012). Digging deeper to find unique microbial communities: The strong effect of depth on the structure of bacterial and archaeal communities in soil. *Soil Biology and Biochemistry*, 50, 58-65. doi:10.1016/j.soilbio.2012.03.011

Fanin, N., Kardol, P., Farrell, M., Nilsson, M.-C., Gundale, M. J., & Wardle, D. A. (2019). The ratio of Gram-positive to Gram-negative bacterial PLFA markers as an indicator of carbon availability in organic soils. *Soil Biology and Biochemistry*, 128, 111-114. doi:10.1016/j.soilbio.2018.10.010

FAOSTAT, http://faostat3.fao.org/faostat-gateway/go/to/home/E, (2016).

Farrar, J., Hawes, M., Jones, D., & Lindow, S. (2003). How roots control the flux of carbon to the rhizosphere. *Ecology*, 84 (4), 827-837. doi:10.1890/0012-9658(2003)084[0827:HRCTFO]2.0.CO;2

Fierer, N. (2017). Embracing the unknown: disentangling the complexities of the soil microbiome. *Nat Rev Microbiol*, 15 (10), 579-590. doi:10.1038/nrmicro.2017.87

Frenzel, P., Rothfuss, F., & Conrad, R. (1992). Oxygen profiles and methane turnover in a flooded rice microcosm. *Biology and Fertility of Soils*, 14, 84-89. doi:10.1007/BF00336255

Ge, T., Li, B., Zhu, Z., Hu, Y., Yuan, H., Dorodnikov, M., . . . Kuzyakov, Y. (2016). Rice rhizodeposition and its utilization by microbial groups depends on N fertilization. *Biology and Fertility of Soils*, 53 (1), 37-48. doi:10.1007/s00374-016-1155-z

Geisseler, D., & Scow, K. M. (2014). Long-term effects of mineral fertilizers on soil microorganisms - A review. Soil Biology & Biochemistry, 75, 54-63. doi:10.1016/j.soilbio.2014.03.023

Gu, Y. F., Wang, Y. Y., Lu, S. E., Xiang, Q. J., Yu, X. M., Zhao, K., . . . Zhang, X. P. (2017). Long-term fertilization structures bacterial and archaeal communities along soil depth gradient in a paddy soil. *Frontiers in Microbiology*, 8, 15. doi:10.3389/fmicb.2017.01516

Huang, X. Z., Wang, C., Liu, Q., Zhu, Z. K., Lynn, T. M., Shen, J. L., . . . Wu, J. S. (2018). Abundance of microbial CO₂-fixing genes during the late rice season in a long-term management paddy field amended with straw and straw-derived biochar. *Canadian Journal of Soil Science*, 98 (2), 306-316. doi:10.1139/cjss-2017-0098

Jackson, L. E., Burger, M., & Cavagnaro, T. R. (2008). Roots, nitrogen transformations, and ecosystem services. Annu Rev Plant Biol, 59, 341-363. doi:10.1146/annurev.arplant.59.032607.092932

Keeler, B. L., Hobbie, S. E., & Kellogg, L. E. (2009). Effects of long-term nitrogen addition on microbial enzyme activity in eight forested and grassland sites: implications for litter and soil organic matter decomposition. *Ecosystems*, 12 (1), 1-15. doi:10.1007/s10021-008-9199-z

Keiluweit, M., Bougoure, J. J., Nico, P. S., Pett-Ridge, J., Weber, P. K., & Kleber, M. (2015). Mineral protection of soil carbon counteracted by root exudates. *Nature Climate Change*, 5 (6), 588-595. doi:10.1038/nclimate2580

Kramer, C., & Gleixner, G. (2008). Soil organic matter in soil depth profiles: Distinct carbon preferences of microbial groups during carbon transformation. *Soil Biology and Biochemistry*, 40 (2), 425-433. doi:10.1016/j.soilbio.2007.09.016

Kramer, S., Marhan, S., Haslwimmer, H., Ruess, L., & Kandeler, E. (2013). Temporal variation in surface and subsoil abundance and function of the soil microbial community in an arable soil. *Soil Biology & Biochemistry*, 61, 76-85. doi:10.1016/j.soilbio.2013.02.006

Li, C., Mosier, A., Wassmann, R., Cai, Z., Zheng, X., Huang, Y., . . . Lantin, R. (2004). Modeling greenhouse gas emissions from rice-based production systems: Sensitivity and upscaling. *Global Biogeochemical Cycles*, 18 (1), n/a-n/a. doi:10.1029/2003gb002045

Li, Z., & Yagi, K. (2004). Rice root-derived carbon input and its effect on decomposition of old soil carbon pool under elevated CO₂. Soil Biology and Biochemistry, 36 (12), 1967-1973. doi:10.1016/j.soilbio.2004.04.041

Liu, C., Lu, M., Cui, J., Li, B., & Fang, C. M. (2014). Effects of straw carbon input on carbon dynamics in agricultural soils: a meta-analysis. *Global Change Biology*, 20 (5), 1366-1381. doi:10.1111/gcb.12517

Loeppmann, S., Blagodatskaya, E., Pausch, J., & Kuzyakov, Y. (2016). Enzyme properties down the soil profile - A matter of substrate quality in rhizosphere and detritusphere. *Soil Biology and Biochemistry*, 103, 274-283. doi:10.1016/j.soilbio.2016.08.023

Lynn, T. M., Liu, Q., Hu, Y., Yuan, H., Wu, X., Khai, A. A., . . . Ge, T. (2017). Influence of land use on bacterial and archaeal diversity and community structures in three natural ecosystems and one agricultural soil. *Arch Microbiol*, 199 (5), 711-721. doi:10.1007/s00203-017-1347-4

Ma, Q., Wu, L., Wang, J., Ma, J., Zheng, N., Hill, P. W., . . . Jones, D. L. (2018). Fertilizer regime changes the competitive uptake of organic nitrogen by wheat and soil microorganisms: An in-situ uptake test using ¹³C, ¹⁵N labelling, and ¹³C-PLFA analysis. *Soil Biology & Biochemistry*, 125, 319-327. doi:10.1016/j.soilbio.2018.08.009

Olsen, S. R., Cole, C. V., Watanabe, F. S., & Dean, L. A. (1954). Estimation of available phosphorus in soils by extraction with sodium bicarbonate. U S Dept Agric Circ, 939, 1-19.

Peacock, A. g., Mullen, M., Ringelberg, D., Tyler, D., Hedrick, D., Gale, P., & White, D. (2001). Soil microbial community responses to dairy manure or ammonium nitrate applications. *Soil Biology and Biochemistry*, 33 (7-8), 1011-1019. doi:10.1016/S0038-0717(01)00004-9

Plaza, C., Hernandez, D., Garcia-Gil, J. C., & Polo, A. (2004). Microbial activity in pig slurry-amended soils under semiarid conditions. Soil Biology & Biochemistry, 36 (10), 1577-1585. doi:10.1016/j.soilbio.2004.07.017

Shen, M.-X., Yang, L.-Z., Yao, Y.-M., Wu, D.-D., Wang, J., Guo, R., & Yin, S. (2007). Long-term effects of fertilizer managements on crop yields and organic carbon storage of a typical rice-wheat agroecosystem of China. *Biology and Fertility of Soils*, 44 (1), 187-200. doi:10.1007/s00374-007-0194-x

Six, J., Frey, S. D., Thiet, R. K., & Batten, K. M. (2006). Bacterial and fungal contributions to carbon sequestration in agroecosystems. Soil Science Society of America Journal, 70 (2), 555. doi:10.2136/sssaj2004.0347

Tang, H., Xiao, X., Xu, Y., Li, C., Cheng, K., Pan, X., & Li, W. (2019). Utilization of carbon sources in the rice rhizosphere and nonrhizosphere soils with different long-term fertilization management. *J Basic Microbiol*, 59 (6), 621-631. doi:10.1002/jobm.201800736

Tang, H. M., Xiao, X. P., Wang, K., Li, W. Y., Liu, J., & Sun, J. M. (2016). Methane and nitrous oxide emissions as affected by long-term fertilizer management from double-cropping paddy fields in Southern China. *The Journal of Agricultural Science*, 154 (8), 1378-1391. doi:10.1017/s0021859615001355

Tang, H. M., Xu, Y. L., Xiao, X. P., Li, C., Li, W. Y., Cheng, K. K., . . . Sun, G. (2018). Impacts of long-term fertilization on the soil microbial communities in double-cropped paddy fields. *The Journal of Agricultural Science*, 156 (7), 857-864. doi:10.1017/s0021859618000825

Thornton, B., Zhang, Z. L., Mayes, R. W., Hogberg, M. N., & Midwood, A. J. (2011). Can gas chromatography combustion isotope ratio mass spectrometry be used to quantify organic compound abundance? *Rapid Communications in Mass Spectrometry*, 25 (17), 2433-2438. doi:10.1002/rcm.5148

van Leeuwen, J. P., Djukic, I., Bloem, J., Lehtinen, T., Hemerik, L., de Ruiter, P. C., & Lair, G. J. (2017). Effects of land use on soil microbial biomass, activity and community structure at different soil depths in the Danube floodplain. *European Journal of Soil Biology*, 79, 14-20. doi:10.1016/j.ejsobi.2017.02.001

Wang, J., Chapman, S. J., & Yao, H. (2016). Incorporation of ¹³C-labelled rice rhizodeposition into soil microbial communities under different fertilizer applications. *Applied Soil Ecology*, 101, 11-19. doi:10.1016/j.apsoil.2016.01.010

Wang, Q. K., Wang, S. L., He, T. X., Liu, L., & Wu, J. B. (2014). Response of organic carbon mineralization and microbial community to leaf litter and nutrient additions in subtropical forest soils. *Soil Biology & Biochemistry*, 71, 13-20. doi:10.1016/j.soilbio.2014.01.004

Wardle, D. A., & Ghani, A. (1995). A critique of the microbial metabolic quotient (qCO₂) as a bioindicator of disturbance and ecosystem development. *Soil Biology & Biochemistry*, 27 (12), 1601-1610. doi:10.1016/0038-0717(95)00093-t

White, D. C., Davis, W. M., Nickels, J. S., King, J. D., & Bobbie, R. J. (1979). Determination of the sedimentary microbial biomass by extractable lipid phosphate. *Oecologia*, 40 (1), 51-62. doi:10.1007/bf00388810

Whiteside, M. D., Digman, M. A., Gratton, E., & Treseder, K. K. (2012). Organic nitrogen uptake by arbuscular mycorrhizal fungi in a boreal forest. *Soil Biol Biochem*, 55. doi:10.1016/j.soilbio.2012.06.001

Wu, J., Joergensen, R. G., Pommerening, B., Chaussod, R., & Brookes, P. C. (1990). Measurement of soil microbial biomass C by fumigation extraction - an automated procedure. *Soil Biology & Biochemistry*, 22 (8), 1167-1169. doi:10.1016/0038-0717(90)90046-3

Xu, Y., Tang, H., Xiao, X., Li, W., Li, C., Sun, G., & Cheng, K. (2018). Effects of Long-Term Fertilization Management Practices on Soil Microbial Carbon and Microbial Biomass in Paddy Soil at Various Stages of Rice Growth. *Revista Brasileira de Ciência do Solo*, 42 (0). doi:10.1590/18069657rbcs20170111

Yan, D., Wang, D., & Yang, L. (2007). Long-term effect of chemical fertilizer, straw, and manure on labile organic matter fractions in a paddy soil. *Biology and Fertility of Soils*, 44 (1), 93-101. doi:10.1007/s00374-007-0183-0

Yuan, H., Ge, T., Wu, X., Liu, S., Tong, C., Qin, H., . . . Wu, J. (2012). Long-term field fertilization alters the diversity of autotrophic bacteria based on the ribulose-1,5-biphosphate carboxylase/oxygenase (RubisCO) large-subunit genes in paddy soil. *Applied Microbiology and Biotechnology*, 95 (4), 1061-1071. doi:10.1007/s00253-011-3760-y

Yuan, H., Zhu, Z., Liu, S., Ge, T., Jing, H., Li, B., . . . Kuzyakov, Y. (2016). Microbial utilization of rice root exudates: 13C labeling and PLFA composition. *Biology and Fertility of Soils*, 52 (5), 615-627. doi:10.1007/s00374-016-1101-0

Zhang, J. G., Bo, G. D., Zhang, Z. F., Kong, F. Y., Wang, Y., & Shen, G. M. (2016). Effects of Straw Incorporation on Soil Nutrients, Enzymes, and Aggregate Stability in Tobacco Fields of China. *Sustainability*, 8 (8). doi:10.3390/su8080710

Zhang, Q., Wang, G., & Yao, H. (2007). Phospholipid fatty acid patterns of microbial communities in paddy soil under different fertilizer treatments. *Journal of environmental sciences*, 19 (1), 55-59. doi:10.1016/S1001-0742(07)60009-6

Zhang, Q. C., Shamsi, I. H., Xu, D. T., Wang, G. H., Lin, X., Jilani, G., . . . Chaudhry, A. N. (2012). Chemical fertilizer and organic manure inputs in soil exhibit a vice versa pattern of microbial community structure. *Applied Soil Ecology*, 57, 1-8. doi:10.1016/j.apsoil.2012.02.012 Zhran, M., Ge, T., Tong, Y., Deng, Y., Wei, X., Lynn, T. M., . . . Gunina, A. (2020). Assessment of depthdependent microbial carbon-use efficiency in long-term fertilized paddy soil using an¹⁸O–H₂O approach. Land Degradation & Development . doi:10.1002/ldr.3708

Zhu, S. S., Vivanco, J. M., & Manter, D. K. (2016). Nitrogen fertilizer rate affects root exudation, the rhizosphere microbiome and nitrogen-use-efficiency of maize. *Applied Soil Ecology*, 107, 324-333. doi:10.1016/j.apsoil.2016.07.009

Zhu, Z., Ge, T., Hu, Y., Zhou, P., Wang, T., Shibistova, O., . . . Wu, J. (2017). Fate of rice shoot and root residues, rhizodeposits, and microbial assimilated carbon in paddy soil - part 2: turnover and microbial utilization. *Plant and Soil*, 416 (1-2), 243-257. doi:10.1007/s11104-017-3210-4

Zhu, Z., Ge, T., Liu, S., Hu, Y., Ye, R., Xiao, M., . . . Wu, J. (2018). Rice rhizodeposits affect organic matter priming in paddy soil: The role of N fertilization and plant growth for enzyme activities, CO_2 and CH_4 emissions. Soil Biology and Biochemistry, 116, 369-377. doi:10.1016/j.soilbio.2017.11.001

Table 1 Total phospholipid fatty acids (PLFAs) (mg C kg⁻¹) in paddy soil under fertilization regimes at different soil depths (mean+- SE, n=3). Statistical significances based on one-way analysis of variance and Duncan's test (P < 0.05). Different lower-case letters represent significant differences in soil depth within a fertilization regime. Different uppercase letters represent significant differences in fertilization regime within soil depth. Fame: fatty acid methyl esters; CK: no fertilizer; NPK: chemical fertilizers; ST: rice straw combined with chemical fertilizers; OM: 70% NPK + 30% chicken manure. G-: Gram-negative bacteria; G+: Gram-positive bacteria.

	Soil depth					General	
Treatments	(cm)	G-	G+	Actinomycetes	Fungi	Fame	Total PLFA
СК	0-10	$13.17 {\pm} 0.50 {\rm Ca}$	$5.92{\pm}0.47\mathrm{Ca}$	$6.11 \pm 0.23 \mathrm{Ca}$	$1.96{\pm}0.10\mathrm{Ba}$	$10.58{\pm}0.56\mathrm{Ca}$	39.36 ± 1.84 Ca
	10 - 20	$5.81{\pm}0.15\mathrm{Db}$	$2.54{\pm}0.26\mathrm{Db}$	$3.33 \pm 0.10 \text{Db}$	$0.62{\pm}0.01{\rm Cb}$	$4.88{\pm}0.19\mathrm{Db}$	$17.83 \pm 0.72 \text{Db}$
	20 - 30	$1.50{\pm}0.08{\rm Cc}$	$0.79{\pm}0.03\mathrm{Dc}$	$1.11{\pm}0.01{\rm Cc}$	$0.15{\pm}0.01\mathrm{Dc}$	$1.62{\pm}0.06{\rm Cc}$	$5.32{\pm}0.16\mathrm{Cc}$
	30 - 40	$0.53{\pm}0.06\mathrm{Cd}$	$0.41{\pm}0.07\mathrm{Cd}$	$0.54{\pm}0.05{\rm Cd}$	$0.05{\pm}0.01\mathrm{Cd}$	$0.59{\pm}0.06\mathrm{Cd}$	$2.20{\pm}0.25\mathrm{Cd}$
NPK	0 - 10	$15.60{\pm}0.18\mathrm{Ba}$	$9.16{\pm}0.12\mathrm{Ba}$	$8.01{\pm}0.25\mathrm{Ba}$	$2.35{\pm}0.03\mathrm{Ba}$	$13.29{\pm}0.40\mathrm{Ba}$	$50.14 {\pm} 0.87 \text{Ba}$
	10 - 20	$9.63{\pm}0.19{\rm Cb}$	$6.08{\pm}0.33\mathrm{Cb}$	$5.66{\pm}0.20{\rm Cb}$	$1.33{\pm}0.05\mathrm{Bb}$	$8.43{\pm}0.30{\rm Cb}$	$32.40{\pm}1.01{\rm Cb}$
	20 - 30	$5.44{\pm}0.18\mathrm{Bc}$	$3.57{\pm}0.10\mathrm{Cc}$	$3.51{\pm}0.01{\rm Bc}$	$0.67{\pm}0.01{\rm Bc}$	$5.09{\pm}0.11\mathrm{Bc}$	$19.03{\pm}0.47\mathrm{Bc}$
	30 - 40	$1.12{\pm}0.03\mathrm{Bd}$	$0.98{\pm}0.01\mathrm{Bd}$	$1.07{\pm}0.01\mathrm{Bd}$	$0.13{\pm}0.00\mathrm{Bd}$	$1.27{\pm}0.04\mathrm{Bd}$	$4.85{\pm}0.09\mathrm{Bd}$
NPK +	0 - 10	$17.01{\pm}0.06\mathrm{Ba}$	$9.36{\pm}0.25\mathrm{Ba}$	$8.44{\pm}0.01\mathrm{Ba}$	$2.31{\pm}0.02\mathrm{Ba}$	$14.53{\pm}0.07\mathrm{Ba}$	$52.78 \pm 1.46 Ba$
ST							
	10 - 20	$15.32{\pm}0.41\mathrm{Bb}$	$8.79{\pm}0.10\mathrm{Ba}$	$8.05{\pm}0.13\mathrm{Ba}$	$2.11{\pm}0.14\mathrm{Aa}$	$13.88{\pm}0.12\mathrm{Ba}$	$50.50{\pm}0.79\mathrm{Ba}$
	20 - 30	$4.96{\pm}0.38\mathrm{Bc}$	$4.08{\pm}0.27\mathrm{Bb}$	$4.04{\pm}0.35{\rm Bb}$	$0.57{\pm}0.04{\rm Cb}$	$5.17{\pm}0.43\mathrm{Bb}$	$19.45 \pm 1.41 Bb$
	30 - 40	$0.93{\pm}0.12\mathrm{Bd}$	$1.04{\pm}0.10\mathrm{Bc}$	$1.00{\pm}0.12\mathrm{Bc}$	$0.11{\pm}0.02\mathrm{Bc}$	$1.04{\pm}0.10\mathrm{Bc}$	$4.02{\pm}0.56\mathrm{Bc}$
NPK+	0 - 10	$19.29{\pm}0.90\mathrm{Aa}$	$13.01{\pm}0.44\mathrm{Aa}$	$10.21{\pm}0.56\mathrm{Aa}$	$2.96{\pm}0.24\mathrm{Aa}$	$17.13 \pm 1.11 Aa$	$64.14{\pm}3.90\mathrm{Aa}$
OM							
	10 - 20	$16.98{\pm}0.64\mathrm{Ab}$	$11.12{\pm}0.37\mathrm{Aa}$	$9.55{\pm}0.30\mathrm{Aa}$	$2.26{\pm}0.03\mathrm{Ab}$	$15.30{\pm}0.60\mathrm{Aa}$	$57.47 \pm 1.90 Aa$
	20 - 30	$6.89{\pm}0.22\mathrm{Ac}$	$5.58{\pm}0.18\mathrm{Ab}$	$5.14 \pm 0.20 \text{BAb}$	$0.82{\pm}0.04{\rm Ac}$	$6.71{\pm}0.27\mathrm{Ab}$	$25.93 \pm 0.89 \mathrm{Ab}$
	30 - 40	$1.82{\pm}0.05\mathrm{Ad}$	$1.63{\pm}0.10{\rm Ac}$	$1.63{\pm}0.05{\rm Ac}$	$0.27{\pm}0.01\mathrm{Ad}$	$2.05{\pm}0.08{\rm Ac}$	$7.65{\pm}0.30\mathrm{Ac}$

Table 2 Basic physicochemical properties of paddy soil under fertilization regimes at different soil depths (mean \pm SE, n=3). CK: no fertilizer; NPK: chemical fertilizers; ST: rice straw combined with chemical fertilizers; OM: 70% NPK + 30% chicken manure; SOC: soil organic content; MBC: microbial biomass carbon; MBN: microbial biomass nitrogen. Statistical significances based on one-way analysis of variance and Duncan's test (P < 0.05). Different lower-case letters represent significant differences in soil property between soil depths within a fertilization regime. Different uppercase letters represent significant differences in a soil property between fertilization regimes within a soil depth.

Treatments	Treatments	pН	$\begin{array}{c} {\rm SOC} \ ({\rm g} \\ {\rm kg}^{-1}) \end{array}$	$\begin{array}{c} {\rm MBC} \ ({\rm mg} \\ {\rm kg}^{-1}) \end{array}$	$\begin{array}{l} \text{MBN (mg} \\ \text{kg}^{-1} \end{array}) \end{array}$	Available N $(mg kg^{-1})$	Olsen-P (mg kg ⁻¹)
СК	0-10	$7.15 \pm 0.01 \mathrm{Ad}$	$18.32 \pm 0.06 Ca$	740.40 ± 32.52	Cal.81±0.00Da	$151.30 \pm 2.20 D_{\odot}$	$a 6.57 \pm 0.56 Ca$
	10 - 20	$7.40{\pm}0.02\mathrm{Ac}$	$15.78{\pm}0.21\mathrm{Cb}$	351.35 ± 18.501	$Dt1.52\pm0.00Cb$	$108.39 \pm 1.18 C$	$b 4.93 \pm 0.17 Bb$
	20 - 30	$7.52{\pm}0.01\mathrm{Ab}$	$6.48{\pm}0.07\mathrm{Dc}$	$123.78 {\pm} 5.45 {\rm D}$	$c 0.79 \pm 0.01 Dc$	$46.71 \pm 1.18 Cc$	$5.56 \pm 0.37 \text{Cab}$
	30 - 40	$7.61{\pm}0.02\mathrm{Aa}$	$5.42{\pm}0.07\mathrm{Bd}$	59.37 ± 1.71 Cd	$0.71 \pm 0.01 \text{Cd}$	$44.92{\pm}0.77\mathrm{Bc}$	$6.48{\pm}0.14\mathrm{Ca}$
NPK	0 - 10	$6.48{\pm}0.01\mathrm{Dd}$	$17.48 {\pm} 0.52 \text{Da}$	$829.72 \pm 8.50 B$	a 1.87 ± 0.03 Ca	$181.69 \pm 5.08 C_{2}$	a 11.37 ± 0.64 Ba
	10 - 20	$6.82{\pm}0.01{\rm Cc}$	$16.13{\pm}0.04\mathrm{Cb}$	$527.14 \pm 4.60 \text{C}$	$b 1.59 \pm 0.14 Cb$	$129.40 \pm 0.80 B$	$b 9.23 \pm 0.14 Bb$
	20 - 30	$7.21{\pm}0.02\mathrm{Bb}$	$9.40{\pm}0.13\mathrm{Cc}$	219.88 ± 6.49 C	$c 1.04 \pm 0.02 Cc$	$71.29{\pm}0.89\mathrm{Bc}$	$6.80{\pm}0.75\mathrm{Cc}$
	30 - 40	$7.25 \pm 0.02 \mathrm{Ca}$	$6.09{\pm}0.00\mathrm{Ad}$	$81.66 \pm 7.02 Bd$	$0.77 \pm 0.01 \mathrm{Ad}$	$43.13 \pm 0.49 \text{Bd}$	$7.01 \pm 0.53 BCc$
NPK +	0 - 10	$6.64{\pm}0.02\mathrm{Cd}$	$22.04{\pm}0.10{\rm Ba}$	880.98 ± 27.07	Ba2.21±0.02Ba	$209.85 \pm 2.05 Ba$	a 10.38 ± 0.49 Bal
ST							
	10 - 20	$6.84{\pm}0.01{\rm Cc}$	$20.07{\pm}0.25\mathrm{Bb}$	807.86 ± 29.72	Bb1.99±0.04Bb	$185.27 \pm 0.45 \text{A}^{-1}$	$b 9.25 \pm 0.15 Bbc$
	20 - 30	$7.12 \pm 0.01 \mathrm{Db}$	$11.90{\pm}0.10\mathrm{Bc}$	259.15 ± 10.781	$Bc1.25\pm0.02Bc$	$96.77 \pm 2.49 \mathrm{Ac}$	$11.04{\pm}0.88\mathrm{Ba}$
	30 - 40	$7.19{\pm}0.02\mathrm{Da}$	$6.08 {\pm} 0.06 \mathrm{Ad}$	$62.55 \pm 3.47 BC$	$Cd0.77 \pm 0.01 Ad$	$49.84{\pm}0.45{\rm Ad}$	$8.00{\pm}0.07\mathrm{Bc}$
NPK +	0 - 10	$6.74{\pm}0.02\mathrm{Bd}$	$25.46{\pm}0.42\mathrm{Aa}$	$1019.9 \pm 52.4 \text{A}$	a 2.60 ± 0.08 Aa	$251.42 \pm 1.34 \text{Az}$	a $111.25 \pm 6.47 \text{A}$
OM							
	10 - 20	$6.93{\pm}0.01{\rm Bc}$	$24.98{\pm}0.68\mathrm{Aa}$	$869.2 \pm 22.1 \text{Ab}$	$2.50 \pm 0.01 \text{Ab}$	$193.32 \pm 6.49 \text{A}^{-1}$	$b116.50{\pm}4.40A$
	20 - 30	$7.17{\pm}0.02\mathrm{Cb}$	$12.12{\pm}0.03\mathrm{Ab}$	$421.3 \pm 1.53 Ac$	$1.32 \pm 0.01 \mathrm{Ac}$	$97.66 {\pm} 4.47 {\rm Ac}$	$56.99 \pm 2.16 \mathrm{Ab}$
	30 - 40	$7.34{\pm}0.02\mathrm{Ba}$	$5.89{\pm}0.15\mathrm{Ac}$	$165.134{\pm}2.1A$	$d 0.75 \pm 0.01 Bd$	$42.24{\pm}2.05\mathrm{Bd}$	$10.19{\pm}0.25\mathrm{Ac}$

Figure captions

FIGURE 1 Microbial community changes to the (a)ratio of fungal to bacterial phospholipid fatty acids (PLFAs); (b) ratio of gram-positive to gram-negative bacterial PLFAs; (c) ratio of gram-positive to actinomycetes PLFAs; and (d) ratio of fungal to gram-negative bacterial PLFAs in paddy soil under fertilization regimes at different soil depths. CK: no fertilizer; NPK: chemical fertilizers; NPK + ST: rice straw combined chemical fertilizers; NPK + OM: 70% NPK + 30% chicken manure. Statistical significance is based on a one-way analysis of variance and Duncan's test (P < 0.05). Different lower-case letters represent significant differences in soil depths within a fertilization regime. Different uppercase letters represent significant differences in fertilization regimes within a soil depth (n= 3)

FIGURE 2 (a) Soil respiration rate and (b) soil metabolic quotient (qCO₂) in paddy soil under fertilization regimes at different soil depths. CK: no fertilizer; NPK: chemical fertilizers; NPK + ST: rice straw combined with chemical fertilizers; NPK + OM: 70% NPK + 30% chicken manure. Statistical significance is based on a one-way analysis of variance and Duncan's test (P < 0.05). Different lower-case letters represent significant differences in soil depths within a fertilization regime. Different uppercase letters represent significant differences in fertilization regimes within a soil depth (n= 3)

FIGURE 3 Redundancy analysis of phospholipid fatty acid content from soil samples at different depths. (a) 0–40 cm; (b) 0–20 cm; (c) 20–40 cm. Shapes represent the type of fertilizer, while colours represent the different soil depths. CK: no fertilizer; NPK: chemical fertilizers; NPK + ST: rice straw combined with chemical fertilizers; NPK + OM: 70% NPK + 30% chicken manure. 0–10 cm (black), 10–20 cm (red), 20–30 cm (yellow), and 30–40 cm (orange)

