Three-dimensional hydraulic tomography analysis of long-term municipal wellfield operations: Validation with synthetic flow and solute transport data

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Abstract

This study proposes the utilization of municipal well records as an alternative dataset for large-scale heterogeneity characterization of hydraulic conductivity () and specific storage () using hydraulic tomography (HT). To investigate the performance of HT and the feasibility of utilizing municipal well records, a three-dimensional aquifer/aquitard system is constructed and synthetic groundwater flow and solute transport experiments are conducted to generate data for inverse modeling and validation of results. In particular, we simultaneously calibrate four groundwater models with varying parameterization complexity using five datasets consisting of different time durations and periods. Calibration and validation results are qualitatively and quantitatively assessed to evaluate the performance of investigated models. The estimated and tomograms from different model cases are also validated through the simulation of independently conducted pumping tests and conservative solute transport. Our study reveals that: 1) the HT analysis of municipal well records is feasible and yields reliable heterogeneous and distributions where drawdown records are available; 2) accurate geological information is of critical importance when data density is low and should be incorporated for geostatistical inversions; 3) the estimated and tomograms from the geostatistical model with geological information are capable in providing robust predictions of both groundwater flow and solute transport. Overall, this synthetic study provides a general framework for large-scale heterogeneity characterization using HT through the interpretation of municipal well records, and provides guidance for applying this concept to field problems.

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13	Key Points:
14	• Hydraulic tomography analysis of long-term municipal wellfield records is feasible
15	although data selection requires careful consideration.
16	• Continuous records with large water-level variations should be included for more
17	accurate estimation of hydraulic parameters.
18	• Results from hydraulic tomography are validated through synthetic flow and solute
19	transport experiments showing robust performance.
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21 Abstract

22 This study proposes the utilization of municipal well records as an alternative dataset for 23 large-scale heterogeneity characterization of hydraulic conductivity (K) and specific storage (S_s) 24 using hydraulic tomography (HT). To investigate the performance of HT and the feasibility of utilizing municipal well records, a three-dimensional aquifer/aquitard system is constructed and 25 26 synthetic groundwater flow and solute transport experiments are conducted to generate data for 27 inverse modeling and validation of results. In particular, we simultaneously calibrate four groundwater models with varying parameterization complexity using five datasets consisting of 28 29 different time durations and periods. Calibration and validation results are qualitatively and quantitatively assessed to evaluate the performance of investigated models. The estimated K and 30 S_s tomograms from different model cases are also validated through the simulation of 31 independently conducted pumping tests and conservative solute transport. Our study reveals that: 32 1) the HT analysis of municipal well records is feasible and yields reliable heterogeneous K and 33 S_s distributions where drawdown records are available; 2) accurate geological information is of 34 critical importance when data density is low and should be incorporated for geostatistical 35 inversions; 3) the estimated K and S_s tomograms from the geostatistical model with geological 36 37 information are capable in providing robust predictions of both groundwater flow and solute transport. Overall, this synthetic study provides a general framework for large-scale 38 39 heterogeneity characterization using HT through the interpretation of municipal well records, and provides guidance for applying this concept to field problems. 40

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42 **1. Introduction**

Planning for the optimized use and management of groundwater resources requires the 43 44 accurate characterization of subsurface heterogeneity in hydraulic conductivity (K) and specific storage (S_s) , which are two important hydraulic properties for the construction of groundwater 45 flow models, and in particular, for solute transport simulations (Ni et al., 2009; Illman et al., 46 47 2012). In the past few decades, numerous efforts have been dedicated to map the spatial distribution of K and S_s . Typically, geostatistical interpretation of small-scale K estimates 48 49 obtained from core samples, slug tests, flowmeter surveys, and single-hole pumping/injection tests is applied to map its spatial distribution (e.g., Salamon et al., 2007; Sudicky et al., 2010; 50 Alexander et al., 2011), while S_s is commonly treated to be homogeneous as its variability is 51 considered to be much less than K in natural geological formations (Gelhar, 1993). Using this 52 approach, a sufficient number of small-scale estimates is required to fully capture the 53 heterogeneity patterns of hydraulic properties (Rehfeldt et al., 1992). On the other hand, 54 Kuhlman et al. (2008) demonstrated that the interpolated K and S_s fields strongly relied on the 55 estimated small-scale values, which may be biased in representing realistic conditions due to the 56 restricted assumptions implied in analytical solutions (e.g., Theis (1935)) for these estimates. 57

An alternative approach to the geostatistical interpretation of small-scale values, hydraulic tomography (HT) was proposed and developed (e.g., Gottlieb and Dietrich, 1995; Yeh and Liu, 2000) for subsurface heterogeneity characterization. Fundamentally, the HT approach involves the inverse modeling of groundwater response data collected at various locations during a series of spatially varying pumping/injection tests. Yeh et al. (2008) concluded that the data collected in such tomographic surveys provide many constraints for model calibration, yielding more 64 accurate estimations of K and S_s fields with less uncertainty in comparison to traditional 65 characterization methods.

66 The robust performance of HT in revealing K and S_s heterogeneity has been demonstrated through numerous numerical (e.g., Yeh and Liu, 2000; Bohling et al., 2002; Zhu and Yeh, 2005), 67 laboratory (e.g., Liu et al., 2002; Illman et al., 2007, 2010, 2015; Berg and Illman, 2011a; Zhao 68 69 et al., 2015; Luo et al., 2017) and dedicated field experiments (e.g., Straface et al., 2007; Bohling et al., 2007; Illman et al., 2009; Berg and Illman, 2011b; Zha et al., 2016; Zhao and Illman, 70 71 2017). Nevertheless, most of these studies were performed at small-scale (limited to tens of square meters) sites, while only a few studies have been carried out for large-scale (several 72 square kilometers) site characterization using the approach of HT (e.g., Illman et al., 2009; Zha 73 et al., 2016, 2019). In small-scale studies, each single-well pumping/injection test is able to stress 74 the entire aquifer and generate a head response throughout the domain which can be monitored 75 with a well-designed monitoring network. However, designing and conducting HT surveys for 76 large-scale heterogeneity characterization is typically expensive, time-consuming, and 77 sometimes impractical. Illman et al. (2009) applied the approach of transient HT to characterize a 78 kilometer-scale fractured granite site at Mizunami, Japan, using data from two large-scale cross-79 80 hole pumping tests. The estimated K and S_s tomograms qualitatively agreed well with observed drawdown records, available fault information, and coseismic groundwater responses during 81 82 several large earthquakes (Niwa et al., 2012). However, they pointed out that the estimated fault 83 and fracture zones might still involve great uncertainty due to the limited hydraulic data for inverse modeling. Through the inclusion of datasets from two additional pumping tests 84 conducted at the same site, Zha et al. (2016) yielded K and S_s tomograms with improved 85 delineation of fault zones in terms of their locations and patterns. However, they indicated that 86

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with limited number of wells, the collected hydraulic data from four large-scale pumping tests
were still insufficient to map the detailed distribution of fractures and faults.

89 Instead of the traditional data collection strategy used for HT surveys, alternative datasets 90 have been proposed and utilized for large-scale heterogeneity characterization using HT (e.g., Kuhlman et al., 2008; Yeh et al., 2009; Wang et al., 2017; Zha et al., 2019). For instance, 91 92 Kuhlman et al. (2008) applied the HT approach to characterize subsurface heterogeneity (*T* and *S*) at the basin scale using head response data collected from multiple simultaneous pumping wells. 93 94 Through cycling the operation of different sets of pumping wells, the regional aquifer is repeatedly stressed to yield groundwater responses that cover most of the simulation domain. 95 They concluded that the head data collected from potentially disparate aquifer tests could be 96 97 jointly interpreted to estimate basin-wide aquifer properties using HT. It was suggested that this characterization approach could be applied to municipal or pump-and-treat wellfields with 98 existing monitoring networks. However, due to the operational requirements of municipal well 99 100 fields, it is unlikely to be able to cease pumping/injection to conduct dedicated pumping tests for aquifer characterization, thus making the application of typical HT methodologies infeasible. 101

Yeh et al. (2008) provided an opinion of using natural stimuli (e.g., river-state variations, 102 lightning, earthquake, barometric variations, storm events, etc.) as sources of excitations for 103 basin-scale subsurface characterization. Unlike traditional single-well or advanced multiple-well 104 105 (Kuhlman et al., 2008) pumping tests, natural stimuli can easily stress the aquifer to yield groundwater responses over the entire basin. They pointed out that groundwater variations at 106 different scales induced by natural stimuli with frequent and spatially varying occurrence is 107 108 analogous to that of HT surveys, and the monitored groundwater responses along with the characterized corresponding natural stimuli can be interpreted for hydraulic properties estimation. 109

110 Following this thought, Yeh et al. (2009) proposed river stage tomography as a new approach for basin-scale subsurface heterogeneity characterization. Specifically, the migration of river 111 stage perturbation along the river was treated as a natural stimuli that induces groundwater 112 fluctuations over the entire basin. The temporal and spatial variations of river stage as well as the 113 corresponding groundwater response data were then incorporated for inverse modeling to 114 estimate the spatial distribution of hydraulic properties (T and S) of the basin. The efficiency of 115 river stage tomography in revealing basin-scale hydraulic heterogeneity was later evaluated 116 117 through a field experiment conducted in Zhoushui River alluvial fan, Taiwan (Wang et al., 2017). 118 Yeh et al. (2008) contended that natural stimuli-based HT surveys should be a future direction for large-scale subsurface characterization; however, significant challenges still exist in 119 accurately characterizing the locations and strengths of natural stimuli. 120

To avoid the uncertainty associated with natural stimuli, existing hydraulic head records in a 121 wellfield with well-characterized artificial stimuli (pumping/injection operations with known 122 locations and rates) can be utilized as alternative datasets for subsurface heterogeneity 123 characterization, as suggested by Yeh and Lee (2007). Such records are typically abundant and 124 can be readily obtained from contaminant monitoring or municipal water-supply wellfields, but 125 126 they are rarely adopted for mapping the heterogeneity of hydraulic properties. Most recently, Zha et al. (2019) exploited the pump-and-treat system for subsurface heterogeneity characterization at 127 the AFP44 site located in Tucson, Arizona, US. In particular, hydraulic head changes during four 128 129 distinct events (e.g., system shutdown and resumption, changes in pumping/injection operations, and significant variations of flow rates) were extracted from the existing head records and 130 utilized for inverse modeling. Here, it should be noted that the characteristics of well 131 hydrographs might be quite different from one wellfield to another, depending on the associated 132

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pumping/injection regime. Without presenting apparent changes of hydraulic head with distinct 133 events, the well hydrographs in a municipal wellfield appear to be highly variable due to the 134 continuous operation of water-supply wells with variable flow rates. Sun et al. (2013) proposed a 135 temporal sampling strategy for HT analysis using dedicated pumping test data; however, 136 questions remain as to which data points should be extracted from the long-term head records 137 138 and utilized for subsurface heterogeneity characterization. On the other hand, the extracted hydraulic head records within selected periods are affected by prior pumping/injection operations 139 at the same site, resulting in significant difficulties in interpretation of head records due to 140 141 unknown initial condition for groundwater modeling.

In this study, a series of numerical experiments that mimic the hydraulic conditions at the 142 Mannheim East site, a municipal water-supply wellfield located in the southwest area of the city 143 of Kitchener, Ontario, Canada, was performed. In particular, a synthetic three-dimensional multi-144 aquifer/aquitard system was developed and characterized using different modeling approaches 145 146 with different head records for groundwater flow and solute transport predictions. Fundamentally, such a synthetic study with minimized sources of error (e.g., model identification and head 147 measurement) yields a general framework for subsurface heterogeneity characterization using the 148 149 existing long-term pumping/injection and water-level records.

The main objectives of this synthetic study were to: 1) explore the feasibility of utilizing municipal well data for subsurface heterogeneity characterization using HT, 2) evaluate the performance of three different modeling approaches (homogeneous, geological, and geostatistical models) for HT analyses with well data from a municipal wellfield, and 3) investigate the effect of data selection for inverse modeling. The computed K and S_s tomograms 155 from the different models were validated through the simulation of nonreactive tracer migration156 through the municipal wellfield.

157 **2. Experimental Setup**

The Mannheim East wellfield is located within the core area of the Waterloo Moraine, 158 which is classified as a kame deposit with three main aquifers separated by two glacial tills 159 (Karrow, 1993). To mimic the multi-aquifer/aquitard system of the study site, a layer-cake 160 geological model was constructed for this synthetic study, as shown in Figures 1a and 1b. The 161 size of the model was set to be 5000 m, 5000 m, and 200 m in X, Y, and Z directions, 162 respectively. In total, seven geological layers were identified beneath the study site with AT and 163 AF representing aquitard and aquifer, respectively. These layers were identified following the 164 165 conceptual hydrogeological model of the Waterloo Moraine constructed by Bajc and Shirota (2007), whereas some layers with thin thicknesses were merged and irregular layer boundaries 166 was not considered. In each geological layer, random K and S_s fields were generated by assuming 167 the Gaussian distributions of $\ln K$ and $\ln S_s$ fields with known information of their means, 168 variances, and correlation lengths using the spectral approach (Robin et al., 1993). The mean 169 values of $\ln K$ and $\ln S_s$ were obtained based on the predominant materials in each geological layer, 170 while the variances and correlation lengths were estimated according to the statistical properties 171 of spatial K and S_s distributions in natural geological formations. The generated nonstationary 172 "true" K and S_s fields for the synthetic study are illustrated as Figures 4d and 5d, respectively, 173 while the statistical details of hydraulic parameters are summarized in Table S1 of the 174 Supplementary Information section. To better evaluate the results, the entire simulation domain 175 176 was subdivided into three zones (ZONE 1, ZONE 2, and ZONE 3, as shown in Figures 1a and 1b) based on the density of well screens. 177



Figure 1: The synthetic layer-cake geological model domain along with the distribution of pumping and monitoring wells. a) and b) illustrate the plan-view and cross-section of the simulation domain, respectively; c) shows spatial distribution of assigned wells with IDs.

To mimic the hydraulic condition in a municipal water-supply wellfield, the same well configuration as the Mannheim East wellfield was applied for the synthetic study. Within the wellfield, a subdivide well site with Aquifer Storage and Recovery (ASR) system was designed to inject and store treated surface water during low water demand periods and extract the stored water during high demand periods. In total, 13 pumping/injection municipal wells screened in the water-supply aquifer (AF2) and 28 water-level monitoring wells screened at different layers were included in the model. The spatial distribution of these wells and their screens are illustrated in Figure 1. Other than the existing wells, five additional water-supply wells (AWSWs
1-5) were included for the purpose of model validation using independent pumping test data.

The synthetic model was discretized into 33,072 triangular prism elements with 18,050 nodes, as shown in Figure S1 of the Supplementary Information section. The mesh was refined around wells, but became coarser when moving towards boundaries. The four lateral boundaries of the model were set as constant head boundaries of 340 m, while the top and bottom boundaries were set as no-flow. Transient groundwater flow was then considered for the generation of synthetic head data, and the governing equation can be expressed as:

$$\nabla \cdot [K(\mathbf{x})\nabla h] + Q(\mathbf{x}_p) = S_s(\mathbf{x})\frac{\partial h}{\partial t}$$
(1)

197 subject to initial and boundary conditions:

$$h|_{t=0} = h_0, \ h|_{\Gamma_1} = h_1, \ and \ [K(\mathbf{x})\nabla h] \cdot \mathbf{n}|_{\Gamma_N} = q$$
 (2)

where, in Eq. (1), ∇ is the gradient operator, $K(\mathbf{x})$ is hydraulic conductivity (L T⁻¹), h is hydraulic head (L), $Q(\mathbf{x}_p)$ is the rate of pumping per unit volume (T⁻¹) at location \mathbf{x}_p , and $S_s(\mathbf{x})$ is specific storage (L⁻¹). In Eq. (2), h_0 represents the initial hydraulic head, h_1 is a constant head (L) at boundary Γ_1 , q is the specific discharge (L T⁻¹) at the Neumann boundary Γ_N , and \mathbf{n} is a unit vector normal to Γ_2 .

In this study, the transient flow equation was solved using the forward simulation code HydroGeoSphere (HGS) (Aquanty, 2019) to generate synthetic head data for analyses. In particular, variable pumping/injection records in all 13 water-supply wells during the years of 2012 and 2013 (Figure 2) were extracted from Water Resources Analysis System (WRAS+) (Regional Municipality of Waterloo, 2014) database and included in the forward model. 208 Simulated head values were report at all 28 monitoring well locations during the year of 2013 (as 209 shown in Figure S2 of the Supplementary Information section) for the analyses presented in this study. The purpose of including pumping/injection information prior to the observation data is to 210 211 mimic the pumping history of the system prior to the calibration period, which leads to uncertain initial conditions at the beginning of observation data. In addition to municipal well data, 212 dedicated pumping test data from additional water-supply wells (AWSWs 1-5) were generated as 213 independent pumping test data and utilized for model validation. In particular, a constant 214 pumping rate of 8,000 m³/day was assigned to each additional well, and drawdown data in all 28 215 monitoring wells were simulated (as shown in Figure S3 of the Supplementary Information 216 section). The utilization of these independent pumping test data is to assess the ability of the 217 obtained hydraulic parameter (K and S_s) fields in guiding the construction of new water-supply 218 219 wells at the study site.



Figure 2: Extraction (positive) and injection (negative) pumping rate records at all 13 municipal
wells during the years of 2012 and 2013 from the WRAS+ database.

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3. Data Utilized for Inverse Modeling

Instead of including all simulated head data (0.1-day interval) for analysis, daily observation 224 data at the beginning of each day (12:00 am) are extracted and utilized for model calibration (0 -225 120 days) and validation (180 - 365 days). To investigate the effect of data selection on inverse 226 modeling, five datasets with different durations and periods are selected for model calibration in 227 the synthetic study, as shown in Figure 3, with their properties summarized in Table S2. In 228 particular, Dataset A includes daily observation data in all 28 monitoring wells during the first 30 229 days, while Datasets B and C extend the simulation durations to 60 and 120 days, respectively. 230 231 Daily observation data during the second 30 days are extracted as Dataset D, which shares the same simulation duration as Dataset A, but with a relatively small magnitude of water-level 232

variations. Dataset E has the same simulation duration as Dataset C (120 days); however, instead
of incorporating all observation points, only the periods with large water-level variations are
selected and utilized for model calibration.

236 As mentioned previously, the interpretation of municipal well data still suffers an issue of uncertain initial conditions for groundwater modeling due to the continuous operation of 237 238 municipal water-supply wells. The effect of uncertain initial condition on groundwater modeling 239 has been investigated by Yu et al. (2019). Based on their results, the proposed spin-up method is adopted here to minimize the effect of uncertain initial conditions. In particular, 240 pumping/injection rate records prior to the observation data are utilized for model spin-up and 241 incorporated for model calibration. The model spin-up time is determined by incorporating 242 different lengths of prior pumping/injection records for forward simulations with known 243 hydraulic parameter (K and S_s) fields. The simulated head variations at monitoring locations are 244 then compared quantitatively to the observed ones (Figure S1), with the comparison results 245 illustrated in Figure S4 of the Supplementary Information section. Results reveal that the 246 discrepancy between simulated and observed head data decreases significantly as the spin-up 247 period increases and stabilizes in magnitude after incorporating pumping/injection records 180 248 249 days prior to the observation data. As a result, pumping/injection rate records for 180 days prior to the observation data are extracted and incorporated for model calibration in this synthetic 250 251 study.



Figure 3: Selected datasets for model calibration. Dataset A (0 – 30 days), Dataset B (0 – 60 days), Dataset C (0 – 120 days), Dataset D (30 – 60 days), Dataset E (0 – 120 days with selected large drawdown variation periods).

4. Groundwater Flow Modeling Approaches

257 **4.1 Cas**

4.1 Case 1: Effective Parameter Model

The synthetic multi-aquifer/aquitard system is first characterized as a homogeneous, isotropic 258 medium to estimate the effective K and S_s values by coupling the groundwater flow model HGS 259 (Aquanty, 2019) with the parameter estimation code PEST (Doherty, 2005), and is referred to as 260 the 'effective parameter' model. The effective parameter model provides zero-resolution on 261 subsurface heterogeneity; however, it may still able to describe the overall behavior of 262 groundwater flow in the system. Furthermore, the estimated effective K and S_s values can be 263 used as the initial estimate of hydraulic parameters to guide the calibration of more sophisticated 264 groundwater flow models. For each dataset, an optimal set of K and S_s is estimated by 265

simultaneously matching all data points. The initial values of *K* and S_s input into PEST are 0.99 m/day and 1.88×10^{-4} /m, respectively, which are the calculated geometric means of the entire "true" *K* and S_s fields.

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9 4.2 Case 2: Geological Model

The response from the synthetic multi-aquifer/aquitard system is then used to calibrate the 270 geological model, which is normally adopted for groundwater flow modeling at large scales (e.g., 271 regional or basin scales). In this approach, each geological layer is characterized as a 272 homogeneous, isotropic medium, and a uniform set of K and S_s is estimated and assigned to 273 describe its hydraulic properties. The effect of the accuracy of constructed geological models on 274 inverse modeling has been previously investigated through sandbox experiments (Zhao et al., 275 276 2016; Luo et al., 2017). To avoid the uncertainty associated with model identification, it is assumed that the hydrostratigraphic contacts are perfectly known for the geological model. In a 277 similar fashion to the effective model, the geological model is calibrated using PEST coupled 278 279 with HGS by simultaneously matching all data points. For each geological layer, the geometric means of K and S_s from the "true" fields are utilized as initial guesses of hydraulic parameters for 280 model calibration. Additional cases were conducted by using the calibrated effective model as 281 initial K and S_s guesses for geological model calibration; however, unrealistic values of hydraulic 282 parameters were obtained in layers where no hydraulic head data was available, a phenomenon 283 284 note in previous studies (Berg and Illman, 2011b).. As a result, the heterogeneous K and S_s fields based on the stratigraphic information are applied as initial guesses. In total, 14 parameters are 285 estimated for subsurface heterogeneity characterization using the geological model. 286

287 **4.3 Case 3: Geostatistical Models**

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288 As a third case, the head response to the synthetic municipal well data are interpreted with highly parameterized geostatistical models for subsurface heterogeneity characterization. All 289 geostatistical inversions are conducted using the Simultaneous Successive Linear Estimator 290 (SimSLE), developed by Xiang et al. (2009) and modified for this study to account for variable 291 pumping/injection records. In this study, geostatistical inversion using SimSLE assumes a 292 293 transient groundwater flow field, and the natural logarithm of K and S_s are both treated as multi-Gaussian, second-order stationary, stochastic processes. Other than estimating the spatial 294 distributions of hydraulic parameters (K and S_s tomograms), SimSLE also provides the variance 295 296 maps of $\ln K$ and $\ln S_s$ to describe the uncertainty of estimated values, with large variance meaning high uncertainty in the estimated parameters and vice versa. 297

Based on the differences in initial K and S_s fields, two geostatistical inversion cases are 298 investigated. For Case 3a, homogeneous initial K and S_s fields are used for model calibration, 299 representing the scenario of calibrating hydraulic data only. In this case, the K and S_s values 300 obtained from the effective parameter model (Case 1) are utilized as initial guesses and assigned 301 to the entire domain. For Case 3b, geological information is incorporated for model calibration. 302 Geostatistical inversions in this case start from heterogeneous initial K and S_s fields which are the 303 304 same as those utilized for geological model calibration. For both cases, the variances of lnK and $\ln S_s$ ($\sigma^2_{\ln K}$, $\sigma^2_{\ln Ss}$) are initially set to be 4.0 and 2.0, respectively, while the correlation scales are 305 set to be $\lambda_x = 400$ m, $\lambda_y = 400$ m, and $\lambda_z = 5$ m for both K and S_s. Due to the fact that the 306 307 statistical properties of heterogeneous K and S_s fields are commonly unknown for field studies, 308 the input properties for geostatistical inversions are set to be different from those utilized for "true" K and S_s fields generation (as shown in Table S1 of the Supplementary Information 309 section). 310

311 **5. Results and Discussion**

In this study, five datasets (Datasets A-E) are interpreted with four different models (Cases 1, 312 313 2, 3a and 3b). Results from all investigated models are summarized and examined. In particular, K and S_s values estimated from different models are first compared to the "true" fields to 314 illustrate the accuracy of these estimates. Then, calibration and validation results are assessed 315 316 qualitatively and quantitatively by plotting scatterplots of simulated versus observed head variations and evaluating model errors, respectively. The evaluation of model errors is performed 317 by computing the mean absolute error (L_1) , mean square error (L_2) , and coefficient of 318 determination (R^2) between simulated and observed head values using: 319

$$L_1 = \frac{1}{n} \sum_{i=1}^{n} |x_i - \hat{x}_i|$$
(3)

$$L_2 = \frac{1}{n} \sum_{i=1}^{n} (x_i - \hat{x}_i)^2$$
(4)

$$R^{2} = \left[\frac{\frac{1}{n}\sum_{i=1}^{n}(x_{i}-\mu_{x})(\hat{x}_{i}-\mu_{\hat{x}})}{\sqrt{\frac{1}{n}\sum_{i=1}^{n}(x_{i}-\mu_{x})^{2}\frac{1}{n}\sum_{i=1}^{n}(\hat{x}_{i}-\mu_{\hat{x}})^{2}}}\right]^{2}$$
(5)

where *n* is the total number of head data, x_i and \hat{x}_i represent *i*th simulated and observed head data, respectively, μ_x and $\mu_{\hat{x}}$ represent averaged simulated and observed head data, respectively. These values were calculated to quantitatively analyze the discrepancy and correspondence between the simulated and observed head data.

In the following sections, calibration and validation results associated with Dataset A are first presented to evaluate the performance of different models in revealing large-scale heterogeneities and predicting groundwater flow, while the summarized results associated with other datasets (Datasets B-E) are provided in the Supplementary Information section. Then, statistical summary (L_1 , L_2 , and R^2) of the validation results obtained from all investigated models is presented to show the effect of data selection on inverse modeling.

5.1 Model Calibration

Through the interpretation of Dataset A, the effective parameter model (Case 1) yields K and 331 S_s estimates as well as their 95% confidence intervals of $K = 9.87 \pm 0.14$ m/day and $S_s = 2.56 \times$ 332 $10^{-4} \pm 2.5 \times 10^{-5}$ /m. Compared to the calculated geometric means from the "true" fields (0.99 333 m/day and 1.88×10^{-4} /m for K and S_s, respectively), the effective K and S_s obtained from the 334 municipal well data are more representative to the effective hydraulic parameters of the layers 335 where most monitoring wells are screened (AF1, AT2, and AF2). This implies that more 336 observation ports in upper and lower geological layers are required to obtain unbiased effective 337 hydraulic parameters for the entire multi-aquifer/aquitard system. 338

Figure 4 illustrates the obtained *K* tomograms from the geological (Case 2) and geostatistical models (Cases 3a and 3b) through the interpretation of Dataset A. The "true" *K* field is included on the bottom right as a reference for comparison. Figure 5 illustrates the same, but for the estimated S_s tomograms.



Figure 4: Estimated K tomograms from three model cases through the interpretation of Dataset A as well as the "true" K field. a) Case 2: geological model, b) Case 3a: geostatistical model without geological information, c) Case 3b: geostatistical model with geological information, and d) "true" K field. In each contour map, small black circles represent the location of monitoring well screens.



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Figure 5: Estimated S_s tomograms from three model cases through the interpretation of Dataset A as well as the "true" S_s field. a) Case 2: geological model, b) Case 3a: geostatistical model without geological information, c) Case 3b: geostatistical model with geological information, and d) "true" S_s field. In each contour map, small black circles represent the location of monitoring well screens.

As shown in Figures 4a and 5a, the estimated K and S_s values from the geological model (Case 2) are found to roughly describe the average hydraulic properties of each geological layer in comparison to the "true" field (Figures 4d and 5d for K and S_s , respectively). Here, it should be noted that geological models in this synthetic study are calibrated with known stratigraphic

359 information and well estimated initial K and S_s values (the geometric means of random K and S_s fields in each geological layer). However, typically there is significant uncertainty in the 360 geological model constructed with sparse boreholes and utilized as prior information. To further 361 evaluate these estimates, natural logarithm of these K and S_s estimates as well as their 95% 362 confidence intervals are plotted as Figure S13 of the Supplementary Information section. Results 363 364 reveal that the narrowest confidence intervals of K estimates are obtained in the water-supply aquifer (AF2), where all pumping/injection and monitoring wells are screened, suggesting the 365 high confidence of the K estimate of this layer. In contrast, when we examine the upper and 366 367 lower layers, the confidence of K estimates decreases resulting in larger confidence intervals due to fact that fewer observation data are available in these layers for estimating reliable K values. 368 This is in line with the conclusion provided by Luo et al. (2017) that when using a zonation 369 model for subsurface characterization, hydraulic head data in each identified zones are required 370 to yield reliable estimates of hydraulic parameters of these zones. 371

372 In comparison to the K estimates, the estimated S_s values are found to have larger confidence intervals, suggesting higher uncertainty associated with these S_s estimates. This may be attributed 373 to the fact that daily observation data are extracted and interpreted for hydraulic parameters 374 375 estimation in this study. The utilization of such data points ignores early-time water-level variations right after the change of pumping/injection rates which are of critical importance for 376 obtaining reliable S_s estimates (Sun et al., 2013). The interpretation of datasets in a denser 377 378 fashion (e.g., hourly observation points) may improve the estimation of S_s . However, due to the computationally intensive nature of geostatistical inversions, such a scenario of including a dense 379 dataset was not included in this study. 380

381 The obtained K and S_s tomograms from the geostatistical models are illustrated as Figures 4b -4c and Figures 5b - 5c, respectively. As shown in Figure 4b, the geostatistical inversion of 382 hydraulic head data only (Case 3a) is able to reveal heterogeneity details, where wells are 383 concentrated with sufficient head data. However, the estimated K tomogram is found with great 384 loss of heterogeneity details in comparison to the "true" K field. Although some major zones are 385 386 delineated, the overall smooth patterns fail to capture the precise shapes of stratigraphic features. Different from the K tomogram, the S_s tomogram estimated from Case 3a does not show any 387 distinct heterogeneity details, as shown in Figure 5b. This result again implies that the selected 388 389 head data for model calibration are restrictive for S_s estimations. After incorporating the geological information for geostatistical inversion (Case 3b), significant improvement in 390 revealing heterogeneity details is observed for both K and S_s tomograms, as shown in Figures 4c 391 and 5c, respectively. In particular, greater detail in K heterogeneity is revealed within the water-392 supply aquifer (AF2), resulting the spatial distribution of K in this layer comparable to that in the 393 "true" field. For upper and lower layers, the loss of heterogeneity details is still observed due to 394 the lack of hydraulic information in these layers. We believe that the estimated K tomogram can 395 further be enhanced if more monitoring wells are available for head response records at different 396 397 layers. The improvement in the S_s tomogram is not as distinct as that in the K tomogram; however, slight patterns of S_s heterogeneities are still revealed after the incorporation of 398 geological information, as shown in Figure 5c. 399

The estimated *K* and S_s values from geostatistical models (Case 3a and 3b) are then evaluated by analyzing the uncertainty associated with these estimates and comparing them to the "true" values. For uncertainty analysis, the corresponding $\ln K$ and $\ln S_s$ variance maps are plotted (as shown in Figure S18 of the Supplementary Information section), with larger variances indicate 404 higher uncertainty of the estimates. For both cases, relatively small lnK variances are obtained in the central area of the simulation domain, where wells are concentrated for hydraulic head data, 405 while variances become larger when moving away from the wells. In general, the $\ln S_s$ variances 406 are computed to be larger than those of $\ln K$, suggesting the higher uncertainty of these S_s 407 estimates in comparison to the K estimates. This may again attribute to the factor of the temporal 408 409 resolution of observation data for model calibration, as discussed above. The estimated K and S_s values are then compared to the "true" values by plotting the scatterplots of corresponding 410 estimated versus "true" $\ln K$ and $\ln S_s$ values (as shown in Figure S23 of the Supplementary 411 412 Information section). Comparison results reveal that the geostatistical inversion of hydraulic head only (Case 3a) is still able to yield relatively reliable K and S_s estimates in the area with 413 sufficient hydraulic head data (ZONE 1), while large discrepancies of these estimates are 414 observed in ZONEs 2 and 3. After incorporating the stratigraphic information, significant 415 improvements are observed for both K and S_s estimates in all three zones. These results indicate 416 that the municipal well data can be used to characterize subsurface heterogeneity with HT 417 methods. However, since such hydraulic data are typically concentrated in the pumping area, 418 accurate stratigraphy information is of critical importance for geostatistical inversions to 419 420 accurately reveal heterogeneity patterns and yield reliable estimates of hydraulic parameters. Earlier studies by Zhao et al. (2016) and Luo et al. (2017) have shown that the inclusion of 421 inaccurate stratigraphy information will have deleterious impacts on parameter estimates. 422

The performance of four different models are then assessed qualitatively and quantitatively by plotting the scatterplots of calibration results, as shown in Figure 6. In each scatterplot, data points corresponding to three subdivided zones are distinguished with different colors. A linear model that fits all data points is provided along with the corresponding coefficient of 427 determination (R^2) , as well as calculated L_1 and L_2 norms. Examination of Figure 6 reveals that 428 the calibration results in terms of head data matching improve when a larger number of estimated 429 parameters are accounted for inverse modeling (from Case 1 to Cases 3). This makes sense since 430 the highly parameterized geostatistical model allows for the adjustment of *K* and *S_s* estimates in 431 each element to fit the observation data. After incorporating the geological information, the 432 geostatistical model (Case 3b) yields the best fit of simulated and observed head variations 433 (Figure 6d).



434

Figure 6: Calibration scatterplots (Dataset A) of simulated versus "observed" drawdowns for
four model cases. a) Case 1: effective parameter model, b) Case 2: geological model, c) Case 3a:

geostatistical model without geological information, d) Case 3b: geostatistical model withgeological information.

439 **5.2 Model Validation**

Model validation in this study is performed in two scenarios. For Scenario 1, the municipal 440 well data during the second half year of 2013 are utilized for model validation. Specifically, the 441 obtained K and S_s tomograms are applied to continuously predict head variations using the same 442 well configuration (water-supply and monitoring wells) as that for model calibration. For 443 444 Scenario 2, the independent pumping test data obtained from additional water-supply wells (AWSWs 1-5) that not used in the calibration effort are utilized for model validation. The 445 validation scatterplots of different model cases associated with Dataset A are illustrated in 446 447 Figures 7 and 8, for Scenarios 1 and 2, respectively.

Examination of Figure 7 reveals that when the municipal well data are utilized for model 448 validation (Scenario 1), the performances of different model cases share the same order as the 449 calibration results. In particular, Case 3d (Figure 7d) performs the best in continuously predicting 450 drawdown variations for the entire domain, followed by Case 3a (Figure 7c), while Case 1 yields 451 the worst prediction results in terms of bias and scatter. Case 2 is found able to adequately 452 predict drawdown variations at monitoring locations in all subdivided zones; however, the lack 453 information of intralayer heterogeneity resulted in relatively large scatter between the simulated 454 455 and observed head variations. It is of interest to note the K and S_s tomograms obtained from the geostatistical model without geological information (Case 3a) reveal greater loss of 456 heterogeneity details (as shown in Figures 4b and 5b for K and S_s , respectively); however, they 457 458 could still be applied to yield adequate predictions of head variations in all monitoring wells. This is because the data utilized for validation in Scenario 1 share the information of well 459

460 configuration as the datasets utilized for model calibration, thus making these validation results461 biased for the assessment of the performance of different models.



Figure 7: Validation scatterplots (Dataset A) of simulated versus "observed" municipal well data
(Scenario 1) for four model cases. a) Case 1: effective parameter model, b) Case 2: geological
model, c) Case 3a: geostatistical model without geological information, d) Case 3b: geostatistical
model with geological information.





Observed Drawdown (m)

Figure 8: Validation scatterplots (Dataset A) of simulated versus "observed" independent 468 pumping test data (Scenario 2) for four model cases. a) Case 1: effective parameter model, b) 469 Case 2: geological model, c) Case 3a: geostatistical model without geological information, d) 470 Case 3b: geostatistical model with geological information. 471

To ensure a more credible validation of the different models, independent pumping test data 472 that was not used in the calibration effort are utilized for model validation (Scenario 2), as 473 suggested by Illman et al. (2010). As shown in Figure 8, Case 1 (Figure 12a) still performs the 474 worst in predicting drawdowns from the independent pumping tests. However, it is surprising to 475 find that Case 2 (Figure 8b) provides much better prediction results in comparison to Case 3a 476 477 (Figure 8c), especially for monitoring wells screened in ZONEs 2 and 3. The comparison result

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478 reveals that the stratigraphic information becomes increasingly important for subsurface heterogeneity characterization when fewer hydraulic head data are available for inverse 479 modeling, which is again in line with the conclusion provided by Luo et al. (2017). After 480 incorporating geological information, Case 3b (Figure 8d) yields the best prediction results with 481 the highest correlation and smallest discrepancy between simulated and observed drawdowns in 482 483 comparison to other model cases. The validation results associated with Case 3b reveal that the K and S_s tomograms obtained from the geostatistical model with geological information cannot 484 only be used to predict water-level variations in the existing municipal wells, but also guide the 485 486 construction of new water-supply wells.

The calibration and validation results presented above reveal that stratigraphic information is 487 of critical importance for large-scale site characterization using the municipal well data. The 488 calibrated geological model yields relatively adequate predictions of water-level variations 489 induced by both the existing (Scenario 1) and additional (Scenario 2) water-supply wells, while 490 remarkable improvements in prediction results are observed when accurate geological 491 information was incorporated for geostatistical inversions. However, it should be noted that the 492 stratigraphic information adopted here is extracted from the synthetic model with no error. 493 494 Following the conclusion provided by Zhao et al. (2016) and Luo et al. (2017), close attention should be paid in constructing accurate geological model when using the actual municipal well 495 data for site characterization. 496

497

7 5.3 Effect of Data Selection on Inverse Modeling

498 To investigate the effect of data selection on inverse modeling, the statistical properties (L_1 , 499 L_2 , and R^2) of the validation results from all investigated models are computed and plotted in 500 Figure 9 with all values summarized in Table S3 of the Supplementary Information section. In 501 general, when different datasets are included for model calibration, the effective parameter 502 model (Case 1) always performs the worst in predicting groundwater flow, while the 503 geostatistical model with geological information (Case 3b) always performs the best. On the 504 other hand, the performance of the geological model (Case 2) and the geostatistical model 505 without geological information (Case 3a) vary from one dataset to another.



Figure 9: Statistical Summary (L_1 , L_2 , and R^2) of validation results for four model cases when different datasets were incorporated for model calibration. a) municipal well data (Scenario 1), b) independent pumping test data (Scenario 2).

510 When more observation points with longer simulation durations are included for model 511 calibration (from Dataset A to Dataset C), the estimated *K* and S_s tomograms from Case 2 show 512 distinct improvement in continuously predicting municipal well data (Scenario 1, as shown in 513 Figure 9a) in terms of computed L_1 , L_2 , and R^2 values. Such improvement is not observed for the 514 prediction of independent pumping test data (Scenario 2, as shown in Figure 9b); however, 515 slightly better prediction results are still obtained when using Dataset C for the geological model calibration. It is interesting to find that Case 3a behaves oppositely to Case 2, in which, worse 516 validation results are obtained for Case 3a after increasing the simulation duration for model 517 calibration. This may be attributed to the fact that with longer simulation durations, 518 pumping/injection influence from the water-supply wells propagates to an area beyond the 519 520 production area, resulting observation data in monitoring wells affected by the heterogeneity of K and S_s in a greater area without any hydraulic information. When interpreting municipal well 521 data with long simulation durations, the calibration of geostatistical models using hydraulic head 522 523 only (Case 3a) is likely solving ill-posed inverse problems, yielding inaccurate estimation of hydraulic parameters. Dataset D is selected to have the same simulation of Dataset A, but with 524 much smaller magnitude of head variations. Results reveal that the validation results associated 525 526 with Dataset D are distinctly worse in comparison to those associated with Dataset A for all model cases, implying that the periods with large water-level variations should be included when 527 interpreting the municipal well data for site heterogeneity characterization. Dataset E shares the 528 same simulation duration as Dataset C, but only the periods with large water-level variations are 529 utilized for model calibration. In comparison to Dataset C, Dataset E yields slightly worse 530 531 validation results for all model cases. This may be the case because the analysis presented in this study aims to estimate hydraulic parameters using long-term pumping/injection and water-level 532 records. Instead of using the periods with large head variations only, continuous data points 533 534 should be included to accurately describe the overall trends of water-level variations in monitoring wells. 535

The results presented above reveal that the effects of data selection on inverse modeling aredifferent for different modeling approaches. Through the comparison of the validation results

538 from all investigated models, the geostatistical model with geological information (Case 3b) is suggested to interpret continuous head data with large variations for large-scale heterogeneity 539 characterization. However, new approaches need to be developed for big data synthesis and 540 intelligent data selection for inverse modeling. 541

542

6. Solute Transport Prediction

One remaining question is whether the estimated K and S_s tomograms from the municipal 543 well data can be applied to predict solute transport. To investigate this issue, additional model 544 runs are performed by simulating solute transport using the estimated K and S_s tomograms. 545 Results are then compared to the scenario simulated using the "true" K and S_s fields to evaluate 546 the performances of these K and S_s estimates in predicting solute transport. For this investigation, 547 548 the estimated K and S_s tomograms from four model cases through the interpretation of Dataset A are utilized. 549

550

6.1 Solute Transport Simulation

To simulate solute transport, a point source of the conservative solute chloride (Cl) was 551 added into the synthetic system, located in the central area of layer AF1 with X, Y, and Z equal 552 553 to 2,750 m, 2,750 m, and 175 m, respectively. The source was assigned with a constant Cl concentration of 1,000 mg/L and removed after 50 years of simulation. The dispersivities of the 554 system were assumed to be 20 m, 5 m, and 0.02 m for longitudinal, transverse, and vertical 555 transverse directions, respectively. The porosity was assigned to be 0.4 throughout the simulation 556 domain. To mimic the migration of plume under real conditions, regional groundwater flow was 557 558 accounted for in the solute transport simulation, in which groundwater was considered to flow from northwest to southeast with a hydraulic gradient of 0.0014. The influence of municipal 559

water-supply pumping was also accounted for by assigning a constant pumping rate in each water-supply well based on its corresponding rate records. A slightly modified form of conventional advection-dispersion equation was adopted in this study for solute transport simulation, following the work of Burnett and Frind (1987). Specifically, it accounts transverse dispersivities at both horizontal and vertical directions and can be expressed as:

$$-\nabla \cdot (\mathbf{q}C - \theta_s \mathbf{D}\nabla C) \pm Q_c = \theta_s \frac{\partial C}{\partial t}$$
(6)

$$D_{ij} = [\alpha_L - \alpha_T] \frac{v_i v_j}{|\mathbf{v}|} + \alpha_T |\mathbf{v}| \delta_{ij} + D_0 \delta_{ij}, \qquad \text{where } i, j = x, y, z$$
(7)

subject to initial and boundary conditions:

 $C(\mathbf{x},t)|_{t=0}=C_0,$

566
$$C(\mathbf{x},t)|_{\Gamma_D} = C_D, [\theta_s \mathbf{D} \nabla C] \cdot \mathbf{n}|_{\Gamma_N} = 0, \text{ and } [-\mathbf{q}C + \theta_s \mathbf{D} \nabla C] \cdot \mathbf{n}|_{\Gamma_C} = \mathbf{q}C_0$$
 (8)

where in Eq. (1), $\mathbf{q} = -K(\mathbf{x})\nabla h(\mathbf{x})$ is the flux (L T⁻¹) and θ_s is the effective porosity (dimensionless). C 567 is the solute concentration (M L^{-3}), and Q_c is the rate at which solutes are injected (source) or 568 extracted (sink) (M $L^{-3} T^{-1}$). **D** is the macrodispersion coefficient ($L^{2} T^{-1}$) evaluated from velocity 569 and dispersivities, as shown in Eq. (2). α_L and α_T are longitudinal and transverse dispersivity (L² 570 L⁻¹), respectively. v_i and v_i are velocities (L T⁻¹) at different directions, and |**v**| is the magnitude 571 of the velocity. D_0 is the effective molecular diffusion coefficient (L² T⁻¹), and δ_{ij} is the identity 572 tensor. In Eq. (3), C_0 is the initial concentration in the entire system, C_D is the specified 573 concentration at the Dirichlet boundary (Γ_D), no dispersive flux is applied at the Neumann 574 boundary (Γ_N), and $\mathbf{q}C_0$ is the mass flux (M L⁻² T⁻¹) at the Cauchy boundary (Γ_C). 575

In this investigation, solute transport within the domain is simulated with HGS, using the "true" and estimated K and S_s fields. The total simulation duration is set to be 300 years. The performance of the different model cases in predicting solute transport are then assessed by comparing simulation results in terms of plume patterns, Cl concentrations at sampling locations, breakthrough curves and their temporal moments.

581 **6.2 Simulation Results**

Figure 10 illustrates the contour maps of the Cl plume simulated for the four model cases 582 along the cross-section Northwest-Southeast at four selected times: Year 5 (early time), Year 50 583 (source removal), Year 100 (peak concentration arrival), and Year 300 (late time). The simulated 584 Cl plumes associated with the "true" K and S_s fields are also included at the bottom for the 585 purpose of comparison. The outer bound of these plumes is set to be 1×10^{-6} . Examination of 586 Figure 10 reveals that Case 3b provides the best prediction results (Figure 10d), yielding Cl 587 plumes quite similar to the observed ones (Figure 10e) at all time stages. Without incorporating 588 589 the stratigraphic information for inverse modeling, Case 1 and Case 3a fail to capture the migration of Cl (shown as Figures 10a and 10c, respectively), especially at the early and late 590 time stages. It is surprising to find that even with known stratigraphic information, Case 2 yields 591 the worst prediction results (Figure 10b) in comparison to other investigated model cases. This 592 may be attributed to the inaccurate estimation of hydraulic parameters (K and S_s) in the source 593 594 layer (AF1), where few hydraulic head data are available for model calibration. The simulated Cl 595 plume using the calibrated geological model is found to be distinctly enlarged with the presence of source (Years 5 and 50), but rapidly diluted after the removal of source (Years 100 and 300). 596 597 These results reveal that solute transport is strongly impacted by the heterogeneity of hydraulic parameters (K and S_s), and the accurate estimation of K and S_s values, as well as their spatial 598

distributions are of critical need for the adequate prediction of solute migration in subsurfaceconditions.



Figure 10: Simulated Cl plumes at four different time stages for four model cases and using the
"true" K and S_s fields. a) Case 1: effective parameter model, b) Case 2: geological model, c)
Case 3a: geostatistical model without geological information, d) Case 3b: geostatistical model
with geological information, and e) "true" K and S_s fields.





Figure 11: Scatterplots of simulated and observed Cl concentrations at all wells for four model
cases at four time stages: a) Year 5 (early time); b) Year 50 (source removal); c) Year 100 (peak
concentration arrival); d) Year 300 (late time).

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610 The simulation results are then assessed by plotting the scatterplots of simulated versus observed Cl concentrations at water-supply and monitoring wells (sampling points) to visualize 611 the spatial distribution of errors in terms of bias and scatter, as shown in Figure 11. Examination 612 of Figure 11 reveals that prediction results for all model cases are improved from the early time 613 stage to peak concentration arrival (Figures 11a through 11c) with all data points approaching the 614 45° line. This makes sense since more heterogeneity information is captured when the plume 615 extends to a larger area, and the heterogeneous system behaves more like a homogeneous model 616 with effective hydraulic parameters for solute transport prediction. After the removal of the 617 source, the impact of heterogeneity in hydraulic properties on solute transport enhanced again, 618 resulting biased prediction results with enlarged scatters for all model cases at the late time stage 619 (Figure 11d). These results reveal that the heterogeneity of hydraulic parameters (K and S_s) 620 621 would strongly impact the removal of solute from the subsurface and should be accurately characterized for site contaminant remediation. 622

623

6.3 Breakthrough Curves

Figure 12 illustrates the breakthrough curves of Cl concentration at three selected sampling points (M8b, M5a, and M4 located in ZONEs 1, 2, and 3, respectively) for four model cases as well as the "true" K and S_s fields. The breakthrough curves of Cl concentration for all sampling points are illustrated in Figure S40 of the Supplementary Information section. In each plot, the "true" breakthrough curve is illustrated as the black dash line, while the simulated ones from different model cases are illustrated as solid lines with different colors.



Figure 12: Simulated and observed breakthrough curves of Cl concentration at selected sampling
locations (one for each subdivided zone) for four model cases.

As shown in Figure 12a, the *K* and S_s tomograms obtained from the geostatistical model with geological information (Case 3b) can be utilized to adequately capture the behavior of solute transport, yielding the simulated breakthrough curve at the sampling points M8b be consistent with the "true" one. In contrast, the geological model (Case 2) yields quite poor prediction result, with much higher peak concentration, earlier arrival time, and shorter late-time tail in comparison to the "true" breakthrough curve. This is the case for most sampling points located

639 within ZONE 1. In ZONE 2, where hydraulic head data are lacking for inverse modeling, Case 3b yields slightly biased prediction results at the late-time simulation period, as shown in Figure 640 12b. Nevertheless, it still performs the best in predicting solute transport in comparison to other 641 model cases. For ZONE 3, the sampling point (M4) located at the bottom layer (Bedrock) is 642 selected and the corresponding breakthrough curves are compared, as shown in Figure 12c. 643 644 Without incorporating geological information for model calibration, Cases 1 and 3a yield significantly enhanced Cl concentrations at the bottom of the simulation domain. In the 645 following section, temporal moment analyses are performed to quantitatively compare the 646 simulated breakthrough curves to the "true" ones. 647

648 6.4 Temporal Moment Analysis

Instead of characterizing the breakthrough curves at all wells, two sampling points (M4 and M8a) at the bottom layer (Bedrock) are excluded for temporal moment analysis since the Cl plume is simulated mainly present in the upper layers. The *n*th temporal moments (M_n) of Cl concentration at location (x, y, z) at time (t) are given by:

$$M_n = \int_0^\infty t^n C(x, y, z, t) dt$$
(9)

where *t* is the time, and *C* is the Cl concentration. The zeroth (M_0) , first (M_1) , and second (M_2) for all characterized breakthrough curves were then computed through numerical integration of the breakthrough data.

For each breakthrough curve, the calculated M_0 is used to describe the total mass of Cl passing through the corresponding well during the simulation duration. The first normalized moment is used to estimate the mean arrival time of the center of Cl mass (μ):

$$\mu = \frac{M_1}{M_0} \tag{10}$$

659 The variance σ^2 of breakthrough curves is calculated through

$$\sigma^2 = \frac{M_2}{M_0} - (\frac{M_1}{M_0})^2.$$
(11)

In general, the σ^2 represents the spread of the concentration distribution and is influenced by mechanical dispersion and molecular diffusion. In other words, this parameter can be used to describe the heterogeneity levels of hydraulic parameters within the simulation domain. The calculated M_0 , μ , and σ^2 of the simulated and "true" breakthrough curves are then compared, with the comparison scatterplots illustrated in Figure 13.



Figure 13: Temporal moment analysis of simulated versus observed breakthrough curves for four model cases. a) total mass (M₀), b) mean arrival time (μ), and c) variance (σ^2).

Figure 13a reveals that at the wells highly impacted by the Cl plume, significantly large M_0 values are estimated from the geological model (Case 2) in comparison to the observed ones. The estimation of M_0 at these wells improves gradually when the effective parameter model (Case 1) and the geostatistical model without geological information (Case 3a) are utilized for prediction, 672 while the geostatistical model with geological information (Case 3b) yields the best estimation of M_0 with smallest discrepancy between the simulated and observed values. To enlarge the 673 comparison results at the wells with small M_0 values, the logarithm of the simulated and 674 "observed" M_0 estimates are computed and compared, as shown in the subplot of Figure 13a. 675 The comparison results show that the geostatistical model with geological information is able to 676 adequately estimate M_0 values at almost all wells with all data points clustered around the 45° 677 678 line; however, biased M_0 estimates with relatively large scatters are obtained from other model 679 cases. These results imply that detailed heterogeneity and accurate K and S_s estimates are 680 required to adequately capture the total solute mass.

The comparison of the simulated and observed mean arrival time (μ) for all model cases is 681 illustrated in Figure 13b. Results show that the estimated mean arrival time at all wells were, on 682 average, shorter in comparison to the observed ones for all model cases. This may be attributed 683 to the poor estimation of K and S_s values in the source layer (AF1), where hydraulic head data 684 685 are limited for detailed heterogeneity characterization, resulting in biased prediction of solute 686 transport at early time. However, Case 3b still yields the best estimation of the mean arrival time 687 with relatively smaller discrepancy between the simulated and observed values in comparison to 688 other model cases. Based on these results, geostatistics-based HT is suggested to reveal 689 heterogeneity details for more accurate estimation of the mean solute arrival time, which is in 690 line with the conclusion provided by Illman et al. (2012) based on a sandbox study.

Figure 13c illustrates the comparison of the simulated and "observed" variances (σ^2). In general, Case 3b still performs the best in estimating the variances, followed by Cases 3a and 1, while Case 2 yields the worst result. However, the computed variances of breakthrough curves are typically smaller with apparent bias for all model cases in comparison to the observed ones, indicating under predictions of temporal spreading of the plume using the estimated K and S_s tomograms. This may be attributed to the loss of heterogeneity details when using municipal well data for large-scale site characterization. Even with geostatistical inversions, heterogeneity details of hydraulic parameters (K and S_s) can only be revealed where there are sufficient hydraulic head data. These results emphasize again that solute transport is strongly impacted by the heterogeneity of hydraulic parameters (K and S_s). Detailed characterization of subsurface heterogeneity at finer scales is suggested for solute transport investigations.

702 **7.** Conclusions

In this study, a synthetic 3D multi-aquifer/aquitard system that mimics the Mannheim East 703 wellfield is characterized using HT-based approaches through the interpretation of long-term 704 705 water-supply pumping/injection records (municipal well data). In particular, pumping/injection rate records from 13 water-supply wells and simulated hydraulic head observations at 28 706 monitoring locations are interpreted to map subsurface heterogeneities in hydraulic conductivity 707 708 (K) and specific storage (S_s) . To investigate the performance of different modeling approaches and the effect of data selection on inverse modeling, the synthetic system is successively 709 characterized using four groundwater models (effective parameter model, geological model, and 710 two geostatistical models with different prior information) through the interpretation of five 711 712 datasets consisting of different time durations and periods within a given year. The estimated Kand S_s tomograms from all investigated models are then applied to predict municipal well data 713 with the existing water-supply wells and independent pumping test data from additional water-714 supply wells for model validation. Additional model runs are performed to investigate the ability 715 716 of estimated K and S_s tomograms in predicting solute transport in subsurface conditions for a stronger form of validation of HT results. Our study results in the following findings andconclusions:

1. Results from all investigated models reveal that HT analysis of long-term pumping/injection and water-level records is feasible and yields reliable K and S_s estimates where hydraulic data are available. In comparison to traditional subsurface characterization with dedicated pumping tests, the utilization of such data is able to reveal large-scale heterogeneities of hydraulic parameters and yield K and S_s estimates representative of aquifer properties during existing pumping/injection event, while reducing cost and time requirements for site characterization.

2. To avoid the effect of uncertain initial conditions on inverse modeling when using long-726 727 term records for site heterogeneity characterization, pumping/injection records prior to 728 the observation data are accounted for during model calibration. In this study, pumping/injection records 180 days prior to the observation data were used; however, the 729 appropriate duration is dependent on site specific conditions. To minimize the effect of 730 uncertain initial conditions, while maintaining computational efficiency for inverse 731 modeling, preliminary characterization of well hydrographs at the study site is suggested 732 733 to select an appropriate length of prior pumping/injection records.

3. The calibration of the effective parameter model yields K and S_s estimates that are more representative to the effective hydraulic parameters of the upper layers, where most monitoring wells are screened with sufficient hydraulic head data. The utilization of these values yields significantly biased predictions of hydraulic head variations at monitoring wells, implying the importance of considering heterogeneity for subsurface characterization. With well identified geological layers and well estimated initial hydraulic parameters, the calibrated geological model is found able to provide relatively
adequate predictions of drawdown variations. However, additional hydraulic data at
different geological layers are still required to obtain reliable estimates of hydraulic
parameters for each hydrostratigraphic unit.

4. Stratigraphy information is verified to be of critical importance for large-scale 744 745 heterogeneity characterization, in which hydraulic data are typically sparsely located with limited number of monitoring wells. The geostatistical inversion of hydraulic head data 746 only, is able to reveal heterogeneity details where head data are concentrated; however, 747 748 the overall smooth patterns and poor predictions of independent pumping test data cause the estimated K and S_s tomograms to fail to represent site specific heterogeneities. After 749 incorporating the geological data as prior information, the geostatistical model reveals 750 751 greater detail of subsurface heterogeneity and yields K and S_s tomograms comparable to the "true" fields. The estimated K and S_s tomograms provide adequate predictions of not 752 only the municipal well data with the existing water-supply wells, but also independent 753 pumping test data from additional pumping wells, implying that these estimated hydraulic 754 parameter fields can be used to guide the construction of new water-supply wells. 755

5. The effect of data selection on inverse modeling is investigated by manually selecting
different datasets, on the basis of duration and period for model calibration. Based on the
comparison results, continuous data points with large water-level variations are suggested
to be incorporated for large-scale heterogeneity characterization using the geostatistical
model with geological information. However, new approaches need to be developed for
big data synthesis and intelligent data selection for inverse modeling.

762 6. Synthetic conservative solute transport simulations conducted with various estimated hydraulic parameter fields (effective parameter model, geological model, and 763 geostatistical models with different prior information) reveal that solute migration is 764 strongly impacted by the heterogeneity of hydraulic parameters (K and S_s). Although the 765 calibrated geological model is able to provide adequate predictions of head variations at 766 monitoring wells, it yields poor predictions of contaminant transport due to the neglect of 767 intralayer heterogeneities and poor estimation of K and S_s values at the source layer 768 where hydraulic head data are few for model calibration. On the other hand, the 769 770 geostatistical inversion of the municipal well data incorporated with geological information yields K and S_s tomograms that can provide adequate predictions of not only 771 drawdown variations at monitoring wells but also solute transport in subsurface 772 773 conditions, indicating the superior application of this approach for large-scale heterogeneity characterization using the long-term water-supply pumping/injection 774 records. 775

The numerical experiments presented in this study provide a general framework for large-776 scale heterogeneity characterization using HT through the interpretation of long-term 777 778 pumping/injection and water-level records. Groundwater variations in this study are considered to be induced by only pumping/injection operations, while ignoring the influence from other 779 complexities (e.g., precipitation, evapotranspiration, surface water/groundwater interaction, long-780 781 term decline of groundwater levels due to dewatering operations, and etc.) which might be very important for long simulation periods. When applying this concept and technique to field 782 783 problems with actual municipal well data, a more sophisticated groundwater model with well 784 characterized sources of groundwater variations should be developed.

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- 792 Supplementary Information section graphically and the corresponding values are available from
- 793 the link 10.5281/zenodo.3723880.
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